# HUMAN HEALTH RISK ASSESSMENT 4835 GLENBROOK ROAD 

SPRING VALLEY
FORMERLY USED DEFENSE SITE (SVFUDS)
OPERABLE UNIT 3 (OU-3)
WASHINGTON, DC
CONTRACT DACA87-02-D-0005
DERP FUDS MEC/CWM PROJECT NO. C03DC091801
DERP FUDS HTRW PROJECT NO. C03DC091802


Prepared for:
U.S. ARMY

CORPS OF ENGINEERS BALTIMORE DISTRICT


Prepared by:
PARSONS
WASHINGTON, D.C. SEPTEMBER 11, 2009

September 28, 2009

Mr. Leland Reeser<br>CENAB-PP-E<br>Baltimore District Corps of Engineers<br>10 South Howard Street<br>Baltimore, MD 21201

Re: Revised Final Risk Assessment for 4835 Glenbrook Road Spring Valley FUDS, Washington, DC
Contract No. DACA87-02-D-0005, Task Order DA01
DERP-FUDS HTRW Project Number C03DC091802

Dear Mr. Reeser:
Enclosed are 3 hard copies of the Revised Final Risk Assessment for 4835 Glenbrook Road, Spring Valley FUDS. A Final Risk Assessment (July 28, 2009) incorporated comments received from USACE and partners was distributed electronically to the partners on July 29, 2009. AU provided additional comments to the Final Risk Assessment report on August 13, 2009. This version incorporates additional AU comments. A CD is included with each copy. This version has been submitted to the distribution list below.

If you have any questions, please call me at 202-714-5364 or Fan Wang-Cahill at 202-4696483.

Sincerely,


Paul Rich, P.E.
Project Manager
Distribution:
CEHNC - Cook (2)
USEPA - Hirsh (1)
USACHPPM - (1)
DCDOE - Sweeney (1)
TAP - deFur (1)
AU - Bridgham (1)

## COMMENT RESPONSE FORM

| Response to Comments on the "Draft Human Health Risk Assessment for 4835 Glenbrook Road" dated May 21, 2009 |  |  |  |
| :---: | :---: | :---: | :---: |
| Name: Steve Hirsh (EPA) |  | Date: 29 June 2009 |  |
| ITEM | REFERENCE | COMMENT | RESPONSE |
| 1 |  | EPA has completed its review of the Draft 4835 Glenbrook Road HHRA. The investigation conducted at this property thoroughly characterized conditions at the site. Further, appropriate remedial measures have been taken to mitigate potential risks associated with environmental exposures. |  |
| 2 |  | With respect to the AU suggestion that arsenic should be evaluated in this assessment, it is EPA's opinion that the report can be finalized with the current arsenic discussion, or could be modified to include a more detailed analysis of arsenic. The PRG for arsenic at this site is $20 \mathrm{mg} / \mathrm{kg}$. Even if $20 \mathrm{mg} / \mathrm{kg}$ does not represent bg conditions at the site, with no other risk drivers, this concentration is close to the risk-based goal that would be established for arsenic in soil under a residential exposure scenario ( $22 \mathrm{mg} / \mathrm{kg}$ ). Further, if the 95th percent UCL for arsenic at 4835 Glenbrook is 11.2 $\mathrm{mg} / \mathrm{kg}$ for a residential receptor, this would approximately equate to an excess cancer risk in the low 10-5 range and a Hazard Quotient of 0.5. | A discussion of the risks from assumed exposures to arsenic has been added to the Uncertainty Analysis. Risk calculations have been added as an appendix (Appendix H). |
| 3 | 5.3.2.2, 5.5.0.2 | EPA has considered the comments provided by Peter deFur. <br> 1. The data in Table 5-2 indicate that a child | Noted. The text has been revised to incorporate EPA's comments. |

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|  |  | exposure with RME assumptions with all of the chemicals present, as is the case at this site, would have an excess risk of developmental problems, and that cobalt and aluminum have combined risk HIs over 1.0. In fact, thyroid and hematopoietic effects are at an HI of 0.9; remarkably close to 1.0. The effects cannot be separated entirely, but are integrated by the child who may suffer from multiple effects, exacerbating each one because of the multiple exposures and effects. <br> 2. The HIs for development for the combination of chemical exposures exceed HI of 1.0, indicating excess risk that should be treated appropriately. The closing sentence of negligible risks is not right and should be modified to include the nonnegligible risks for RME children for developmental effects from the multiple chemicals. <br> To address concerns raised by the first comment, for exposure to mixed soils ( $0-10$ feet) under a residential exposure scenario (child), the combined HI for cobalt and aluminum is 1 (developmental effects). This value is appropriately rounded from 1.058. This does not constitute an unacceptable non-cancer risk. <br> With regard to the second part of the first comment, the HI values for thyroid and |  |

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|  |  | hematopoietic effects are each 0.9 ; while this value is close to 1 , it does not trigger the need for action. Approaching a benchmark value (such as an HI of 1), even for several target organ endpoints, does not imply that a threat is imminent or cumulative. EPA believes adequate margins of safety are built in to both the dose equations and the toxicity criteria relied upon to estimate potential risk; this provides protectiveness under circumstances such as this. <br> Regarding the second comment, as stated above, the highest calculated HI for developmental effects is 1 . This does not indicate the need for action. The risk calculations in the report were properly performed, and EPA believes the conclusions are valid. EPA agrees with Peter that the risks should not be referred to as "negligible", a better term might be below EPA's established risk threshold. |  |

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| Name: Jim Sweeney (DDOE)) |  | Date: 29 June 2009 |  |
| ITEM | REFERENCE | COMMENT | RESPONSE |
| 1 |  | As you know, due to the lack of a staff toxicologist, DDOE has, in the past, relied on EPA's expertise in reviewing risk assessments for the Spring Valley project and as such we agree with the comments that EPA has made on the risk assessment for 4835 Glenbrook. Further, we agree with American University's request that arsenic be more fully addressed in the risk assessment. Arsenic is and always has been the main Chemical of Concern at the site and as such the potential risk from arsenic should be completely assessed, regardless of the perceived probable outcome of the assessment. | A discussion of the risks from assumed exposures to arsenic will be added to the Uncertainty Analysis. Arsenic risk calculations have been performed and included in an Appendix (Appendix H) for informational purposes. As shown by these calculations, risks for arsenic are within acceptable ranges. |

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Response to Comments on the "Draft Human Health Risk Assessment for 4835 Glenbrook Road" dated May 21, 2009

| Name: Bethany Bridgeham (AU) |  | Date: 29 June 2009 |  |
| :---: | :---: | :---: | :---: |
| ITEM | REFERENCE | COMMENT | RESPONSE |
| 1 |  | This risk assessment is one of the elements that AU will need to judge the efficacy of the USACE cleanup efforts at this location. The focus of the risk assessment is on the hypothetical human health risk associated with exposure to chemical residuals that remain at the site. Other information required for final decision making includes historical documents, reports of site geophysics, a report on the test pitting analysis, the results of the geotechnical borings, and maps showing 4835 Glenbrook in the context of USACE activities on neighboring portions of the SVFUDS. <br> The risk assessment generally followed a protocol that was the subject of several rounds of comments by AU. The risk assessment concludes that the lifetime excess upperbound cancer risks based on conditions of exposure are less than the CERCLA target risk range of 10-6 for all receptors and that the non-cancer hazard indices are less than or equal to 1 (for developmental effects) which is borderline to the CERCLA acceptable range. Unfortunately, the risk assessment did not include arsenic which is the most significant chemical of potential concern at the site. The omission of arsenic seriously underestimates the potential risk and is a fatal flaw for the risk assessment. | The Spring Valley arsenic remediation goal agreed upon by USACE, USEPA, and DDOE is $20 \mathrm{mg} / \mathrm{kg}$. This was jointly proposed by the Partners. This concentration is considered protective of human health and the environment. The Scientific Advisory Panel, established to assist the community in understanding the overall approach to technical issues affecting Spring Valley, recommended adopting this remediation goal, saying that "the level should not pose a health hazard to the community and should not threaten the natural ecological systems of northwest Washington, DC." (Scientific Advisory Panel Report, May 29, 2002 Meeting). <br> However, the cancer risk and hazard index for arsenic have been calculated for adult residents, child residents and outdoor workers and included in an Appendix (Appendix H) for informational purposes. As shown by these calculations, risks for arsenic are within acceptable ranges. A discussion of the risks from exposures to arsenic has been added to the Uncertainty Analysis. |
| 2 |  | The activities of the Army during the period of occupancy of the AUES were primarily directed toward the testing and development of chemical | See response to Comment 1. |

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| :---: | :---: | :---: | :---: |
| Name: Bethany Bridgeham (AU) |  | (AU) $\quad$ Date: 29 June 2009 | Date: 29 June 2009 |
| ITEM | REFERENCE | COMMENT | RESPONSE |
|  |  | agents such as Mustard, Lewisite, and Adamsite. Many of the materials used in these tests were organoarsenicals or contained inorganic arsenic. Over the years, most of these arsenic-containing compounds have degraded and released inorganic arsenic into the environment. Inorganic arsenic, which has been found on campus at levels exceeding $1,000 \mathrm{ppm}$, is the most significant COPC at the site and has triggered the largest amount of response action. Early in the process, the USACE and other interested parties agreed on a preliminary remediation goal (PRG) for arsenic of 20 ppm . AU's concurrence with this PRG was predicated on a statistical analysis of data available at the time which suggested that cleanup of individual arsenic concentrations of 20 ppm or greater would result in an aggregate average soil concentration equal to the background concentration. AU has a fundamental risk management philosophy that people occupying AU property should not be at greater health risk than people occupying uncontaminated property in the NW District of Columbia/Adjoining Montgomery County, MD area. This position has been enunciated many times by AU in comment documents and at meetings. EPA guidance is clear that PRGs can be modified during the CERCLA process, and indeed this is often the case, especially when the PRG is not risk- or background-based. |  |

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| Name: | hany Bridgeha | (AU) $\quad$ Date: 29 June 2009 |  |
| ITEM | REFERENCE | COMMENT | RESPONSE |
| 3 |  | In this risk assessment, the USACE elected to remove arsenic from consideration by comparing the residual arsenic concentrations at the site to the PRG rather than to a risk-based value. To put things into context, the $95 \%$ upper confidence limit on the mean arsenic concentration as calculated by the USACE is 11.2 ppm with a range of 0.69 ppm to 19.9 ppm and a mean assuming a normal distribution of 9.3 ppm . The current risk-based concentration (September 2008 RBC Tables) is 0.39 ppm for residential soil and the SSL for protection of groundwater is $1.3 \times 10-3 \mathrm{ppm}$. Using EPA's standard residential default values, the cancer risk associated with the 95\% UCL residual arsenic concentration is $3 \times 10-5$, which is many orders of magnitude higher than the risks presented in the risk assessment. An estimate of the non-cancer hazard quotient for arsenic is 0.52 . Arsenic exerts its non-carcinogenic toxicity through a variety of endpoints, however, based on information contains in IRIS, one is developmental effects. If this is indeed the case, the hazard index for the site would exceed the value of 1 , possibly resulting in a different risk management decision based on the potential for developmental effects. | See response to Comment 1. |

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| Name: Bethany Bridgeham (AU) |  | Date: 29 June 2009 |  |
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| 4 |  | The arsenic concentrations remaining at the site also exceed background levels as shown in the 2007 Background Soil Sampling Report. According to the data in this report as input to EPA's ProUCL software, the $95 \%$ UCL for background is 6.7 ppm and the arithmetic mean background is 5.9 ppm , both significantly lower than the corresponding residual concentrations at 4835 Glenbrook. The background concentration is equivalent to a cancer risk of $2 \times 10-5$ based on EPA defaults, therefore, to a first approximation, an individual occupying the property at 4835 Glenbrook may be subjected to an additional cancer risk of $1 \times 10-5$, over and above the background risk associated with arsenic. | A discussion of the risks from exposures to arsenic has been added to the Uncertainty Analysis. |
| 5 |  | The exclusion of arsenic from a risk assessment is also unprecedented in risk assessments performed for the SVFUDS including the first risk assessment for this property. It should be noted that The exposure point concentration for arsenic at Lot 18 was substantially lower than for 4835 Glenbrook ( 6.9 ppm at Lot 18 compared to 11.2 ppm at 4835 Glenbrook), however, arsenic was appropriately retained in the risk assessment for Lot 18 and discarded at 4835 Glenbrook. | The same approach was used in the Lot 18 risk assessment. The maximum detected arsenic concentration was compared to the screening value of $20 \mathrm{mg} / \mathrm{kg}$. However, arsenic was detected at concentrations in excess of the screening value at both $0-2 \mathrm{ft}$ bgs ( $28 \mathrm{mg} / \mathrm{kg}$ ) and 10 ft bgs ( $42.3 \mathrm{mg} / \mathrm{kg}$ ). Therefore, arsenic was retained as a COPC in the Lot 18 risk assessment. This is shown in Tables D.1A and D.1B of the Lot 18 risk assessment. In contrast, the maximum detected concentration of arsenic at 4835 Glenbrook Rd. (i.e., $19.9 \mathrm{mg} / \mathrm{kg}$ ) was below the screening value. |

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| Name: Bethany Bridgeham (AU) |  | (AU) $\quad$ Date: 29 June 2009 |  |
| ITEM | REFERENCE | COMMENT | RESPONSE |
| 6 |  | The absence of the required documentation for this site makes a risk management decision by AU difficult; but the lack of a risk assessment including the most significant chemical of concern at the site makes this decision all but impossible. The degree of risk associated with arsenic at this site may be acceptable to AU, however, this cannot be determined until the appropriate calculations are performed. The calculations presented in these comments are only estimates, based on EPA defaults. AU strongly recommends that this risk assessment be redone incorporating arsenic in the calculations so that we may come to a decision at this site. | The risks and HIs due to exposure of arsenic detected in soil have been calculated for adult residents, child residents, and outdoor workers. The arsenic risk calculations are included in Appendix H. A discussion of the risks from exposures to arsenic has been added to the Uncertainty Analysis. |

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| Name: Peter L. deFur, PH.D (TAPP) |  | Date: 29 June 2009 |  |
| ITEM | REFERENCE | COMMENT | RESPONSE |
| 1 | General | The ES says that the risks do not exceed the benchmarks and then says the HI is 2 , but none of the individual chemicals is HI> 1. This explanation does not fly. If the cumulative risk $\mathrm{HI}>2$, then the risks are NOT acceptable. | RAGS A (USEPA 1989) indicates that if the overall HI is greater than one, the HIs should be separated by toxic endpoint. Then, the HIs for the individual toxic endpoints are compared to the threshold value of 1 . None of the HIs for the individual toxic endpoints exceeded 1 and, therefore, the risks are acceptable. <br> During finalization of the risk assessment report, it was noted that the vegetable intake rate was incorrect and \% consumption was not applied to the intake rate. Therefore, the vegetable ingestion pathway was recalculated. Following this change, the HIs for all receptors at the site do not exceed the benchmark level of concern of 1 . |
| 2 | Section 1.0.1.1- | change the wording - not "procedures to perform," but the results of performing- this text is a holdover from the work plan. | Noted. The text in Paragraph 1.1.01 has been changed to the following: "The purpose of this report is to present the results of a human health risk assessment (RA) that estimated the potential risks/hazards to current and future receptors from site-related contamination in the soil at the 4835 Glenbrook Road property..." |
| 3 | 1.3.0.8 | delete the following: "that can also originate from sources other than mustard agent" because this information can be put somewhere else and it does not matter for this site HHRA-we are assuming, rightly so, that the mustard | Noted. The indicated text has been deleted, as requested. The text has been expanded to include the locations of the four samples in which |

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| Name: Peter L. deFur, PH.D (TAPP) |  | Date: 29 June 2009 |  |
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|  |  | breakdown products come from old mustard gas that was used/disposed at this site. This paragraph should refer to the location of the 4 samples where thiodiglycol was found. | thiodiglycol was detected, as requested. |
| 4 | $\begin{aligned} & \text { Section } \\ & \text { 2.2.0.2 } \end{aligned}$ | So how will the four chemicals be handled without EPA Region III RBC's? I see that the uncertainty section has them in it, but this section needs to indicate that these four will be covered in that section. Not all 4 should be relegated to the ranks of unimportant. <br> Iodine pentaflouride has some toxicity information and it actually breaks down into hydrogen fluoride (HF) - which is a potent acid. The other chemicals do have some information and one, 1,2,3,4 tetrahydro-1,6-dimethyl-4-(1-methylethyl)-napthalene, seems to have little if any toxicological information. The other two chemicals are insect pheromones. | The four chemicals that do not have Regional Screening Levels also do not have toxicity values that can be used to quantitatively evaluate the risks from exposures. These toxicity values must be taken from the primary sources that are listed in, or fit the descriptions in, USEPA $(2003,2009)$ guidance. This list of sources includes (see paragraphs 4.1.0.3 and 4.2.0.3 in the report): <br> - USEPA's IRIS <br> - USEPA's Provisional Peer Reviewed Toxicity Values (PPRTVs) <br> - Agency for Toxic Substances and Disease Registry's (ATSDR) Minimal Risk Levels (MRLs) <br> - OEHHA's Toxicity Criteria Database <br> - USEPA's Health Effects Summary Tables <br> Chemicals without peer-reviewed toxicity values from the sources listed in USEPA $(2003,2009)$ guidance may only be evaluated qualitatively in the Uncertainty Analysis, unless USEPA provides an interim toxicity value to use or agreement is reached with USEPA on an interim toxicity value. Therefore, no change is proposed to the report. |

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| Name: Peter L. deFur, PH.D (TAPP) |  | Date: 29 June 2009 |  |
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|  |  |  | target organs that could sum to greater than 1 without adverse effects. <br> No change is proposed. |
| 8 | 5.4.2.7- | Where does the RA include both child and adult exposures for a resident? These two exposures are added somewhere, but I do not see where the addition is done. | This was not done. USEPA $(1991,2002,2004)$ guidance recommends that residential exposures be evaluated for a total exposure duration of 30 years, with the first 6 years as a child and the remaining 24 years as an adult. Since the workplan called for the evaluation of adult exposures for 30 years, child and adult residents were evaluated separately, as stated in the Executive Summary (paragraph 6) and Section 3.3.5. No change is proposed. |
| 9 | 5.4.3.5 | The risks from iodine pentaflouride and from tellurium need to be added to the total and neither is trivial. | Following USEPA $(2003,2009)$ guidance, there are no toxicity values for iodine pentaflouride and tellurium that may be used to quantitatively estimate the risks for human exposures. No change is proposed. |
| 10 | 5.5.0.2 | The HIs for development for the combination of chemical exposures exceed HI of 1.0, indicating excess risk that should be treated appropriately. The closing sentence of negligible risks is not right and should be modified to include the non-negligible risks for RME children for developmental effects from the multiple chemicals. | See response to comment on Section 5.3.2.2 above. |
| 11 | Table 2-2 | has an error for tellurium. The background is 5.0 ppm , there is no RPG and the maximum detected is 6.6 ppm . The screening level is listed as 39.11 ppm , when the selection criteria indicate that 5.0 ppm should be the | Noted. A PRG of $39.11 \mathrm{mg} / \mathrm{kg}$ should have been listed for tellurium, as indicated by footnote 3. The table has been revised. |

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## References

USEPA, 1989. Risk Assessment Guidance for Superfund (RAGS), Volume 1 - Human Health Evaluation Manual (Part A). Interim final. Office of Emergency and Remedial Response. Washington, DC. EPA/540/1-89/002.

USEPA, 1991. Risk Assessment Guidance for Superfund; Volume 1 - Human Health Evaluation Manual Supplemental Guidance. Standard Default Exposure Factors. Interim Final. Office of Solid Waste and Emergency Response. OSWER Directive 9285.6-03. March 25.

USEPA, 2002. Supplemental guidance for developing soil screening levels for Superfund sites. OSWER 9355.4-24.
USEPA, 2003. Human health toxicity values in Superfund risk assessments. OSWER Directive 9285.7-53.

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USEPA, 2004. Risk Assessment Guidance for Superfund, Volume I: Human Health Evaluation Manual, Part E, Supplemental Guidance for Dermal Risk Assessment, Final. U.S. Environmental Protection Agency, Office of Superfund Remediation and Technology Innovation. July 2004.

USEPA, 2009. Regional Screening Levels for Chemical Contaminants at Superfund Sites. Available online at: http://www.epa.gov/reg3hwmd/risk/human/rbconcentration_table/index.htm

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| Name: Bethany Bridgeham (AU) |  | Date: 13 August 2009 |  |
| ITEM | REFERENCE | COMMENT | RESPONSE |
| 1 |  | Item No. 4 of Dr. DeFur's comments was cut off and incomplete. | Second portion of Dr. DeFur's comment Item No. 4 is "Iodine pentaflouride has some toxicity information and it actually breaks down into hydrogen fluoride (HF) which is a potent acid. The other chemicals do have some information and one, 1,2,3,4 tetrahydro-1,6-dimethyl-4-(1-methylethyl)-napthalene, seems to have little if any toxicological information. The other two chemicals are insect pheromones." Parsons' previous response to this complete comment in Item No. 4 remains the same. |
| 2 | Section 2.2.0.2 | Iodine pentafluoride is not an organic compound and therefore should not be in this section. $\mathrm{IF}_{5}$ is not stable in the environment - it will hydrolyze rapidly to hydrofluoric acid and iodic acid. It is unlikely that this was a correct identification. Dr DeFur has noted that no risk was calculated for $\mathrm{IF}_{5}$. This situation could be easily remedied by recognizing that each mol of $\mathrm{IF}_{5}$ hydrolyzes to form 5 mols for HF which could be addressed in the risk assessment (at least as an uncertainty). | Section 2.2.0.2 will be revised to state that three organic compounds and Iodine pentafluoride were detected. A discussion on $\mathrm{IF}_{5}$ will also be included in the Uncertainty Analysis as follow: "Iodine pentafluoride (as iodate) was detected in both of the soil samples that were analyzed for this chemical. Although the lab reported the detection was iodine pentafluoride, it is more likely that an iodate salt was detected; e.g., sodium iodate $\left(\mathrm{NaIO}_{3}\right)$, silver iodate $\left(\mathrm{AgIO}_{3}\right)$, and calcium iodate $\left(\mathrm{Ca}\left(\mathrm{IO}_{3}\right)_{2}\right)$. In addition to the uncertain identity of the actual iodate present, there are no toxicity values available from the approved sources listed in USEPA (2003) guidance. Thus, the effects from assumed exposures to iodates can not be quantified.." |

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| Name: Bethany Bridgeham (AU) |

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| Name: Bethany Bridgeham (AU) |  | Date: 13 August 2009 |  |
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| 4 | Section 3.3.2.1. | Several metals were screened out of a plant bioaccumulation assessment on the basis of an EPA document that concerns mostly sediment and fish. EPA's HHRAP (Human health risk assessment protocol for hazardous waste combustion facilities, EPA530-R-05-006) contains soil to plant transfer factors for numerous metals and other relevant chemicals. This resource should be used to make sure that some of the uncertainties have been eliminated. | Although the basis of the EPA guidance document is sediment and fish, this is the most current USEPA guidance for all bioaccumulative pathways. No change is proposed. |
| 5 | Section 3.4.1.1.2. | The PEF was calculated based on Philadelphia International Airport data. The uncertainties of assuming that Spring Valley meteorology and Philadelphia meteorology are the same should be discussed. | The PEF was calculated based on the Philadelphia International Airport data, following USEPA guidance. This was the closest location given in USEPA guidance to Spring Valley. No change is proposed. |
| 6 |  | For this version, arsenic was included as an uncertainty. The analysis and the document would have been more useful if arsenic had been integrated into the text. Arsenic soil concentrations at this location exceed background levels. As stated in the risk assessment report, the central tendency concentration for arsenic at the site is $9.1 \mathrm{mg} / \mathrm{kg}$ and the $95 \%$ UCL is 11.17 $\mathrm{mg} / \mathrm{kg}$. The corresponding background values are $5.59 \mathrm{mg} / \mathrm{kg}$ and $6.69 \mathrm{mg} / \mathrm{kg}$, respectively. USACE calculated an RME risk for arsenic exposure of 2E05 . This is within EPA's generic risk range for Superfund, however, it is higher than AU's target risk. It should also be recognized that the current | Noted. The risk assessment was performed according to the current EPA guidance that specifies the sources of toxicity values that may be used in a risk assessment. Please note that interim or proposed toxicity values are not acceptable for use in a risk assessment, as per EPA guidance. The risk assessment results are within the USEPA acceptable risk range. The target end point of 20 $\mathrm{mg} / \mathrm{kg}$ was the Spring Valley arsenic remediation goal agreed upon by USACE, USEPA, and DDOE. This concentration is considered protective of human health and the environment. The Scientific Advisory Panel, established to assist the community in understanding the overall approach to technical issues affecting Spring Valley, recommended adopting this remediation goal |

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| ITEM | REFERENCE | COMMENT | RESPONSE |
|  |  | cancer slope factor for arsenic is very controversial and is highly likely to increase. USACE has used the current IRIS value o 1.5 per mg/kg/day. EPA's office of water uses a slope factor that is somewhat over 2X higher than this ( 3.67 per $\mathrm{mg} / \mathrm{kg} / \mathrm{day}$ ). EPA's has currently proposed increasing the slope factor to somewhere in the neighborhood of 25 per $\mathrm{mg} / \mathrm{kg} /$ day. This would be a 15 -fold increase over the value used by USACE and would put the resultant risk for 4835 Glenbrook out of EPA's target risk range for superfund. The new slope factor is designed to be protective against internal cancers such as lung and bladder cancer whereas the old slope factor is based on skin cancer alone. EPA's plans are consistent with the recommendations of the National Academy of Sciences and EPA’s Science Advisory Board. This proposed action is another argument for cleanup to background. | saying that "the level should not pose a health hazard to the community and should not threaten the natural ecological systems of northwest Washington, DC." (Scientific Advisory Panel Report, May 29, 2002 Meeting). No change is proposed. |
| 7 | Additional comment | As a follow-up to the discussion we had last week, about the University's belief that additional remediation at 4835 Glenbrook is necessary to reduce the arsenic to background as was done at the majority of properties in Spring Valley, I asked Paul for his assessment of the data from 4835. <br> Paul performed several simulations of soil removal at 4835 GB to see how it might be possible to attain statistical equivalence with background. Based on his calculations, we would suggest that the easiest way is to replace soil represented by | Noted. The risk assessment results are within the USEPA acceptable risk range. The target end point of 20 $\mathrm{mg} / \mathrm{kg}$ was the Spring Valley arsenic remediation goal agreed upon by USACE, USEPA, and DDOE. This concentration is considered protective of human health and the environment. The Scientific Advisory Panel, established to assist the community in understanding the overall approach to technical issues affecting Spring Valley, recommended adopting this remediation goal, saying that "the level should not pose a health hazard to the community and should not threaten the natural |

## COMMENT RESPONSE FORM

| Response to Comments on the "Final Human Health Risk Assessment for 4835 Glenbrook Road" dated July 28, 2009 |  |  |  |
| :---: | :---: | :---: | :---: |
| Name: Bethany Bridgeham (AU) |  | Date: 13 August 2009 |  |
| ITEM | REFERENCE | COMMENT | RESPONSE |
|  |  | samples over 18 ppm and replace them with clean fill. <br> Paul created a table with locations, depths, and concentrations of the samples that could be removed and a sketch of the areas that would be included (marked in red). Basically this involves the area on the north side of the house plus another isolated area in the southwest portion of the lot. Paul obtained the locations from the tables in the risk assessment. His simulations are based on removal of at least 2 ft of soil over the area shaded in red next to the house and 5 ft of soil in the isolated area. If the clean fill has an arsenic concentration at or below the background average, we belive this work would bring the whole lot to below background. | ecological systems of northwest Washington, DC." <br> (Scientific Advisory Panel Report, May 29, 2002 <br> Meeting). <br> Arsenic impacted soil removal has been performed at the site. All arsenic impacted soil detected at concentrations exceeding $20 \mathrm{mg} / \mathrm{kg}$ were removed from the site and replaced with backfill soil. The arsenic concentrations detected in the four backfill soil samples were 2.28 $\mathrm{mg} / \mathrm{kg}, 1.77 \mathrm{mg} / \mathrm{kg}, 2.04 \mathrm{mg} / \mathrm{kg}$, and $2.21 \mathrm{mg} / \mathrm{kg}$. The backfill soil analytical report is included as Attachment 1. The removed full and partial arsenic grids are illustrated in the figure in Attachment 2. As shown in the figure, grids $(-150,50),(-150,30),(-170,30)$ and $(-170,10)$, and portion of grids $(-150,10)$ and $(-150,-10)$ were removed from 2' to 5' bgs. Approximately $75 \%$ of the area adjacent to the garage that was marked red by Dr. Chrostowski in Attachment 3 and suggested to be removed, was previously excavated and replaced with backfill soil. Considering all the arsenic grids that were removed from the site, the actual overall exposure to arsenic impacted soil is further reduced because the arsenic concentrations in the backfill soil are less than the background level. <br> No change is proposed. |

## COMMENT RESPONSE FORM

ATTACHMENT 1
BACKFILL SOIL ANALYTICAL LAB REPORT

# WASTE STREAM TECHNOLOGY, INC. 

302 Grote Street
Buffalo, NY 14207
(716) 876-5290

Analytical Data Report
Report Date: 09/03/09
Work Order Number: 7H07004

Prepared For<br>Scott Burns<br>Sevenson/G-Jobs<br>2749 Lockport Road<br>Niagara Falls, NY 14305

Fax: (202) 237-5895
Site: Spring Valley G-203

Enclosed are the results of analyses for samples received by the laboratory on 08/07/07. If you have any questions concerning this report, please feel free to contact me.

Sincerely,


Daniel W. Vollmer, Laboratory QA/QC Officer

ENVIRONMENTAL LABORATORY ACCREDITATION CERTIFICATION NUMBERS
NYSDOH ELAP \#11179 NJDEPE \#73977 PADEP \#68757 CTDPH \#PH-0306 MADEP \#M-NY068


| Sevenson/G-Jobs | Project: | Spring Valley G-203 Backfill |
| :--- | :---: | :---: |
| 2749 Lockport Road | Project Number: | Spring Valley G-203 |

ANALYTICAL REPORT FOR SAMPLES

| Sample ID | Laboratory ID | Matrix | Date Sampled | Date Received |
| :---: | :---: | :---: | :---: | :---: |
| 203-BF(G1)-2751-0 | 7H07004-01 | Soil | 08/02/07 14:35 | 08/07/07 08:45 |
| 203-BF(G2)-2752-0 | 7H07004-02 | Soil | 08/02/07 14:45 | 08/07/07 08:45 |
| 203-BF(G3)-2753-0 | 7H07004-03 | Soil | 08/02/07 14:55 | 08/07/07 08:45 |
| 203-BF(C)-2754-0 | 7H07004-04 | Soil | 08/02/07 15:05 | 08/07/07 08:45 |


| Sevenson/G-Jobs | Project: | Spring Valley G-203 Backfill |
| :--- | :---: | :--- |
| 2749 Lockport Road | Project Number: | Spring Valley G-203 |
| Niagara Falls NY, 14305 | Project Manager: | Scott Burns |

## Metals by EPA 6000/7000 Series Methods

Waste Stream Technology

| Analyte | Result | Reporting |  |  |  |  |  |
| :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- |
| Nimit | Units | Dilution | Batch | Prepared | Analyzed | Method | Notes |


| 203-BF(G1)-2751-0 (7H07004-01) Soil | Sampled: 08/02/07 14:35 | Received: | 08/07/07 0 |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Boron | ND | 25.0 | $\mathrm{mg} / \mathrm{kg}$ dry | 5 | AH70720 | 08/07/07 | 08/07/07 | EPA 6010B | U |
| Mercury | 0.036 | 0.010 | " | 1 | AH71803 | 08/18/07 | 08/18/07 | EPA 7471A |  |
| Silver | 3.35 | 2.50 | " | 5 | AH70720 | 08/07/07 | 08/07/07 | EPA 6010B |  |
| Aluminum | 11800 | 250 | " | 100 | " | " | 08/08/07 | " |  |
| Arsenic | 2.28 | 1.70 | " | 1 | " | " | 08/07/07 | " |  |
| Barium | 23.2 | 5.00 | " | 5 | " | " | " | " |  |
| Beryllium | ND | 2.50 | " | " | " | " | 08/07/07 | " | U |
| Calcium | 295 | 12.5 | " | " | " | " | " | " | B, J-06 |
| Cadmium | ND | 5.00 | " | " | " | " | " | " | U |
| Cobalt | ND | 5.00 | " | " | " | " | 08/07/07 | " | U |
| Chromium | 23.6 | 5.00 | " | " | " | " | 08/07/07 | " |  |
| Copper | 38.3 | 5.00 | " | " | " | " | " | " |  |
| Iron | 40700 | 830 | " | 100 | " | " | 08/08/07 | " |  |
| Magnesium | 305 | 60.0 | " | 5 | " | " | 08/07/07 | " |  |
| Manganese | 64.6 | 5.00 | " | " | " | " | " | " |  |
| Nickel | 5.81 | 5.00 | " | " | " | " | 08/07/07 | " |  |
| Lead | ND | 20.5 | " | " | " | " | " | " | U |
| Antimony | ND | 7.00 | " | " | " | " | " | " | U |
| Selenium | ND | 7.00 | " | " | " | " | " | " | U |
| Thallium | ND | 5.00 | " | " | " | " | " | " | U |
| Vanadium | 32.5 | 5.00 | " | " | " | " | 08/07/07 | " |  |
| Zinc | 31.2 | 20.0 | " | ${ }^{\prime \prime}$ | " | " | 08/07/07 | " |  |
| Potassium | 290 | 14.0 | " | 1 | AH70723 | 08/07/07 | 08/13/07 | " |  |
| Sodium | 21.2 | 12.0 | " | " | " | " | " | " | B |
| Tin | ND | 5.00 | " | 5 | AH70720 | 08/07/07 | 08/07/07 | " | U |


| Sevenson/G-Jobs | Project: | Spring Valley G-203 Backfill |
| :--- | :---: | :---: |
| 2749 Lockport Road | Project Number: | Spring Valley G-203 |
| Niagara Falls NY, 14305 | Project Manager: | Scott Burns |

## Metals by EPA 6000/7000 Series Methods

Waste Stream Technology

| Analyte | Result | Reporting Limit | Units | Dilution | Batch | Prepared | Analyzed | Method | Notes |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 203-BF(G2)-2752-0 (7H07004-02) Soil | Sampled: 08/02/07 14:45 | Received: | 08/07/07 |  |  |  |  |  |  |
| Boron | ND | 25.0 | $\mathrm{mg} / \mathrm{kg}$ dry | 5 | AH70720 | 08/07/07 | 08/07/07 | EPA 6010B | U |
| Mercury | 0.028 | 0.010 | " | 1 | AH71803 | 08/18/07 | 08/18/07 | EPA 7471A |  |
| Silver | ND | 2.50 | " | 5 | AH70720 | 08/07/07 | 08/07/07 | EPA 6010B | U |
| Aluminum | 9700 | 250 | " | 100 | " | " | 08/08/07 | ${ }^{\prime \prime}$ |  |
| Arsenic | 1.77 | 1.70 | " | 1 | " | " | 08/07/07 | " |  |
| Barium | 28.5 | 5.00 | " | 5 | " | " | " | " |  |
| Beryllium | ND | 2.50 | " | " | " | " | 08/07/07 | " | U |
| Calcium | 247 | 12.5 | " | " | " | " | " | " | B, J-06 |
| Cadmium | ND | 5.00 | " | " | " | " | 08/07/07 | " | U |
| Cobalt | 9.64 | 5.00 | " | " | " | " | " | " |  |
| Chromium | 15.5 | 5.00 | " | " | " | " | " | " |  |
| Copper | 22.2 | 5.00 | " | " | " | " | 08/07/07 | " |  |
| Iron | 26700 | 830 | " | 100 | " | " | 08/08/07 | " |  |
| Magnesium | 1460 | 60.0 | " | 5 | " | " | 08/07/07 | " |  |
| Manganese | 164 | 5.00 | " | " | " | " | " | " |  |
| Nickel | 7.43 | 5.00 | " | " | " | " | 08/07/07 | " |  |
| Lead | ND | 20.5 | " | " | " | " | " | " | U |
| Antimony | ND | 7.00 | " | " | " | " | " | " | U |
| Selenium | ND | 7.00 | " | " | " | " | " | " | U |
| Thallium | ND | 5.00 | " | " | " | " | " | " | U |
| Vanadium | 26.5 | 5.00 | " | " | " | " | 08/07/07 | " |  |
| Zinc | 28.5 | 20.0 | " | " | " | " | 08/07/07 | " |  |
| Potassium | 1260 | 14.0 | " | 1 | AH70723 | 08/07/07 | 08/13/07 | " |  |
| Sodium | 24.1 | 12.0 | " | " | " | " | " | " | B |
| Tin | ND | 5.00 | " | 5 | AH70720 | 08/07/07 | 08/07/07 | " | U |


| Sevenson/G-Jobs | Project: | Spring Valley G-203 Backfill |
| :--- | :---: | :---: |
| 2749 Lockport Road | Project Number: | Spring Valley G-203 |
| Niagara Falls NY, 14305 | Project Manager: | Scott Burns |

## Metals by EPA 6000/7000 Series Methods

Waste Stream Technology

| Analyte | Result | Reporting Limit | Units | Dilution | Batch | Prepared | Analyzed | Method | Notes |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 203-BF(G3)-2753-0 (7H07004-03) Soil | Sampled: 08/02/07 14:55 | Received: | 08/07/07 |  |  |  |  |  |  |
| Boron | ND | 25.0 | $\mathrm{mg} / \mathrm{kg}$ dry | 5 | AH70720 | 08/07/07 | 08/07/07 | EPA 6010B | U |
| Mercury | 0.021 | 0.010 | ${ }^{\prime \prime}$ | 1 | AH71803 | 08/18/07 | 08/18/07 | EPA 7471A |  |
| Silver | ND | 2.50 | " | 5 | AH70720 | 08/07/07 | 08/07/07 | EPA 6010B | U |
| Aluminum | 11500 | 250 | " | 100 | " | " | 08/08/07 | " |  |
| Arsenic | 2.04 | 1.70 | " | 1 | " | " | 08/07/07 | " |  |
| Barium | 41.4 | 5.00 | " | 5 | " | " | " | " |  |
| Beryllium | ND | 2.50 | " | " | " | " | " | " | U |
| Calcium | 8290 | 250 | " | 100 | " | " | 08/08/07 | " | J-06 |
| Cadmium | ND | 5.00 | " | 5 | " | " | 08/07/07 | " | U |
| Cobalt | 10.2 | 5.00 | " | " | " | " | ${ }^{\prime}$ | " |  |
| Chromium | 46.0 | 5.00 | " | " | ${ }^{\prime \prime}$ | " | " | " |  |
| Copper | 26.1 | 5.00 | " | " | " | " | " | " |  |
| Iron | 24800 | 830 | ${ }^{\prime \prime}$ | 100 | " | " | 08/08/07 | " |  |
| Magnesium | 6370 | 1200 | " | ${ }^{\prime}$ | " | " | " | " |  |
| Manganese | 258 | 5.00 | " | 5 | " | " | 08/07/07 | " |  |
| Nickel | 41.1 | 5.00 | " | " | " | " | " | " |  |
| Lead | ND | 20.5 | " | " | " | " | " | " | U |
| Antimony | ND | 7.00 | " | " | " | " | " | " | U |
| Selenium | ND | 7.00 | " | " | " | " | " | " | U |
| Thallium | ND | 5.00 | " | " | " | " | " | " | U |
| Vanadium | 33.4 | 5.00 | " | " | " | " | " | " |  |
| Zinc | 41.2 | 20.0 | " | " | " | " | " | " |  |
| Potassium | 2290 | 70.0 | " | " | AH70723 | 08/07/07 | 08/13/07 | " |  |
| Sodium | 58.6 | 12.0 | " | 1 | " | " | 08/13/07 | " | B |
| Tin | ND | 5.00 | " | 5 | AH70720 | 08/07/07 | 08/07/07 | " | U |


| Sevenson/G-Jobs | Project: | Spring Valley G-203 Backfill |
| :--- | :---: | :---: |
| 2749 Lockport Road | Project Number: | Spring Valley G-203 |
| Niagara Falls NY, 14305 | Project Manager: | Scott Burns |

## Metals by EPA 6000/7000 Series Methods

Waste Stream Technology

| Analyte | Result | Reporting <br> Limit | Units | Dilution | Batch | Prepared | Analyzed | Method | Notes |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 203-BF(C)-2754-0 (7H07004-04) Soil | Sampled: 08/02/07 15:05 | Received: 0 | 08/07/07 08 |  |  |  |  |  |  |
| Boron | ND | 25.0 | $\mathrm{mg} / \mathrm{kg}$ dry | 5 | AH70720 | 08/07/07 | 08/07/07 | EPA 6010B | U |
| Mercury | 0.031 | 0.010 | " | 1 | AH71803 | 08/18/07 | 08/18/07 | EPA 7471A |  |
| Silver | ND | 2.50 | " | 5 | AH70720 | 08/07/07 | 08/07/07 | EPA 6010B | U |
| Aluminum | 12900 | 250 | " | 100 | " | " | 08/08/07 | " |  |
| Arsenic | 2.21 | 1.70 | " | 1 | " | " | 08/07/07 | " |  |
| Barium | 40.9 | 5.00 | " | 5 | " | " | " | " |  |
| Beryllium | ND | 2.50 | " | " | " | " | " | " | U |
| Calcium | 15200 | 250 | " | 100 | " | " | 08/08/07 | " | J-06 |
| Cadmium | ND | 5.00 | " | 5 | " | " | 08/07/07 | " | U |
| Cobalt | 8.35 | 5.00 | " | " | " | " | " | " |  |
| Chromium | 22.6 | 5.00 | " | " | " | " | " | " |  |
| Copper | 28.8 | 5.00 | " | " | " | " | " | " |  |
| Iron | 31700 | 830 | " | 100 | " | " | 08/08/07 | " |  |
| Magnesium | 2750 | 60.0 | " | 5 | " | " | 08/07/07 | " |  |
| Manganese | 186 | 5.00 | " | " | " | " | " | " |  |
| Nickel | 15.3 | 5.00 | " | " | " | " | " | " |  |
| Lead | ND | 20.5 | " | " | " | " | " | " | U |
| Antimony | ND | 7.00 | " | " | " | " | " | " | U |
| Selenium | ND | 7.00 | " | " | " | " | " | " | U |
| Thallium | ND | 5.00 | " | " | " | " | " | " | U |
| Vanadium | 31.4 | 5.00 | " | " | " | " | " | " |  |
| Zinc | 36.8 | 20.0 | " | " | " | " | " | " |  |
| Potassium | 2040 | 70.0 | " | " | AH70723 | 08/07/07 | 08/13/07 | " |  |
| Sodium | 62.2 | 12.0 | " | 1 | " | " | 08/13/07 | " | B |
| Tin | ND | 5.00 | " | 5 | AH70720 | 08/07/07 | 08/07/07 | " | U |



## COMMENT RESPONSE FORM

ATTACHMENT 2
ARSENIC REMOVAL EXTENT FIGURE


## COMMENT RESPONSE FORM

ATTACHMENT 3
FIGURE ATTACHED TO THE ADDITIONAL COMMENT


## REVISED FINAL

## 4835 GLENBROOK ROAD HUMAN HEALTH RISK ASSESSMENT

## SPRING VALLEY FORMERLY USED DEFENSE SITE (SVFUDS) WASHINGTON, DC

Prepared For:

## U.S. ARMY CORPS OF ENGINEERS BALTIMORE DISTRICT



Prepared by:

## PARSONS

WASHINGTON, DC 20003

Primary Author:
Mark Rigby, PhD, REA

September 11, 2009

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## LIST OF ACRONYMS AND ABBREVIATIONS

| ABP | Agent Breakdown Products |
| :---: | :---: |
| AF | Soil-to-Skin Adherence Factor |
| AT | Averaging Time |
| ATSDR | Agency for Toxic Substances and Disease Registry |
| AU | American University |
| AUES | American University Experiment Station |
| bgs | Below Ground Surface |
| BW | Body Weight |
| $\mathrm{C}_{\text {air }}$ | COPC Concentration in Airborne Dust or Outdoor Air |
| CDI | Chronic Daily Intake |
| CENAB | United States Army Corps of Engineers, Baltimore District |
| CERCLA | Comprehensive Environmental Response, Compensation, and Liability Act |
| CF | Conversion Factor |
| cm | Centimeter |
| CNS | Central Nervous System |
| COPC | Chemical of Potential Concern |
| CSM | Conceptual Site Model |
| CSS | Chemical Safety Submission |
| CT | Central Tendency |
| CWM | Chemical Warfare Materiel |
| DAF | Dermal Absorption Fraction |
| DDOE | District Department of the Environment |
| DERP | Defense Environmental Restoration Program |
| DW | Dry Weight |
| ECBC | Edgewood Chemical and Biological Center |
| ED | Exposure Duration |
| EE/CA | Engineering Evaluation/Cost Analysis |
| EF | Exposure Frequency |
| EMS | Environmental Management Systems |
| EPC | Exposure Point Concentration |
| ERG | Emergency Removal Guideline |
| ET | Exposure Time |
| EV | Event Frequency |
| FI | Fraction Ingested |
| FS | Feasibility Study |
| ft | Feet |
| $\mathrm{F}(\mathrm{x})$ | Windspeed Distribution Function |
| FUDS | Formerly Used Defense Site |
| HEAST | Health Effects Assessment Summary Tables |
| HI | Hazard Index |
| HQ | Hazard Quotient |
| HTRW | Hazardous, Toxic and Radioactive Waste |
| IR | Soil Ingestion Rate |


| IR $_{\text {veg }}$ | Home-Grown Vegetable Ingestion Rate |
| :--- | :--- |
| IRIS | Integrated Risk Information System |
| kg | Kilogram |
| LOAEL | Lowest-Observed-Adverse-Effect Level |
| m | Meter |
| MD | Munitions Debris |
| MEC | Munitions and Explosives of Concern |
| mg | Milligram |
| MRL | Minimal Risk Level |
| NOAEL | No-Observed-Adverse-Effect Level |
| OAF | Oral Absorption Factor |
| OEHHA | Office of Environmental Health Hazard Assessment |
| OSWER | Office of Solid Waste and Emergency Response |
| PAHs | Polycyclic Aromatic Hydrocarbons |
| PEF | Particulate Emission Factor |
| PID | Photoionization Detector |
| PL | Preparation and Cooking Loss |
| PPRTV | Provisional Peer Reviewed Toxicity Value |
| Q/C wind | Dispersion factor |
| RA | Risk Assessment |
| RAGS | Risk Assessment Guidance for Superfund |
| RCRA | Resource Conservation and Recovery Act |
| RfC | Reference Concentration |
| RfD | Reference Dose |
| RfD | Dermal Reference Dose |
| RfD | Oral Reference Dose |
| RI | Remedial Investigation |
| RME | Reasonable Maximum Exposure |
| RSL | Regional Screening level |
| SA | Skin Surface Area |
| SF | Slope Factor |
| SF | Dermal Slope Factor |
| SF | Oral Slope Factor |
| SVFUDS | Spring Valley Formerly Used Defense Site |
| SVOC | Semi-Volatile Organic Compounds |
| TIC | Tentatively Identified Compound |
| Um $_{\mathrm{m}}$ | Mean Annual Wind Speed |
| Ut $_{\mathrm{t}}$ | Equivalent Threshold Value of Windspeed at 7 m |
| UCL | Upper Confidence Limit |
| URF | Unit Risk Factor |
| US | United States |
| USACE | United States Army Corps of Engineers |
| USEPA | United States Environmental Protection Agency |
| UTL | Upper Tolerance Limit |
| Fraction of Vegetative Cover |  |
| Volatile Organic Compound |  |
| UOC |  |

[^0] Contract No. DACA87-02-D-0005

## EXECUTIVE SUMMARY

ES. 1 A human health risk assessment (RA) was performed to estimate the potential risks/hazards to current and future receptors from site-related contamination in the soil at the 4835 Glenbrook Road property, located in Spring Valley, Washington, D.C. The type and magnitude of exposures to Chemicals of Potential Concern (COPCs) at the site were estimated, potential exposure pathways, receptors, and exposure scenarios were identified, and exposure was quantified. This RA was performed under contract DACA87-02-D-0005, Task Order DA01, DERP/FUDS MEC/CWM project no. C03DC091801 and DERP/FUDS HTRW project no. C03DC091802, for the U.S. Army Corps of Engineers, Baltimore District (CENAB).

ES. 24835 Glenbrook Road is part of the Spring Valley Formerly Used Defense Site (SVFUDS), an area of northwest Washington, D.C., that was formerly occupied by the American University Experiment Station (AUES). During World War I, the U.S. government established AUES to investigate the testing, production, and effects of noxious gases, antidotes, and protective masks and to conduct research and development on chemical warfare materiel (CWM), including mustard and lewisite agents, as well as adamsite, irritants, and smokes.

ES. 3 Test pit investigation and arsenic removal were performed at 4835 Glenbrook Road in accordance with the Amendment 1 Site-Specific Work Plan for the Test Pit Investigations at 4825 and 4835 Glenbrook Road Properties, March 18, 2008. 76 Test pits were investigated at the 4835 Glenbrook Road property. 62 test pits yielded no debris or cultural debris only; and 14 test pits included suspect AUES related items. 13 test pits yielded suspected AUES-related labware components (i.e., glass tubing, stoppers, glass fragments, etc.) and one test pit (TP49) yielded a Livens projectile. Although a Livens projectile was found in TP49, it should be noted that a Livens projectile is merely a gas drum. The gas drum contained no explosives, as those were external to the projectile. Low level analysis of the liquid in the projectile and the soil sample revealed no agents of concern. An x-ray of the projectile concluded that there were no explosives present. No other munitions debris (MD), munitions and explosives of concern (MEC) or CWM were found at the site.

ES. 4 Ten full or partial arsenic contaminated soil grids in the northern portion and three full or partial arsenic contaminated soil grids in the southern portion of the house were excavated. The grids, and associated extensions, were excavated until the arsenic concentrations in the confirmation samples were acceptable. More than 500 cubic yards of arsenic impacted soil were removed and disposed off-site.

ES. 5 A total of 185 soil samples were collected at the site, these soil samples are representative of soil still in place at the site. These samples were analyzed variously for the Spring Valley comprehensive list of parameters, including mustard, lewisite, agent breakdown products, VOCs, SVOCs, metals, explosives, and pesticides, and PCBs. The results of these samples guided interim removal measures to address potential residual risk while teams were still mobilized in the field. These analytical results were also used to identify the COPCs that were the focus of the investigation from that point forward, and which were evaluated in this RA.

ES. 6 The receptors evaluated in this RA include adult and child residents, as well as outdoor on-site workers. For future residents and outdoor workers, the risks associated with incidental ingestion of soil, inhalation of particulates from soil, and dermal contact with soil, were calculated. Ingestion of home-grown vegetables was also evaluated for residents. The residential pathway conservatively evaluates childhood exposure separately from adult exposures. Two depth intervals were evaluated for both receptors at this site: 0-2 feet below ground surface (bgs) to evaluate the risk associated with exposure to surface soil and $0-10$ feet bgs to account for the potential mixing of soil that may occur in the future due to excavation and/or construction at the site. Since outdoor workers were evaluated for assumed exposures to soils at $0-10 \mathrm{ft} \mathrm{bgs}$, the exposures estimated for this receptor are assumed to be protective of a construction worker as well.

ES. 7 The cumulative cancer risk estimates for child residents, adult residents, and outdoor workers are all well below the USEPA point of departure of $1 \times 10^{-6}$. Thus, unacceptable cancer risks to the receptors resident are not expected from assumed exposures to COPCs in soils at the site. Additionally, the hazard indexes (HIs) estimated for assumed exposures at the site do not exceed the benchmark level of concern of 1 . This indicates that unacceptable noncarcinogenic health effects are not expected from assumed exposures to COPCs in soils at the site.

ES. 8 Arsenic was not selected as a COPC because the maximum arsenic concentration remaining at the site is below the Spring Valley arsenic remediation goal of $20 \mathrm{mg} / \mathrm{kg}$, which was agreed upon by United States Army Corps of Engineers (USACE), United States Environmental Protection Agency (USEPA), and District Department of the Environment (DDOE). However, the cancer risk and hazard index for arsenic were calculated for adult residents, child residents and outdoor workers and included in Appendix H as requested by American University and DDOE for informational purposes. A discussion of the risks from exposures to arsenic is included in Section 5.4.4.3 of the Uncertainty Analysis. The combined RME risk and hazards of arsenic and the identified COPCs show that the risk estimates including arsenic are within the USEPA (1990) target risk range of $1 \times 10^{-6}$ to $1 \times 10^{-4}$ and the noncancer hazards do not exceed the threshold value of 1 (when summed by toxic endpoint for children). This indicates that assumed exposures to COPCs and arsenic at the site are unlikely to result in adverse noncarcinogenic health effects.

## SECTION 1 INTRODUCTION

### 1.1 PROJECT OVERVIEW

1.1.0.1 The purpose of this report is to present the results of a human health risk assessment (RA) that estimated the potential risks/hazards to current and future receptors from site-related contamination in the soil at the 4835 Glenbrook Road property, located in Spring Valley, Washington, D.C. This property is owned by American University (AU). The RA is based on analytical data, historical information, and recommendations/conclusions presented in previous investigation reports.
1.1.0.2 As described in detail in Section 2, an RA evaluating risk associated with soil contamination was previously performed for 4835 Glenbrook Road (Parsons, 2002). The RA concluded that the risk estimates did not exceed USEPA's target risk range and that no adverse health effects were expected for human receptors at the 4835 Glenbrook Road property. Those findings are re-evaluated here using additional data that has been collected since the last RA (Parsons, 2002).
1.1.0.3 This RA report was prepared under contract DACA87-02-D-0005, Task Order DA01, DERP/FUDS MEC/CWM project no. C03DC091801 and DERP/FUDS HTRW project no. C03DC091802, for the U.S. Army Corps of Engineers, Baltimore District (CENAB).

### 1.2 SVFUDS BACKGROUND

4835 Glenbrook Road is an approximately 0.5 acre private residential property within Operable Unit 3 (OU-3) of the Spring Valley Formerly Used Defense Site (SVFUDS). The SVFUDS is an area of northwest Washington, DC, that was formerly occupied by the American University Experiment Station (AUES). During World War I, the U.S. government established the AUES to investigate the testing, production, and effects of noxious gases, antidotes, and protective masks. The AUES was located on the grounds of the current AU and used additional property in the vicinity to conduct this research and development on chemical warfare materiel (CWM), including mustard and lewisite agents, as well as adamsite, irritants, and smokes. After the war, these activities were transferred to other locations and the site was returned to the owners. The SVFUDS location map is presented as Figure 1-1.

### 1.3 4835 GLENBROOK ROAD BACKGROUND

1.3.0.1 Over the years, numerous investigations have been performed at the 4835 Glenbrook Road property. These were conducted at different times, by different parties, and with different sampling objectives and analytical parameters. These include:

- 1992, Environmental Management Systems (EMS)
- 1996, Apex Environmental
- 1999, USEPA
- 1999-present, USACE/Parsons

1.3.0.2 In 1992, AU contracted Environmental Management Systems (EMS) to investigate conditions discovered during construction activities in the vicinity of what would become the 4825 and 4835 Glenbrook Road properties. At that time, the properties were under construction and the EMS letter reports from May and June 1992 are not sufficiently detailed to determine the exact locations of the incidents described or the sampling performed. Workers reportedly experienced eye and respiratory irritation during construction activities (EMS, 1992). A rusted drum, laboratory glassware, and a white granular material were reportedly encountered. EMS (1992) conducted soil gas probes, hand excavations around the drum, and collected various samples, including the white powder, which they concluded was the herbicide Silvex. Although it is now believed that the areas investigated were actually in the vicinity of the current driveway of 4825 Glenbrook Road, the investigation is discussed here because of the uncertainty associated with the specifics of the letter reports.
1.3.0.3 In June 1996, landscape workers at 4835 Glenbrook Road were excavating a large hole (about 6 feet in diameter and 4 feet in depth) to plant a tree in the front yard near the southwest corner of the house. They were overcome by an odor, experiencing eye and respiratory irritation, forcing activity to cease (Apex, 1996).
1.3.0.4 In the front yard, Apex Environmental, Inc. (Apex, 1996) advanced 24 soil probes to a depth of 4 feet on 2.5 -foot centers. As the probes were completed, the probe holes were screened for VOCs using a PID. Based on PID reading and visual inspection of the soil probes, an additional four soil borings were advanced and one soil sample was collected from each boring. Elevated levels of certain metals (with arsenic being of most concern) and volatile organic compounds (VOC) were found in the soil samples. The pH of some of the samples was low (i.e., as low as 3.9), indicating the presence of acids. Apex (1996) over-excavated the hole where the tree was to be planted to approximately 12 feet in diameter and 6 feet in depth, and removed laboratory glassware about 2 feet below grade. Remediation was confirmed with five post-excavation soil samples.
1.3.0.5 In the backyard, Apex (1996) advanced 91 soil probes to a depth of 4 feet on 10 -foot centers. Again, as the probes were completed, the probe holes were screened for VOCs using a PID. Based on PID readings, six soil samples, one from each location with high PID readings, were collected. Finally, Apex (1996) dug two test pits to a depth of 9 feet and a third test pit to a depth of 7 feet. During the excavation of the test pits, a PID was used to monitor air in the breathing zone of the workers and screen soils. There were no elevated air or soil PID readings and no visual or olfactory indications of contamination. The current layout of the property is shown in Figure 1-2.
1.3.0.6 As part of a larger investigation at the SVFUDS, USEPA Region III collected surface soil and subsurface soil samples in and around 4801, 4825, and 4835 Glenbrook Road to supplement their risk assessment (USEPA, 1999a). These three properties (Figure 1-2) form OU-3. At 4835 Glenbrook Road, USEPA collected three surface soil samples (i.e., 0 to 6 inches bgs ) in April 1999: G-01, G-02, and G-03.
1.3.0.7 Based on the results of the USEPA (1999a) Region III sampling, it was determined that the soil at these properties could have been affected by AUES activities in the vicinity of Burial Pits 1 and 2 at 4801 Glenbrook Road. Consequently, the USACE performed an Engineering Evaluation/Cost Analysis (EE/CA) for the three OU-3 properties (USACE, 2000). The EE/CA included extensive sampling to determine the nature and extent of contamination found in the surface and subsurface soils of the three OU-3 properties.

[^1]
1.3.0.8 In October 2000, in support of the EE/CA, Parsons used the quadrant procedure to collect four surface soil samples at 4835 Glenbrook Road for the mustard agent breakdown products (ABPs) dithiane, oxathiane, and thiodiglycol. Thiodiglycol, a non-specific mustard agent breakdown product, was detected at low levels in all four samples (i.e., OU3 MTL-4835-1, OU3 MTL-4835-2, OU3 MTL-4835-3, OU3 MTL-4835-4). However, dithiane and oxathiane were not detected in any of the samples. OU3 MTL-4835-SB was also sampled at 0-2, 2-4, and 4-6 feet bgs near the southeast corner of the house. These subsurface samples were also analyzed for the three mustard ABPs, all of which were non-detect in all three samples. Grid sampling for arsenic was also performed at 4835 Glenbrook Road in October 2000.
1.3.0.9 A RA for 4835 Glenbrook Road was conducted to evaluate the risk associated with exposure to contaminated soil at this property and completed in April 2002 (Parsons, 2002). This RA concluded that there was no actionable risk or hazard at this property. However, subsequent to that document (Parsons, 2002), the Spring Valley Remediation Endpoint for arsenic ( $20 \mathrm{mg} / \mathrm{kg}$ ) took effect. Eight full or partial grids on the 4835 Glenbrook Road property had arsenic levels exceeding this value and were identified for removal.

### 1.4 OBJECTIVES AND SCOPE

1.4.0.1 The objective of this effort is to conduct a site-specific quantitative RA for human receptors at the 4835 Glenbrook Road property. All previously collected data was evaluated following guidance from United States Environmental Protection Agency (USEPA, 1992a) to determine whether it was acceptable for use in an RA (Parsons 2009). Data that were considered acceptable were used to identify and screen chemicals of potential concern (COPCs). For the receptors present at the site, the RA estimated the magnitude of exposure to COPCs, identified potential exposure pathways, and quantified exposure. This information, in conjunction with toxicity information for the COPCs, provides a quantitative post-interim removal measures risk assessment and determines if there is potential unacceptable risk to human health associated with exposure to chemicals in the soil remaining at 4835 Glenbrook Road.
1.4.0.2 The RA was conducted using techniques and methodology recognized by the USACE and the USEPA. Reference and guidance documents used or consulted in preparation of the RA include:

- Risk Assessment Guidance for Superfund (RAGS), Volume 1, Human Health Evaluation Manual (Part A), interim final (USEPA, 1989a);
- Risk Assessment Guidance for Superfund (RAGS), Volume 1, Human Health Evaluation Manual, Supplemental Guidance, Standard Default Exposure Factors (USEPA, 1991a);
- Role of the Baseline Risk Assessment in Superfund Remedy Selection Decisions (USEPA, 1991b).
- Guidance for Data Usability in Risk Assessment (Part A) (USEPA, 1992a);
- Supplemental Guidance to RAGS: Calculating the Concentration Term, (USEPA, 1992b);
- Soil Screening Guidance: Technical Background Document (USEPA, 1996a);
- Exposure Factors Handbook (USEPA, 1997a);
- Supplemental Guidance for Developing Soil Screening Levels for Superfund Sites (USEPA, 2002);
- Human Health Toxicity Values in Superfund Risk Assessments (USEPA, 2003);
- Risk Assessment Guidance for Superfund (RAGS), Volume I, Human Health Evaluation Manual, Part E, Supplemental Guidance for Dermal Risk Assessment (USEPA, 2004a);
- On the computation of a $95 \%$ upper confidence limit of the unknown population mean based upon data sets with below detection limit observations (USEPA, 2006a); and
- Risk Assessment Guidance for Superfund (RAGS). Volume I: Human health evaluation manual. Part F, Supplemental Guidance for Inhalation Risk Assessment (USEPA, 2009a).
1.4.0.3 This RA only evaluates the risk associated with human exposure to soil contamination. Groundwater exposure pathways at the 4835 Glenbrook Road property are incomplete since municipal water is provided to the property and no springs are identified at this location. Therefore, the evaluation of potential risk from groundwater will be addressed separately, if necessary.


### 1.5 TECHNICAL APPROACH OVERVIEW

1.5.0.1 The four-step RA procedure recommended by USEPA (1989a) was used for this evaluation. The four steps are as follows:

1. data evaluation;
2. exposure assessment;
3. toxicity assessment; and
4. risk characterization.
1.5.0.2 The first step of the RA process involves an evaluation of available data. Section 2.1 describes the data from previous site investigations that were used in this evaluation. The data is also screened to identify the COPCs that will be evaluated in the subsequent steps.
1.5.0.3 The second step in the RA process is the exposure assessment. The purpose of the exposure assessment is to identify and evaluate the nature of the chemical releases, potential exposure pathways, potential receptors, and exposure scenarios. This involves the preparation of a Conceptual Site Model (CSM) to help determine which potential exposure pathways will be evaluated. The CSM is site specific and can be used to identify all potentially complete exposure pathways (for current and future human receptors).
1.5.0.4 Steps 3 and 4 (toxicity assessment and risk characterization) are performed for those chemicals identified as COPCs in step 1. The toxicity assessment involves researching available toxicity data for those chemicals retained for further evaluation (i.e., COPCs) and is conducted concurrently with step 2 , the exposure assessment. If toxicity data are available, step 4 is conducted and cancer risk estimates and noncancer estimates (also referred to as hazard estimates) are determined for each COPC for each complete exposure pathway. The risk/hazard estimates for each chemical are summed for each receptor to determine the cumulative potential health threat to a potential receptor exposed to site-related contamination (i.e., risk characterization). The risk characterization step also includes an evaluation of the uncertainties associated with steps 1 through 4, including a qualitative description of the inherent and sitespecific uncertainties of each component of the process. The uncertainty evaluation also presents the potential effects on the risk estimates (i.e., the calculated risk may be over- or underestimated depending on the uncertainties).

### 1.6 ORGANIZATION OF THE RISK ASSESSMENT

1.6.0.1 This report consists of seven sections, including this introduction, and seven appendices. The overall format of the report follows the four-step RA paradigm described in Section 1.5. Section 2 reports on the data evaluation step; summarizes analytical results of the field investigations; summarizes the results of the statistical calculations (including the site-tobackground comparison and derivation of exposure-point concentrations), and presents the results of the risk-based concentration screening step. The human health exposure assessment is presented in Section 3; Section 4 presents the toxicity assessment; and Section 5 provides the methodology to characterize potential human health risks, including a qualitative analysis of the uncertainties in the RA process. Section 6 presents the conclusions of the RA, while Section 7 lists references cited in this report.
1.6.0.2 Appendix A presents data summary tables. Appendix B presents an evaluation of the number of samples collected at the site, Appendix C presents a statistical analysis of the data, Appendix D presents the derivation of the particulate emissions factors (PEFs), Appendix E presents the RAGS Part D Tables, Appendix F presents the risk characterization tables, and Appendix G presents homegrown vegetable intake parameters.

[^2]
# SECTION 2 <br> DATA EVALUATION AND IDENTIFICATION OF CHEMICALS OF POTENTIAL CONCERN 

2.0.0.1 The first step of the RA process involves review of available site data that can be used in the RA. This step includes:

- Data gathering and review of existing reports;
- Development of data sets for potentially complete exposure pathways (performed in conjunction with the human health exposure assessment - discussed in Section 3 of this report); and
- Identification of COPCs to be included in the RA.
2.0.0.2 The section below describes the process for evaluating site data and developing the data sets for the RA in more detail and presents the specific COPCs that were evaluated in the RA.


### 2.1 SUMMARY OF EXISTING DATA

2.1.0.1 4835 Glenbrook Road (4835GR) is an approximately 0.5 acre private residential property within OU-3 of the SVFUDS. Over the years, numerous previous investigations have been performed at the 4835 GR property, at different times, by different parties, and with different sampling objectives and analytical parameters. These efforts include:

- 1992, EMS
- 1996, Apex
- 1999, USEPA
- 1999-present, Parsons
2.1.0.2 The existing data in summarized in Appendix A (see Table A.1) with regard to sample numbers, sample dates, and analytical parameters sampled over the various investigations. Each of the data sets associated with those investigations is presented in Appendix A as separate data summary tables (Tables A. 2 through A.10).
2.1.0.3 The locations of the samples, color coded to match the various investigations, are shown in Figure 2-1. Note that because of scale and the number of samples, not every sample is individually identified on the figure. Additionally, some samples are shown in approximate locations (for example, the blue dots are specific to the excavated grid with which they are associated, but the discrete location is approximate).
2.1.0.4 Aside from arsenic, the primary data set is for the 12-metals suite (see Table A.7). Collection of 109 soil samples for the 12 -metals suite was based on a request from American University to reflect the findings of the nearby Lot 18 as arsenic remediation and associated sampling was beginning at the 4835 Glenbrook Road property. The 12 metals represent constituents that have historically exceeded their applicable standards. The intent was to compile data while field crews were mobilized to excavate arsenic grids. Should levels of concern have been found, a potential interim action similar to that at Lot 18 would have been


Figure 2-1

## Legend

## ${ }^{\text {Test Pits }}$

(Compled with No Significant Findings) APEX Soil Probes (6 of these converted to

- Organics and Metals - Surface (APEX 1996)
- Organics and Metals - Sub-Surface (APEX 1996)
- Organics and Metals - Surface (EPA 1999)
- Organics and Metals - Sub-Surface (EPA 1999)
- Metals - Surface (PARSONS 2007-08)
- Metals - Sub-Surface (PARSONS 2007-08)
- Agent (HD,L) + ABPs + Metals
- 

HD ABPs and Arsenic - Sub-Surface (2000)
Property Boundaries
Buildings
$\square 20$ Grid
APEX Tree Removal Perimeter

Note
1.) During Test Pit Operations, Air Monitoring was performed for Mustard, Lewisite, Arsine, Phosgene, Chloropicrin, and Cyanogen Chloride with No Confirmed Detections.
2.) Additional Sampling Not Shown: 1992 EMS Investigation (Specifics Unavailable). 2000 Quadrant Sampling for HD ABPs (Composited Locations).

performed. However, the data did not indicate metals concerns beyond arsenic. Appendix B presents a statistical analysis of sample quantity for these samples.
2.1.0.5 The Edgewood Chemical and Biological Center (ECBC) analyzed chemical agent (mustard and lewisite) and their breakdown products (1,4-dithiane and 1,4-oxathiane). Those data are included in Table A.8. Note that although only four locations (orange colored) for these samples are shown on Figure 2-1, six samples are represented; test pits 40 and 49 each contained two samples.
2.1.0.6 Ten full or partial arsenic contaminated soil grids in the northern portion and three full or partial arsenic contaminated soil grids in the southern portion of the house were excavated. The grids, and associated extensions, were excavated until the arsenic concentrations in the confirmation samples were acceptable (i.e. below the remedial action level of $20 \mathrm{mg} / \mathrm{kg}$ ). All arsenic impacted soil detected at concentrations exceeding $20 \mathrm{mg} / \mathrm{kg}$ was removed from the site. This included the removal of 26 sample locations (see Appendix A) and more than 500 cubic yards of soil. In addition, Apex (1996) removed approximately 25 cubic yards of soil and one soil sample location where laboratory waste was found. For metals, only samples collected by Parsons were used in this assessment. These data were collected under the appropriate QAPP (Parsons, 2007) and meet all of the QAQC requirements for data to be used in a risk assessment. Although EMS (1992), Apex (1996), and USEPA (1999) also collected and analyzed samples from the site, those reports contain inadequate information to accurately assess the data quality and QAQC procedures. However, the data collected by Parsons focused on metals and only a very few samples were analyzed for non-metals; e.g., VOCs, SVOCs, PCBs, and PAHs. Therefore, to supplement the non-metal data from Parsons, the non-metals data from EMS (1992), Apex (1996), and USEPA (1999) was also used in this RA.
2.1.0.7 A total of 185 soil samples representative of soils still in place at the site (Table 2-1) were collected at the site. The number of samples that were analyzed for the various groups of analytes are as follows:

- Metals:

152 (Parsons data only)

- VOCs: 22
- SVOCs: 7
- Pesticides: 13
- Herbicides: 10
- PCBs: 4
- Explosives: 3
- Agent Breakdown products: 8
- TICs: 1
2.1.0.8 Note that within each analyte group, the numbers of samples analyzed for an individual chemical may vary, as the analyte list differed among the multiple investigations that have been performed at the site.

Table 2-1
Unexcavated Samples Used in the Risk Assessment 4835 Glenbrook Rd.

| Metals data (from Parsons only) | Non-Metals data |  |
| :---: | :---: | :---: |
| Sample ID | Sample ID | Collected By |
| SW-4835GB-(-170,10)SW-E(5) | 052692-1CM | EMS (1992) |
| SW-4835GB-(-170,10)SW-S | 9005 | Apex (1996) |
| SW-4835GB-(-170,10)SW-W | 9006 | Apex (1996) |
| SW-4835GB-(-190,10)-2 | 9007 | Apex (1996) |
| SW-4835GB-(-190,10)-N | 9008 | Apex (1996) |
| SW-4835GB-(-130,-30)-1.5 | 9009 | Apex (1996) |
| SW-4835GB-(-130,-30)SW-N | 9010 | Apex (1996) |
| SW-4835GB-(-130,-30)SW-W | 9011 | Apex (1996) |
| SW-4835GB-(-190,90)-2 | 9012 | Apex (1996) |
| SW-4835GB-(-190,90)SW-E(5) | 9013 | Apex (1996) |
| SW-4835GB-(-190,90)SW-S | 9014 | Apex (1996) |
| SW-4835GB-(-250,70)-2 | 9015 | Apex (1996) |
| SW-4835GB-(-250,70)SW-E | 9016 | Apex (1996) |
| SW-4835GB-(-250,70)SW-S | 9017 | Apex (1996) |
| SW-4835GB-(-150,50)-2 | 9018 | Apex (1996) |
| SW-4835GB-(-150,50)SW-E | 9019 | Apex (1996) |
| SW-4835GB-(-150,50)SW-N | G-01 | EPA (1999) |
| SW-4835GB-(-90,50)-2 | G-02 | EPA (1999) |
| SW-4835GB-(-90,50)SW-E | G-03 | EPA (1999) |
| SW-4835GB-(-90,30)SW-W(5) | OU3-SB02 | EPA (1999) |
| SW-4835GB-(-130,-30)SW-S(2.5) | OU3 MTL-4835-1 | Parsons |
| SW-4835GB-(190,90)SW-E(5)LC | OU3 MTL-4835-2 | Parsons |
| SW-4835GB-(-190,90)SW-E(5)LN | OU3 MTL-4835-3 | Parsons |
| SW-4835GB-(190,90)SW-E(5)LS | OU3 MTL-4835-4 | Parsons |
| SW-4835GB-(-190,90)SW-N(6) | OU3 MTL-4835-SB-(0-2) | Parsons |
| SW-4835GB-(-90,50)SW-N(5) | OU3 MTL-4835-SB-(2-4) | Parsons |
| SW-4835GB-(-170,10)-4 | OU3 MTL-4835-SB-(4-6) | Parsons |
| SW-4835GB-(-170,10)SW-E(5)LC-4 | SW-4835GB-01 (assoc w/TP-17) | Parsons |
| SW-4835GB-(-170,10)SW-E(5)LS | SW-4835GB-04 (assoc w/ TP-40) | Parsons |
| SW-4835GB-(-170,10)SW-S3.5 | SW-4835GB-02 (assoc w/ TP-40) | Parsons |
| SW-4835GB-(-170,10)-SW-W3.5 | SW-4835GB-TP56-001 (assoc w/ TP-56) | Parsons |
| SW-4835GB-(-190,10)SW-E(7) | SW-4835GB-TP49-001 (assoc w/ TP-49) | Parsons |
| SW-4835GB-(-190,90)SW-N(6)LC | SW-4835GB-16 (assoc w/ TP-49) | Parsons |
| SW-4835GB-(-190,90)SW-N(6)LE | 4835GB(-190,50) SW-N(5)LW-5 | Parsons |
| SW-4835GB-(-150,50)SW-S(8) |  |  |
| SW-4835GB-(-90,50)SW-N(5)LE |  |  |
| SW-4835GB(-90,50)-SW-N(5)LE2.5 |  |  |
| SW-4835GB(-90,50)-SW-N(5)LC-3 |  |  |
| SW-4835GB(-150,50)-SW-S(8)LE |  |  |
| SW-4835GB(-150,50)-SW-S(8)LC-3 |  |  |
| SW-4835GB(-190,10)SW-E(7)LN |  |  |
| SW-4835GB(-190,10)SW-E(7)LC |  |  |
| SW-4835GB(-150,50)SWS(8)2.5 |  |  |
| SW-4835GB(-150,50)SWS(8)LE2.5 |  |  |
| SW-4835GB(-90,50)SWN(5)2.5 |  |  |

Table 2-1

## Unexcavated Samples Used in the Risk Assessment

 4835 Glenbrook Rd.```
Metals data (from Parsons only) Non-Metals data
SW-4835GB-(-170,50)
SW-4835GB-(-150,10)
SW-4835GB-(-150,30)-2
SW-4835GB-(-150,-10)SW-E
SW-4835GB-(-170,30)-4
SW-4835GB-(-150,-10)-2
SW-4835GB-(-190,50)-5
SW-4835GB-(-170,-10)-3
SW-4835GB-(-190,50)SW-N(5)
SW-4835GB-(-190,50)SW-S(5)
SW-4835GB-(-170,30)SW-E
SW-4835GB-(-170,30)SW-E-3.5
SW-4835GB-(-150,30)SW-E(5)LN
SW-4835GB-(-150,30)SW-W(5)
SW-4835GB-(-170,10)SW-N
4835GB-(-190,30)-5
4835GB-(-190,30)-SW-N(4.5)
4835GB-(-190,30)-SW-N
4835GB-(-170,30)SW-S(5)-3.5
4835GB-(-170,30)SW-S(5)LW
4835GB-(-150,30)SW-W(5)LC
SW-4835GB-(190,50)SW-S(5)LC
SW-4835GB-(170,30)SW-S(5)-LC5
SW-4835GB(-170,30)SW-S(5)LW4.5
SW-4835GB(-170,30)SW-S(5)-4.5
SW-4835GB(-150,10)SW-W(10)LC3
SW-4835GB(-150,-10)SW-W(10)LC4
4835GB(-150,-10)SW-W(10)LS-2.5
4835GB(-150,-10)SW-W(10)-2.5
4835GB(-190,50)-SW-N(5)-4.5
4835GB(-190,50)-SW-N(5)LC
4835GB(-190,50)SW-S(5)-4.5
4835GB(-170,-10)SW-S-3
SW-4835GB-(-190,50)SWN(5)LW(5)
SW-4835GB-(-190,50)SWN(5)LW(5)-4.5
SW-4835GB-(-190,50)SWN(5)LW(5)LN
SW-4835GB-(-190,50)SWN(5)LW(5)LN-4.5
SW-4835GB-(-190,50)SWN(5)LW(5)LE
SW-4835GB-(-190,50)SWN(5)LW(5)LE-4.5
SW-4835GB-(-170,10)SWN(5)-3.5
SW-4835(-170,10)SWN(5)LC5
SW-4835(-170,10)SWN(5)-4.5
SW-4835(-170,10)SWN(5)LE-4.5
SW-4835(-170,10)SWN(5)LW-4.5
(-170,30)SW-S(5)LE-4.5
(-150,30)SW-E(5)LN-2.5
```

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Table 2-1
Unexcavated Samples Used in the Risk Assessment 4835 Glenbrook Rd.

| Metals data (from Parsons only) | Non-Metals data |
| :---: | :---: |
| (-150,30)SW-E(5)LC-3.0 |  |
| SW-4835GB(-90,30)-4 |  |
| SW-4835GB-(90,30)-SW-W(15)-3.5 |  |
| SW-4835GB-(90,30)-SW-W(15)-0.5 |  |
| SW-4835GB-(90,30)-SW-W(15)LE-4.0 |  |
| SW-4835GB-01 |  |
| SW-4835GB-04 |  |
| 4835GB(-190,50)-SW-N(5)LW-5 |  |
| OU3-MTL-4835(-100,0) |  |
| OU3-MTL-4835(-100,100) |  |
| OU3-MTL-4835(-100,120) |  |
| OU3-MTL-4835(-100,140) |  |
| OU3-MTL-4835(-100,20) |  |
| OU3-MTL-4835(-100,80) |  |
| OU3-MTL-4835(-120,100) |  |
| OU3-MTL-4835(-120,120) |  |
| OU3-MTL-4835(-120,140) |  |
| OU3-MTL-4835(-140,100) |  |
| OU3-MTL-4835(-140,120) |  |
| OU3-MTL-4835(-140,140) |  |
| OU3-MTL-4835(-160,100) |  |
| OU3-MTL-4835(-160,120) |  |
| OU3-MTL-4835(-160,140) |  |
| OU3-MTL-4835(-180,120) |  |
| OU3-MTL-4835(-180,140) |  |
| OU3-MTL-4835(-200,120) |  |
| OU3-MTL-4835(-200,140) |  |
| OU3-MTL-4835(-220,120) |  |
| OU3-MTL-4835(-220,140) |  |
| OU3-MTL-4835(-240,120) |  |
| OU3-MTL-4835(-240,140) |  |
| OU3-MTL-4835(-260,120) |  |
| OU3-MTL-4835(-260,140) |  |
| OU3-MTL-4835(-280,120) |  |
| OU3-MTL-4835(-320,0) |  |
| OU3-MTL-4835(-340,0) |  |
| OU3-MTL-4835 $(280,140)$ |  |
| OU3-MTL-4835-(-100,-20) |  |
| OU3-MTL-4835-(-100,-40) |  |
| OU3-MTL-4835-(-120,-20) |  |
| OU3-MTL-4835-(-120,-40) |  |
| OU3-MTL-4835-(-120,0) |  |
| OU3-MTL-4835-(-140,-40) |  |
| OU3-MTL-4835-(-140,0) |  |
| OU3-MTL-4835-(-160,80) |  |
| OU3-MTL-4835-(-180,100) |  |

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Table 2-1

## Unexcavated Samples Used in the Risk Assessment

 4835 Glenbrook Rd.| Metals data (from Parsons only) | Non-Metals data |
| :--- | :--- |
| OU3-MTL-4835-(-180,60) |  |
| OU3-MTL-4835-(-180,80) |  |
| OU3-MTL-4835-(-200,60) |  |
| OU3-MTL-4835-(-200,80) |  |
| OU3-MTL-4835-(-220,100) |  |
| OU3-MTL-4835-(-220,40) |  |
| OU3-MTL-4835-(-220,60) |  |
| OU3-MTL-4835-(-220,80) |  |
| OU3-MTL-4835-(-240,100) |  |
| OU3-MTL-4835-(-240,60) |  |
| OU3-MTL-4835-(-240,80) |  |
| OU3-MTL-4835-(-260,100) |  |
| OU3-MTL-4835-(-280,100) |  |
| OU3-MTL-4835-(-300,0) |  |

### 2.2 IDENTIFICATION OF CHEMICALS OF POTENTIAL CONCERN

2.2.0.1 COPCs were identified from the 185 samples representative of soils remaining in place at the site (Tables 2-2 and 2-3). The data from these samples were screened as follows to identify COPCs (USACE, 2009):

- Essential nutrients were removed from further consideration. Essential nutrients include calcium, sodium, iron, potassium, magnesium, iodine, chloride, and phosphorus (USEPA, 1989).
- For non-metals (excluding PAHs), the maximum detected concentration (from up to 10 feet bgs) of each chemical in soil was compared to the USEPA residential Regional Screening Levels (RSLs; USEPA, 2009d). For carcinogens, the RSL is protective of a risk level of $1 \times 10^{-6}$. For noncarcinogens, the RSL is protective of a hazard quotient of 1 . To account for potential cumulative effects, the RSLs for non-carcinogens were divided by 10 to be protective of a hazard quotient of 0.1 . Only chemicals that exceed the RSLs were retained as COPCs.
- For metals and PAHs, the maximum detected concentration was compared to the greater of the residential RSL (as described above) and the background Upper Tolerance Limit (UTL, or the upper $95^{\text {th }}$ percentile with $95 \%$ confidence), as established in the Background Soil Sampling Report for SVFUDS (USACE, 2008A). Metals and PAHs were eliminated as COPCs if the maximum detected concentration was less than the greater of the background UTL or RSL. Comparisons to background UTLs, to determine which metals are elevated over background, are consistent with USEPA (1989b, 1992c, 2006b, 2009b, c) guidance.
2.2.0.2 The 3 organics below and iodine pentafluoride (as iodate) that were detected in soils at the site do not have RSLs (USEPA, 2009d):
- (+)-Cycloisosativene
- 1,2,3,4-Tetrahydro-1,6-dimethyl-4-(1-methylethyl)-naphthalene
- E-11,13-Tetradecadien-1-ol
2.2.0.3 Additionally, the metal tellurium was detected in soils at concentrations exceeding its background UTL. However, there is no RSL (USEPA, 2009d) for tellurium. Further, there is no toxicity data for these 5 chemicals from the hierarchy of sources listed in SECTION 4. Therefore, these chemicals cannot be quantitatively evaluated in this risk assessment and were not identified as COPCs.
2.2.0.4 Following this procedure, the following 8 COPCs were identified (Tables 2-2 and 23):
- Aluminum;
- Cobalt;
- Copper;
- Manganese;
- Mercury;
- Nickel;
- Thallium; and
- Vanadium.

Table 2-2
COPC Selection for Metals and PAHs
4835 Glenbrook Road
Spring Valley, Washington, D.C.

| Metal | Site |  |  |  |  |  |  | $\begin{gathered} \text { Background } \\ \text { UTL } \\ (\mathrm{mg} / \mathrm{kg}) \end{gathered}$ | Residential PRG $^{3}$ (mg/kg) | Screening Level ${ }^{4}$ ( $\mathrm{mg} / \mathrm{kg}$ ) | Maximum Detect <br> Greater Than Screening Level? |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | Sample Size <br> (-) | Number of NDs (-) | Frequency of Detection $(-)$ | Arithmetic Average of Detected Concentrations ( $\mathrm{mg} / \mathrm{kg}$ ) | Range of Detection Limits ${ }^{1}$ $(\mathrm{mg} / \mathrm{kg})$ | $\begin{gathered} \hline \text { Minimum } \\ \text { Detect } \\ (\mathrm{mg} / \mathrm{kg}) \end{gathered}$ | Maximum Detect ( $\mathrm{mg} / \mathrm{kg}$ ) |  |  |  |  |
| Aluminum | 97 | 1 | 99\% | 24,177 | 18,600 | 8,960 | 55,900 | 19,100 | 7,700 | 19,100 | Yes |
| Anthracene | 6 | 5 | 17\% | 0.052 | ND - 0.4 | 0.052 | 0.052 | 0.51 | 1,700 | 1,700 | No |
| Antimony | 99 | 34 | 66\% | 0.96 | 0.53-56.3 | 0.25 | 3.8 | 5.2 | 3.1 | 5.2 | No |
| Arsenic | 151 | 0 | 100\% | 9.27 | NA | 0.69 | 19.9 | 12.6 | 20 | 20 | No |
| Barium | 99 | 0 | 100\% | 90.84 | NA | 18.2 | 254 | 172 | 1,500 | 1,500 | No |
| Benzo(a)anthracene | - | 5 | 17\% | 0.11 | ND - 0.4 | 0.11 | 0.11 | 0.36 | 0.15 | 0.36 | No |
| Benzo(a)pyrene | 6 | 5 | 17\% | 0.083 | ND - 0.4 | 0.083 | 0.083 | 0.40 | 0.015 | 0.40 | No |
| Benzo(b)fluoranthene | 6 | 5 | 17\% | 0.072 | ND - 0.4 | 0.072 | 0.072 | 0.37 | 0.15 | 0.37 | No |
| Benzo(k)fluoranthene | 6 | 5 | 17\% | 0.092 | ND - 0.4 | 0.092 | 0.092 | 0.37 | 1.5 | 1.50 | No |
| Beryllium | 3 | 0 | 100\% | 1.01 | NA | 0.73 | 1.3 | 1.90 | 16 | 16 | No |
| Cadmium | 99 | 66 | 33\% | 0.32 | 0.025-5.2 | 0.037 | 0.92 | 2.36 | 7 | 7 | No |
| Chromium | 1 | 0 | 100\% | 448 | NA | 448 | 448 | 51.3 | 12,000 | 12,000 | No |
| Chrysene | 6 | 5 | 17\% | 0.1 | ND - 0.4 | 0.1 | 0.1 | 0.40 | 15 | 15 | No |
| Cobalt | 3 | 0 | 100\% | 28 | NA | 18.4 | 42 | 17.80 | 2.3 | 17.80 | Yes |
| Copper | 99 | 0 | 100\% | 78.73 | NA | 16.2 | 444 | 49.65 | 310 | 310 | Yes |
| Fluoranthene | 5 | 2 | 60\% | 0.10 | ND - 0.4 | 0.005 | 0.23 | 0.70 | 230 | 230 | No |
| Lead | 99 | 7 | 93\% | 14.36 | 4.3-13.8 | 2.9 | 67.7 | 194 | 400 | 400 | No |
| Manganese | 99 | 1 | 99\% | 670.44 | 1290.00 | 133 | 4,110 | 968 | 180 | 968 | Yes |
| Mercury | 99 | 31 | 69\% | 0.12 | . 001 - . 12 | 0.013 | 0.83 | 0.25 | 0.78 | 0.78 | Yes |
| Nickel | 99 | 0 | 100\% | 66.05 | NA | 12.3 | 345 | 33.5 | 150 | 150 | Yes |
| Phenanthrene ${ }^{5}$ | 6 | 5 | 17\% | 0.22 | ND - 0.4 | 0.22 | 0.22 | 0.41 | 170 | 170 | No |
| Pyrene | 6 | 2 | 67\% | 0.144 | ND - 0.4 | 0.048 | 0.24 | 0.63 | 170 | 170 | No |
| Selenium | 3 | 1 | 67\% | 0.71 | 5.7 | 0.59 | 0.83 | 1.20 | 39 | 39 | No |
| Silver | 3 | 2 | 33\% | 0.12 | 0.91-0.91 | 0.12 | 0.12 | 0.87 | 39 | 39 | No |
| Strontium | 3 | 0 | 100\% | 19.37 | NA | 14.5 | 26.1 | 53.0 | 4,700 | 4,700 | No |
| Tellurium | 3 | 0 | 100\% | 3.77 | NA | 2.2 | 6.6 | 5.0 | 39.11 | 39.11 | No |
| Thallium | 98 | 64 | 35\% | 1.43 | 0.6-23.4 | 0.55 | 8.7 | 2.2 | 0.6 | 2.2 | Yes |
| Tin | 3 | 2 | 33\% | 14.6 | 1.4-4.6 | 14.6 | 14.6 | 8.4 | 4,700 | 4,700 | No |
| Titanium | 3 | 0 | 100\% | 614.67 | NA | 325 | 867 | 2,690 | NA | 2,690 | No |
| Vanadium | 99 | 0 | 100\% | 100.83 | NA | 33.2 | 345 | 75.5 | 39 | 75.5 | Yes |
| Zinc | 99 | 0 | 100\% | 70.62 | NA | 31.7 | 180 | 158 | 2,300 | 2,300 | No |
| Zirconium | 3 | 1 | 67\% | 12.9 | 16.9 | 12.2 | 13.6 | 48.3 | NA | 48 | No |

1 - For the NDs
2 - All background UTLs are from USACE (2008a), except for benzo(a)pyrene and pyrene, which are from USACE (2009)
3 - The residential PRGs listed here are the lesser of the cancer-based and non-cancer based 2009 USEPA Regional Screening Levels (RSL), except for arsenic and tellurium. Note that non-
 PRG is from a toxicological literature review (USACE, 2008b).
4 - The greater of the background UTL and the residential PRG
Definitions:
PRG - Preliminary Remediation goal
NA - Not Applicable
RSL - USEPA (2009d) Regional Screening Levels
UTL - Upper tolerance limit

## Table 2-3

COPC Selection for VOCs
4835 Glenbrook Road
Spring Valley, Washington, D.C.

| Chemical | Site |  |  |  |  |  |  | $\begin{gathered} \text { Residential } \\ \text { PRG }^{2} \\ (\mathrm{mg} / \mathrm{kg}) \\ \hline \end{gathered}$ | Maximum Detect Greater Than Screening Level? |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | Sample Size (-) | Number of NDs $\qquad$ (-) | Frequency of Detection (-) | Arithmetic Average of Detected Concentrations (mg/kg) | Range of Detection Limits ${ }^{1}$ $(\mathrm{mg} / \mathrm{kg})$ | Minimum Detect ( $\mathrm{mg} / \mathrm{kg}$ ) | ```Maximum Detect (mg/kg)``` |  |  |
| Acetone | 6 | 5 | 17\% | 0.045 | ND - 0.012 | 0.045 | 0.045 | 6,100 | No |
| bis(2-Ethylhexyl)phthalate | 6 | 3 | 50\% | 0.052 | ND - 0.4 | 0.044 | 0.067 | 35 | No |
| alpha-Chlordane | 3 | 1 | 67\% | 0.00485 | ND - ND | 0.0018 | 0.0079 | 1.6 | No |
| gamma-Chlordane | 3 | 1 | 67\% | 0.00515 | ND - ND | 0.0019 | 0.0084 | 1.6 | No |
| Chloroform | 6 | 5 | 17\% | 0.01 | ND - 0.012 | 0.01 | 0.01 | 0.3 | No |
| (+)-Cycloisosativene | 1 | 0 | 100\% | 0.56 | NA | 0.56 | 0.56 | NA | NA |
| 4,4'-DDT | 13 | 12 | 8\% | 0.0031 | ND - 0.1 | 0.0031 | 0.0031 | 1.7 | No |
| 1,3-Dichlorobenzene ${ }^{3}$ | 20 | 19 | 5\% | 0.0015 | ND - 0.380 | 0.0015 | 0.0015 | 2.6 | No |
| 1,4-Dichlorobenzene | 20 | 19 | 5\% | 0.0016 | ND - 0.380 | 0.0016 | 0.0016 | 2.6 | No |
| Di-n-butylphthalate | 5 | 4 | 20\% | 0.079 | ND - 0.38 | 0.079 | 0.079 | 610 | No |
| Fluoride | 2 | 0 | 100\% | 9.5 | NA | 8 | 11 | 470 | No |
| Heptachlor epoxide | 13 | 12 | 8\% | 0.0023 | ND - 0.1 | 0.0023 | 0.0023 | 0.053 | No |
| Iodine Pentafluoride (as lodate ${ }^{\text {' }}$ | 2 | 0 | 100\% | 82.5 | NA | 55 | 110 | NA | NA |
| p-Isopropyltoluene ${ }^{4}$ | 16 | 15 | 6\% | 0.004 | 0.005-0.005 | 0.004 | 0.004 | 220 | No |
| Methylene chloride | 21 | 19 | 10\% | 0.0378 | ND - . 001 | 0.0014 | 0.074 | 11 | No |
| Naphthalene, 1,2,3,4-tetrahydrı | 1 | 0 | 100\% | 0.24 | NA | 0.24 | 0.24 | NA | NA |
| Perchlorate | 2 | 1 | 50\% | 0.00174 | 0.002 | 0.00174 | 0.00174 | 5.5 | No |
| 1,1,2,2-Tetrachloroethane | 20 | 19 | 5\% | 0.38 | ND - 0.012 | 0.38 | 0.38 | 0.59 | No |
| E-11,13-Tetradecadien-1-ol | 1 | 0 | 100\% | 0.14 | NA | 0.14 | 0.14 | NA | NA |
| Thiodiglycol | 8 | 4 | 50\% | 0.9555 | 0.575-0.61 | 0.792 | 1.19 | 39.1 | No |
| Toluene | 21 | 20 | 5\% | 0.002 | ND - 0.13 | 0.002 | 0.002 | 500 | No |
| 2,4,5-TP (silvex) | 10 | 9 | 10\% | 0.013 | ND - ND | 0.013 | 0.013 | 49 | No |
| 1,1,2-Trichloroethane | 20 | 19 | 5\% | 0.32 | ND - 0.012 | 0.32 | 0.32 | 1.1 | No |
| Xylenes (Total) | 21 | 20 | 5\% | 0.0027 | ND - 0.015 | 0.0027 | 0.0027 | 60 | No |

1 - For the NDs
2 - The residential PRGs listed here are the lesser of the cancer-based and non-cancer based April 2009 USEPA Regional Screening Levels (RSL), except for thiodiglycol, which is from "Remedial Investigation Report for the Operation Safe Removal Formerly Used Defense Site, Washington, D.C." (Parsons, 1995). Note that non-cancer RSLs were divided by 10 to be protective of an HQ of 0.1.
3 - No RSL; the RSL for 1,4-dichlorobenzene was used.
4 - No RSL; the RSL for cumene (isopropylbenzene) was used. Also listed as "1-Methyl-4-(1-methylethyl)benzene" in previous reports.
Definitions:

$$
\begin{aligned}
& \text { NA - Not Applicable } \\
& \text { ND - Non-detect } \\
& \text { RSL - USEPA (2009d) Regional Screening Levels } \\
& \text { PRG - Preliminary Remediation goal }
\end{aligned}
$$

2.2.0.5 Appendix E contains the RAGS Part D tables presenting the occurrence, distribution, and selection of COPCs, as well as the EPC summary, for both $0-2$ and $0-10 \mathrm{ft}$ bgs.

### 2.3 STATISTICAL EVALUATION OF DATA

2.3.0.1 The $95 \%$ Upper Confidence Limit of the mean ( $95 \%$ UCL) of each COPC can be used to estimate the concentration of a contaminant that a receptor would be exposed to over a length of time. This exposure point concentration (EPC) can then be used to estimate risk. All UCLs were calculated using the latest version of ProUCL from USEPA (2009b,c); i.e., ProUCL v4.00.04. Refer to the ProUCL User's and Technical Guides (USEPA 2009b,c) for a detailed discussion of the statistical methods that it uses. Criteria for the selection of the computational method, as well as the formulae for the computational methods, are provided in USEPA (2002b, 2006, 2009b,c) and are not repeated here. ProUCL uses the Kaplan-Meier method to account for non-detects in the calculation of UCLs (USEPA 2009b,c). The default of 2000 iterations was used for all bootstrapping methods. The first UCL recommended by ProUCL, based on the assumed distribution type, was used as the EPC in this risk assessment. In most cases, the recommended UCLs were calculated with a $95 \%$ confidence. For highly skewed datasets, UCLs were calculated with either 97.5 or $99 \%$ confidence, as recommended by ProUCL (USEPA $2009 \mathrm{~b}, \mathrm{c}$ ). The EPCs for reasonable maximum exposure (RME) and central tendency (CT) scenarios calculated using ProUCL are summarized in Table 2-4. RME is the $95 \%$ UCL and CT is the mean or median depending on the distribution of the data. The summary statistics of the RME and CT EPC values are presented in Table C. 1 of Appendix C. The detailed output from ProUCL is also included in Appendix C.

[^3]Table 2-4

## RME and CT Exposure Point Concentrations <br> 4835 Glenbrook Road <br> Spring Valley, Washington, D.C.

RME Exposure Point Concentrations

| RME Exposure Point Concentrations |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | Surface Soils: 0-2 ft bgs | Mixed Soils: 0-10 ft bgs | Dust in Outdoor Air ( $\mathrm{mg} / \mathrm{m}^{3}$ ) |  | Homegrown Vegetables (mg/kg) |  |
| COPC | (mg/kg) | (mg/kg) | Surface Soils | Mixed Soils | Surface Soils | Mixed Soils |
| Aluminum | 23,116 | 25,533 | 7.15E-06 | $7.90 \mathrm{E}-06$ | - | - |
| Cobalt | 42 | 42 | $1.30 \mathrm{E}-08$ | $1.30 \mathrm{E}-08$ | - | - |
| Copper | 79 | 108 | $2.45 \mathrm{E}-08$ | $3.33 \mathrm{E}-08$ | $1.09 \mathrm{E}+01$ | 1.23E+01 |
| Manganese | 604 | 773 | 1.87E-07 | $2.39 \mathrm{E}-07$ | - | - |
| Mercury | 0.15 | 0.12 | $4.51 \mathrm{E}-11$ | $3.72 \mathrm{E}-11$ | - | - |
| Nickel | 74 | 72 | $2.28 \mathrm{E}-08$ | $2.23 \mathrm{E}-08$ | $2.70 \mathrm{E}+00$ | $2.65 \mathrm{E}+00$ |
| Thallium | 1.09 | 1.35 | $3.37 \mathrm{E}-10$ | $4.16 \mathrm{E}-10$ | - | - |
| Vanadium | 94 | 109 | 2.92E-08 | 3.36E-08 | - | - |


| CT Exposure Point Concentrations |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | Surface Soils: 0-2 ft bgs | Mixed Soils: 0-10 ft bgs | Dust in Outdoor Air ( $\mathrm{mg} / \mathrm{m}^{3}$ ) |  | Homegrown Vegetables (mg/kg) |  |
| COPC | (mg/kg) | (mg/kg) | Surface Soils | Mixed Soils | Surface Soils | Mixed Soils |
| Aluminum | 21,000 | 24,020 | $6.50 \mathrm{E}-06$ | $7.43 \mathrm{E}-06$ | - | - |
| Cobalt | 42 | 28 | $1.30 \mathrm{E}-08$ | 8.66E-09 | - | - |
| Copper | 70 | 63 | $2.18 \mathrm{E}-08$ | $1.93 \mathrm{E}-08$ | $1.04 \mathrm{E}+01$ | $9.95 \mathrm{E}+00$ |
| Manganese | 543 | 669 | $1.68 \mathrm{E}-07$ | 2.07E-07 | - | - |
| Mercury | 0.11 | 0.10 | $3.32 \mathrm{E}-11$ | $3.08 \mathrm{E}-11$ | - | - |
| Nickel | 63.72 | 66.05 | $1.97 \mathrm{E}-08$ | $2.04 \mathrm{E}-08$ | $2.42 \mathrm{E}+00$ | $2.49 \mathrm{E}+00$ |
| Thallium | 0.97 | 1.17 | 2.99E-10 | 3.62E-10 | - | - |
| Vanadium | 83.80 | 93.70 | 2.59E-08 | 2.90E-08 | - | - |

## SECTION 3 EXPOSURE ASSESSMENT

3.0.0.1 The objective of the exposure assessment is to estimate the type and magnitude of potential exposures to COPCs at the site. The exposure assessment includes identification of potential exposure pathways, receptors, and exposure scenarios, as well as quantification of exposure. Following USEPA (1989a) guidance, exposure assessment is a three-step process involving characterization of the exposure setting, identification of exposure pathways, and quantification of exposure. To complete these three steps, it is important to 1 ) finalize the CSM; 2) estimate exposure-point concentrations (EPC); 3) determine exposure assumptions; and 4) quantitatively estimate exposure.
3.0.0.2 The following sections present the human health exposure assessment conducted for 4835 Glenbrook Road. It should be noted that this RA evaluates only assumed exposures to soils per the site conceptual model.

### 3.1 CONCEPTUAL SITE MODEL

3.1.0.1 A CSM is an effective tool to define site dynamics, streamline the risk evaluation, and develop appropriate response actions. Specifically, such models are mechanisms for identifying complete exposure pathways between environmental media affected by site-related contamination and potential receptors.
3.1.0.2 The CSM (Figure 3-1) is intended to present and clarify assumptions regarding:

- Suspected sources and types of contaminants present;
- Contaminant release and transport mechanisms;
- Affected media (e.g., soil);
- An exposure or contact point with the contaminated medium (e.g., direct contact with soil);
- An exposure route for chemical intake by a receptor (e.g., dermal uptake); and
- Potential receptors that could contact site-related contaminants in affected media under current or future land use scenarios.
3.1.0.3 Designation of an exposure pathway as complete indicates that human exposure is possible but does not necessarily mean that exposure will occur, nor that exposure will occur at the levels estimated here. When any one of the factors listed above is missing in a pathway, it is considered to be incomplete. Incomplete exposure pathways do not pose a potential risk and were not evaluated in this risk assessment.
3.1.0.4 CSMs are dynamic tools that can be updated as necessary. For example, if changes in site conditions occur, or additional site characterization information is collected, the CSM can be revised to more accurately reflect the most current information. Understanding site conditions and land uses helps to accurately identify potential receptors under current and likely future scenarios, as well as the most appropriate corrective action(s), if necessary.

3.1.0.5 In addition, a potential receptor evaluation should consider criteria such as:
- Current and future land use on and near the site;
- Zoning status and/or deed restrictions of the site and adjacent properties;
- Current and future access to the site and to the affected media;
- Existing and/or planned exposure controls (e.g., engineered containment structures);
- Present and planned site activities;
- Extent that the site is developed and vegetated; and
- Potential for soils to be disturbed (e.g., excavation at the site, installation of a swimming pool, digging trenches for utility lines, etc.).
3.1.0.6 Potential human receptors are defined as individuals who may be exposed to siterelated contaminants in environmental media at a site. Consistent with USEPA (1989a, 1995a) guidance, current and reasonably anticipated land uses were considered in the receptor selection process.
3.1.0.7 Based on previous investigations (EMS, 1992; APEX, 1996; USACE, 2000, 2009; Parsons, 2002) the observations and reasonable assumptions for the potential human receptors for 4835 Glenbrook Road are as follows:
- Current Receptors - The 4835 Glenbrook Road property is designated as the residence for the AU President, but is currently vacant and used periodically for university functions (e.g., meetings and parties). Future land use of the 4835 Glenbrook Road property is expected to be residential. In addition to occasional university functions, current land use includes visits to the site, such as those that would occur by outdoor workers (e.g., routine landscaping activities).
- Future Receptors - The 4835 Glenbrook site is not currently used for residential purposes. However, a residence is located on the lot and it is anticipated that the lot will be returned to residential use. Therefore, the residential exposure scenario is evaluated. Additionally, future receptors would include the outdoor worker as indicated above, as well as a construction worker. Since it is anticipated that a future exposure scenario would include mixing of the soil column within the top 10 feet of soil, the construction worker and outdoor worker were assumed to have similar exposure parameters. Therefore, the risk assessment includes evaluation of a future outdoor worker. Conservative exposure assumptions were used for the outdoor worker so that it is anticipated to be protective of a construction worker as well.
3.1.0.8 In summary, the following receptors were selected for evaluation at the site: 1) current and future residents, and 2) current and future outdoor workers.


### 3.2 EXPOSURE PATHWAYS

3.2.0.1 USEPA (1989a) defines an exposure pathway as: "The course a chemical or physical agent takes from a source to an exposed organism. An exposure pathway describes a unique mechanism by which an individual or population is exposed to chemicals or physical agents at or originating from a site. Each exposure pathway includes a source or release from a source, an exposure point, and an exposure route. If the exposure point differs from the source, a transport/exposure medium (e.g., air) or media (in cases of intermedia transfer) is also included."
3.2.0.2 Currently, the site is an uninhabited residential lot that is occasionally used for university functions. Based on current and future land use assumptions, exposure to site-related contaminants will be evaluated for (see Section 5):

- Current and future on-site workers (i.e., groundskeepers, landscapers); and
- Future residents
3.2.0.3 The potential soil exposure routes that are evaluated here for all receptors include the following:
- Incidental ingestion;
- Dermal contact;
- Inhalation of particulates; and
- Inhalation of volatiles.
3.2.0.4 Additionally, it was assumed that residents at the site may consume home grown vegetables and that those vegetables may take up the COPCs from the soils at the site.
3.2.0.5 None of the COPCs identified at the site are classified by USEPA (1991c, 2009d) as volatiles; i.e., have a molecular weight of less than 200 and a Henry's law constant greater than 1 $\times 10^{-5} \mathrm{~atm}-\mathrm{m}^{3} / \mathrm{mole}$. Therefore, inhalation of volatiles at the site is an incomplete pathway and was not evaluated further.
3.2.0.6 Assumed exposures to two different soil depth intervals were evaluated for the receptors at the site. The current on-site worker and future resident were evaluated using an exposure interval of 0 to 2 feet bgs, to represent routine landscaping and gardening activities. Additionally, current and future residents and future on-site workers were evaluated for assumed exposures to mixed soils, 0 to 10 feet bgs. This depth interval takes into account soil mixing that may occur due to regular outdoor activities (e.g., gardening, lawn maintenance, etc.). Construction workers may encounter soil deeper than 2 feet bgs during excavation activities. Therefore, the exposure parameters selected for the outdoor worker are anticipated to be protective of construction workers, as well.


### 3.2.0.7 Each of these exposure pathways is discussed in detail below.

### 3.3 QUANTIFICATION OF EXPOSURE

3.3.0.1 Human intakes over a long-term period of exposure, called chronic daily intakes (CDIs), were calculated for each COPC identified. Intake is defined as "a measure of exposure expressed as the mass of a substance in contact with the exchange boundary per unit body weight per unit time (e.g., mg chemical/kg body weight-day)" (USEPA, 1989a). Calculation of the CDI also takes into account exposure variables (i.e., assumptions about patterns of exposure to contaminated media), and whether the chemical is a carcinogen or a noncarcinogen. The total exposure is divided by the time period of interest to obtain an average exposure over time. The averaging time is a function of the toxic endpoint: for carcinogenic effects it is the lifetime of an individual; for noncarcinogenic effects it is the exposure duration.
3.3.0.2 The following subsections provide the exposure equations for each of the designated pathways. Appendix F provides the detailed calculations using these equations for each receptor scenario.

### 3.3.1 Incidental Ingestion of Contaminants in Soil

To estimate an oral CDI for the incidental ingestion of COPCs in soil by residential receptors and on-site outdoor workers, the following equation (USEPA, 1989a) was used:

$$
\mathrm{CDI}=\frac{\mathrm{EPC} \times \mathrm{IR} \times \mathrm{FI} \times \mathrm{EF} \times \mathrm{ED} \times \mathrm{CF}}{\mathrm{BW} \times \mathrm{AT}}
$$

Where:

| CDI | $=$ Chronic daily intake $(\mathrm{mg} / \mathrm{kg}-\mathrm{d})$ |
| :--- | :--- |
| EPC | $=$ Exposure point concentration in soil $(\mathrm{mg} / \mathrm{kg})$ |
| IR | $=$ Soil ingestion rate $(\mathrm{mg} /$ day $)$ |
| FI | $=$ Fraction ingested from contaminated source (unitless) |
| EF | $=$ Exposure frequency (days/yr) |
| ED | $=$ Exposure duration $(\mathrm{yrs})$ |
| CF | $=$ Conversion factor, $1 \mathrm{E}-06(\mathrm{~kg} / \mathrm{mg})$ |
| BW | $=$ Body weight $(\mathrm{kg})$ |
| AT | $=$ Averaging time (days) |

### 3.3.2 Ingestion of home-grown vegetables

3.3.2.1 USEPA Region III (2008) guidance states "All bioaccumulative compounds need to be assessed in the food chain exposure evaluation". The list of compounds that the Region III BTAG considers to be bioaccumulative is on Table 4-2 in Bioaccumulative Testing and Interpretation for the Purpose of Sediment Quality Assessment, Status and Needs, EPA-823-R-$00-001$, February 2000." Of the COPCs identified at the site, only copper and nickel are considered by the Region III BTAG to be bioaccumulative. Therefore, exposures from the ingestion of these COPCs in homegrown vegetables were assessed. The other COPCs were not evaluated for exposures via the ingestion of homegrown vegetables.
3.3.2.2 To estimate an oral CDI for the ingestion of COPCs in home-grown vegetables by residential receptors, the following equation (USEPA, 2004b) was used:

$$
\mathrm{CDI}=\frac{\mathrm{EPC} \times \mathrm{IR}_{\mathrm{veg}} \times \mathrm{DW} \times(1-\mathrm{PL}) \times \mathrm{EF} \times \mathrm{ED} \times \mathrm{CF}}{\mathrm{AT}}
$$

| Where: |  |
| ---: | :--- |
| CDI | $=$ Chronic daily intake (absorbed dose) $(\mathrm{mg} / \mathrm{kg}-\mathrm{d})$ |
| EPC | $=$ Exposure point concentration in vegetables $(\mathrm{mg} / \mathrm{kg})$ |
| $\mathrm{IR}_{\mathrm{veg}}$ | $=$ Home-grown vegetable ingestion rate $(\mathrm{mg} / \mathrm{kg}-\mathrm{day})$ |
| DW | $=$ Dry weight percentage $(\%)$ |
| PL | $=$ Preparation and cooking loss $(\%)$ |
| EF | $=$ Exposure frequency (days $/ \mathrm{yr})$ |
| ED | $=$ Exposure duration (yrs) |
| CF | $=$ Conversion factor, $1 \mathrm{E}-06(\mathrm{~kg} / \mathrm{mg})$ |
| AT | $=$ Averaging time (days) |

3.3.2.3 Note that home-grown vegetable intake rates are available on a per capita basis and on a "consumer only" basis. The "consumer only" intake rates exclude individuals that do not

[^4]consume home-grown vegetables. To be health-protective, the home-grown vegetable intake rates used here are the consumer only home-grown vegetable ingestion rates for the south from USEPA (1997a; Table 13-16). Following USEPA (2004b) guidance, the intake rates are multiplied by the average percent of individuals "consuming homegrown vegetables during the survey period." Therefore, the vegetable intake rates were calculated as follows:

- RME
o $95^{\text {th }}$ percentile consumption rate for central cities in the south $=3.7 \mathrm{~g} / \mathrm{kg}$-day
o Percent consuming: 6.63\%
o Consumption rate $=3.7 \mathrm{~g} / \mathrm{kg}$-day $\times 6.63 \%=0.245 \mathrm{~g} / \mathrm{kg}$-day or $245 \mathrm{mg} / \mathrm{kg}$-day
- CT
o $50^{\text {th }}$ percentile consumption rate for central cities in the south $=0.615 \mathrm{~g} / \mathrm{kg}$-day
o Percent consuming: 6.63\%
o Consumption rate $=0.615 \mathrm{~g} / \mathrm{kg}$-day x $6.63 \%=0.041 \mathrm{~g} / \mathrm{kg}$-day or $41 \mathrm{mg} / \mathrm{kg}$-day
3.3.2.4 Since vegetable intake rates have been provided by USEPA (1997a) in terms of wet weight, the intake rates must be converted to dry weight, as the soil and vegetable EPCs are in terms of dry weight. This is accomplished in the equation above by multiplying the vegetable ingestion rate by the average dry weight percentage of vegetables ( $15.57 \%$; see Appendix G). Additionally, the vegetable intake rates from USEPA (1997a) are for raw vegetables. To account for the weight of the food item lost in preparation, the vegetable intake rate is multiplied by the percentage lost during preparation/cooking. For homegrown vegetables, USEPA (2004b) provides a preparation loss of $12 \%$.


### 3.3.3 Dermal Contact with Contaminants in Soil

To estimate a dermal CDI for COPCs in soil, the following equation was used (USEPA, 2004a):

$$
\mathrm{CDI}=\frac{\mathrm{EPC} \times \mathrm{AF} \times \mathrm{DAF} \times \mathrm{CF} \times \mathrm{EV} \times \mathrm{EF} \times \mathrm{ED} \times \mathrm{SA}}{\mathrm{BW} \times \mathrm{AT}}
$$

Where:

| CDI | $=$ Chronic daily intake (absorbed dose) $(\mathrm{mg} / \mathrm{kg}-\mathrm{d})$ |
| :--- | :--- |
| EPC | $=$ Exposure point concentration in soil ( $\mathrm{mg} / \mathrm{kg})$ |
| AF | $=$ Soil-to-skin adherence factor $\left(\mathrm{mg} / \mathrm{cm}^{2}\right.$-day); |
| DAF | $=$ Dermal absorption fraction (unitless); and |
| CF | $=$ Conversion factor (1E-06 $\mathrm{kg} / \mathrm{mg})$. |
| EV | $=$ Event frequency (events $/ \mathrm{day})$ |
| EF | $=$ Exposure frequency (days $/ \mathrm{yr})$ |
| ED | $=$ Exposure duration (yrs) |
| SA | $=$ Skin surface area available for contact $\left(\mathrm{cm}^{2}\right)$ |
| BW | $=$ Body weight $(\mathrm{kg})$ |
| AT | $=$ Averaging time (days) |

As shown in Table 4-2, there are no DAFs for the COPCs at the site. Therefore, it was assumed here that dermal exposures to the COPCs at the site did not result in any intake.

### 3.3.4 Inhalation

Although body weight normalized CDIs (i.e., $\mathrm{mg} / \mathrm{kg}$-day) as are used to estimate intakes for ingestion and dermal absorption, current USEPA (1996a, 2002, 2009a,d) guidance does not recommend this approach for estimating inhalation exposures. Instead, current guidance (USEPA 1996a, 2002, 2009a,d) recommends that risks and hazards be estimated from exposure frequency and duration normalized air concentrations. This is presented in detail in Sections 5.1 and 5.2.

### 3.3.5 Residential Exposure

To better protect human health, exposure to carcinogenic compounds is often assumed to occur during the first 30 years of life. Thus, exposure is assumed to occur during childhood when the intake is greater and the child is more susceptible to the effects of carcinogenic compounds. Therefore, to provide a health protective risk assessment of the site, residents are evaluated separately as adults and children.

### 3.4 EXPOSURE PARAMETERS AND ASSUMPTIONS

3.4.0.1 USEPA (1992b, 1995b) typically requires two types of exposure evaluations: a reasonable maximum exposure (RME) and average, or central tendency (CT), estimate. The RME is defined as the maximum level of exposure that is reasonably expected to occur (USEPA, 1989a), whereas the CT is the typical level of exposure that is expected to occur. In accordance with USEPA (1992b) recommendations, exposure parameters were chosen with the understanding that the combination of variables for a given pathway would result in an estimate of the RME for that pathway. Under this approach, some variables may not be at their individual maximum values, but when combined with other variables they will result in estimates of the RME. This is in contrast to the historical worst-case or bounding approach in which all variables are maximized, resulting in an estimated exposure well above actual levels seen in the population. Studies of the compounding of conservatism in probabilistic risk assessments show that setting as few as two factors at RME levels or high end (e.g., near the $90^{\text {th }}$ percentile), while the remaining variables are set at less conservative, typical or CT values, results in a product of all input variables at an approximate RME level (e.g., $99^{\text {th }}$ percentile value) (Cullen, 1994). CT risk estimates were calculated using central tendency estimates for each of the exposure parameters (USEPA, 1992b, 1995b).
3.4.0.2 Generally, contact rate, exposure frequency, and exposure duration are the most sensitive parameters (i.e., most likely to drive exposure estimates). When statistical data were available, $90^{\text {th }}$ or $95^{\text {th }}$ percentile values were used for exposure duration. If distributions were not available (e.g., for workers), high-end estimates were made using best professional judgment. Typically, distributional data are not available for exposure frequency; therefore, high-end estimates have been made using available site-specific information and best professional judgment. The following subsections discuss the justification for each parameter.
Table 3-1 summarizes the RME and CT exposure parameters used to evaluate receptors at the site.

Table 3-1
RME and CT Exposure Parameters
4835 Glenbrook Road
Spring Valley, Washington, D.C.

| Parameter |  | RME Value | CT Value | Units | RME Source | CT Source |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| $\mathrm{AF}_{\text {s }}$ Soil adherence factor |  |  |  |  |  |  |
|  | Ourdoor Worker Residential | 0.07 | 0.02 | $\mathrm{mg} / \mathrm{cm}^{2}$ | USEPA (2004a) | USEPA (2004a) |
|  | adult | 0.07 | 0.01 | $\mathrm{mg} / \mathrm{cm}^{2}$ | USEPA (2004a) | USEPA (2004a) |
|  | child | 0.2 | 0.04 | $\mathrm{mg} / \mathrm{cm}^{2}$ | USEPA (2004a) | USEPA (2004a) |
| AT | Averaging time |  |  |  |  |  |
|  | Carcinogens ( $\mathrm{AT}_{\mathrm{c}}$ ) | 25,550 | 25,550 | days | Lifetime of 70 ye | (USEPA 1989a) |
|  | Noncarcinogens ( $\mathrm{AT}_{\text {nc }}$ ) |  |  |  |  |  |
|  | Ourdoor Worker Residential | 9,125 | 3,285 | days | ED $\times 365$ days/yr (USEPA 1989a) |  |
|  |  | Residential |  |  |  |  |
|  | adult | 10,950 | 3,285 | days | ED $\times 365$ days/yr (USEPA 1989a) |  |
|  | child | 2,190 | 2,190 | days | ED $\times 365$ days/yr (USEPA 1989a) |  |
| BW | Body weight |  |  |  |  |  |
|  | Outdoor Worker Residential | 70 | 70 | kg | USEPA (1997a) | USEPA (1997a) |
|  | adult | 70 | 70 | kg | USEPA (1997a) | USEPA (1997a) |
|  | child | 15 | 15 | kg | USEPA (2008) | USEPA (2008) |
| $\mathrm{C}_{\text {air }}$ | Concentration in air | chemical-specific |  | $\mathrm{ug} / \mathrm{m}^{3}$ | see Table 2-3 | see Table 2-3 |
| $\mathrm{C}_{\text {soil }}$ | Concentration in soil | chemical-specific |  | $\mathrm{mg} / \mathrm{kg}$ | see Table 2-3 | see Table 2-3 |
| DAF | Dermal absorption fraction | chemical-specific |  | unitless | see Table 4-2 | see Table 4-2 |
| DW | Dry weight | 0.16 | 0.16 | unitless | see Appendix F | see Appendix F |
| ED | Exposure duration |  |  |  |  |  |
|  | Outdoor Worker | 25 | 9 | yrs | USEPA (1989a) | USEPA (2004a) |
|  | Residential |  |  |  |  |  |
|  | adult | 30 | 9 | yrs | USEPA (1997a) | USEPA (2004a) |
|  | child | 6 | 6 | yrs | USEPA (1997a) | USEPA (2004a) |
| EF | Exposure frequency |  |  |  |  |  |
|  | Outdoor Worker | 250 | 219 | days/yr | USEPA (1991a) | USEPA (2004a) |
|  | Residential | 350 | 350 | days/yr | USEPA (1991a) | USEPA (2004a) |
| ET | Fraction of EF breathing contaminated outdoor air |  |  |  |  |  |
|  | Outdoor Worker | 1 | 0.333 | unitless | Assumes on $100 \%$ of the day spent outdoors | Assumes 7.9 hrs/day spent outdoors |
|  | Residential |  |  |  |  |  |
|  | adult | 0.0625 | 0.0625 | unitless | $1.5 \mathrm{hrs} /$ day spent outdoors (US | 1997a) |
|  | child | 0.074 | 0.074 | unitless | 1.776 hrs/day spent outdoors; | PA (2008) |
| FI | Fraction Ingested |  |  |  |  |  |
|  | Outdoor Worker | 1 | 1 | unitless | Conservative assumption | Conservative assumption |
|  | Residential adult | 1 | 1 | unitless | Conservative assumption | Conservative assumption |
|  | child | 1 | 1 | unitless | Conservative assumption | Conservative assumption |
| $\mathrm{IR}_{\text {soil }}$ | Soil ingestion rate |  |  |  |  |  |
|  | Outdoor Worker Residential | 480 | 100 | mg/day | USEPA (1997a) | USEPA (1997a) |
|  | adult | 100 | 50 | mg/day | USEPA (1997a) | USEPA (1997a) |
|  | child | 100 | 100 | mg/day | USEPA (2008) | USEPA (2008) |
| $\mathrm{IR}_{\text {veg }}$ | Homegrown vegetable intake rate |  |  |  |  |  |
|  | Residential | 245 | 41 | $\mathrm{mg} / \mathrm{kg} / \mathrm{day}$ | USEPA (1997a, 2004b) | USEPA (1997a, 2004b) |
| PEF | Particulate emissions factor | 3.23E+09 | $3.23 \mathrm{E}+09$ | $\mathrm{m}^{3} / \mathrm{kg}$ | See Appendix D | See Appendix D |
| PL | Vegetable preparation loss | 0.12 | 0.12 | unitless | USEPA (2004b) | USEPA (2004b) |
| $\mathrm{SA}_{\text {soil }}$ | Skin surface area for soil |  |  |  |  |  |
|  | Outdoor Worker Residential | 3,300 | 3,300 | $\mathrm{cm}^{2}$ | USEPA (2004a) | USEPA (2004a) |
|  | adult | 5,700 | 5,700 | $\mathrm{cm}^{2}$ | USEPA (2004a) | USEPA (2004a) |
|  | child | 2,800 | 2,800 | $\mathrm{cm}^{2}$ | USEPA (2004a) | USEPA (2004a) |

### 3.4.1 Exposure-Point Concentrations

Exposure-point concentrations (EPCs) are intended to be representative of the concentrations of chemicals in a given medium to which a receptor may be chronically exposed at the site (i.e., the exposure point). EPCs were calculated for RME and CT scenarios using ProUCL (see Section 2.3). For incidental ingestion and dermal contact with soils, the soil data collected at the site were used to calculate the EPCs, as described below. For the inhalation of COPCs in dusts and the ingestion of COPCs in homegrown vegetables, fate and transport models were used to estimate the EPCs, as described below.

### 3.4.1.1 Exposure-Point Concentrations for Airborne Fugitive Dust

3.4.1.1.1 Following USEPA (1996a, 2002) guidance, EPCs for COPCs in airborne fugitive dust should be based on soil EPCs and estimated using the following equation:

$$
\mathrm{C}_{\mathrm{air}}=\frac{\mathrm{EPC}}{\mathrm{PEF}}
$$

Where:

$$
\begin{aligned}
& \mathrm{C}_{\text {air }} \quad=\text { COPC concentration in air at the exposure point }\left(\mathrm{mg} / \mathrm{m}^{3}\right) ; \\
& \mathrm{EPC}=\text { Exposure point concentration in soil }(\mathrm{mg} / \mathrm{kg}) \\
& \mathrm{PEF}=
\end{aligned}
$$

3.4.1.1.2 The PEF relates the concentration of soil COPCs to the concentration of dust particles in the air. This calculation addresses dust generated from open sources, which is termed "fugitive" because it is not discharged into the atmosphere in a confined flow. PEF calculations include the $\mathrm{Q} / \mathrm{C}$ term (i.e., dispersion) specific to the site's size and meteorological conditions. The PEF is calculated using the following equation (USEPA 1996a, 2002):

$$
\mathrm{PEF}=\mathrm{Q} / \mathrm{C}_{\text {wind }} \times \frac{3,600 \mathrm{~s} / \mathrm{h}}{0.036 \times(1-\mathrm{V}) \times\left(\frac{\mathrm{U}_{\mathrm{m}}}{\mathrm{U}_{\mathrm{t}}}\right)^{3} \times \mathrm{F}(\mathrm{x})}
$$

Where:
$\mathrm{Q} / \mathrm{C}_{\text {wind }}=87.37 \mathrm{~g} / \mathrm{m}^{2}$-s per $\mathrm{kg} / \mathrm{m}^{3}$, based on a 0.5 acre source in Philadelphia (PA) (USEPA 2002)
$\mathrm{V} \quad=0.5$, fraction of vegetative cover (USEPA, 1996a, 2002)
$\mathrm{U}_{\mathrm{m}} \quad=4.29 \mathrm{~m} / \mathrm{s}$, mean annual wind speed in Philadelphia (PA) (USEPA, 1996a)
$\mathrm{U}_{\mathrm{t}} \quad=11.32 \mathrm{~m} / \mathrm{s}$, equivalent threshold value of windspeed at 7 m (USEPA, 1996a)
$\mathrm{F}(\mathrm{x}) \quad=0.0993$, windspeed distribution function for Philadelphia (PA) (USEPA, 1996a).
3.4.1.1.3 Using this equation results in a PEF of $3.23 \times 10^{9} \mathrm{~m}^{3} / \mathrm{kg}$ (see Appendix D). The EPCs for COPCs in dusts that were estimated using this PEF are presented in Table 2-4.

[^5]
### 3.4.1.2 Exposure-Point Concentrations for Homegrown Vegetables

3.4.1.2.1 In order to predict the concentrations of chemicals in homegrown produce at the site, screening-level bioaccumulation models were used. These models were selected from the following hierarchy of sources:

- USEPA's (2007) ecological soil screening levels (Eco-SSLs)
- Bechtel Jacobs (1998)
- USEPA's (1999b) screening level ecological risk assessment protocol for hazardous waste combustion facilities
- Baes et al. (1984)
3.4.1.2.2 The selected bioaccumulation models are shown in Table 3-2 and the EPCs calculated using these models are shown in Table 2-4.

[^6]Table 3-2
Homegrown Vegetables Bioaccumulation Factors
4835 Glenbrook Road
Spring Valley, Washington, D.C.

| COPC | Bioaccumulation Factors for Vegetables |  |
| :---: | :---: | :---: |
|  | Transfer Equation | Source |
| Copper | $\ln (\mathrm{Cp})=0.394 * \ln (\mathrm{Cs})+0.668$ | USEPA (2007) |
| Nickel | $\ln (\mathrm{Cp})=0.748 * \ln (\mathrm{Cs})-2.223$ | USEPA (2007) |
| Notes: |  |  |
| Cp - Concentration of contaminant in the homegrown vegetables Cs - Concentration of contaminants in soil |  |  |

## SECTION 4 TOXICITY ASSESSMENT

The third step of the RA is the toxicity assessment. The objective of the toxicity assessment is to weigh available evidence regarding the potential for particular chemicals to cause adverse effects in exposed individuals and to provide, where possible, an estimate of the relationship between the extent of exposure to a chemical and the increased likelihood and/or severity of adverse effects. The types of toxicity values used in risk assessment include oral reference doses (RfDs), inhalation reference concentrations (RfCs), oral slope factors (SFs), and inhalation unit risk factors (URFs). RfDs and RfCs are used to evaluate noncarcinogenic effects. SFs and URFs are used to evaluate carcinogenic effects.

### 4.1 TOXICITY VALUES FOR CARCINOGENIC EFFECTS

4.1.0.1 The SF is the toxicity value used to estimate the lifetime excess cancer risk associated with oral exposure (ingestion) to a known or suspected carcinogen (assuming a 70-year average life span). SFs are derived for those chemicals that have been shown to cause an increased incidence of tumors in either human or animal studies. Generating a dose-response relationship between tumor incidence and exposure using human epidemiologic or animal studies is used to derive the SF. This dose-response curve is then assumed to be linear at low doses (e.g., those found in situations of environmental contamination) and is used to predict tumor incidence at low exposure levels.
4.1.0.2 In this RA, the chemical-specific SFs for COPCs were used to evaluate potential carcinogenic risk due to incidental ingestion of soil and dermal exposure to individual COPCs in soil. The SF is reported in terms of risk per milligrams (of chemical) per kilogram (unit body weight) per day $(\mathrm{mg} / \mathrm{kg}-\mathrm{d})^{-1}$. In addition, a chemical specific URF was used to evaluate the potential carcinogenic risk due to inhalation of COPCs. The URF is reported in terms of risk per milligrams (of chemical) in a cubic meter of air $\left(\mathrm{mg} / \mathrm{m}^{3}\right)^{-1}$.
4.1.0.3 Following USEPA (2003, 2009a,d) guidance, SFs and URFs were obtained from the following hierarchy of primary sources:

- USEPA's IRIS (USEPA 2009e)
- USEPA's Provisional Peer Reviewed Toxicity Values (PPRTVs)
- OEHHA's (2009) Toxicity Criteria Database
- USEPA’s Health Effects Summary Tables (USEPA 1997b)
4.1.0.4 The SFs and URFs used in this evaluation are shown in Tables 4-1 and 4-2.


### 4.2 TOXICITY VALUES FOR NONCARCINOGENIC EFFECTS

4.2.0.1 For chemicals that exhibit noncarcinogenic effects, the USEPA assumes that organisms have repair and detoxification capabilities that must be exceeded by some critical concentration (threshold) before the health effect is manifested. This threshold theory assumes that the receptor can tolerate a range of exposures from just above zero to some finite value with no appreciable risk of adverse effects.

[^7]Table 4-1
Oral and Inhalation Toxicity Values 4835 Glenbrook Road
Spring Valley, Washington, D.C.

|  | SF。 | URF |  |  | RfD ${ }_{\text {o }}$ |  |  | RfC |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| COPC | (mg/kg-day) ${ }^{-1}$ | $\left(\mu \mathrm{g} / \mathrm{m}^{3}\right)^{-1}$ | Source | Date | (mg/kg-day) | Source | Date | $\left(\mu \mathrm{g} / \mathrm{m}^{3}\right)$ | Source | Date |
| Aluminum | - | - | - | - | 1.00E+00 | PPRTV | Oct-06 | $5.00 \mathrm{E}+00$ | PPRTV | Oct-06 |
| Cobalt | - | $9.00 \mathrm{E}-03$ | PPRTV | Aug-08 | $3.00 \mathrm{E}-04$ | PPRTV | Aug-08 | $6.00 \mathrm{E}-03$ | PPRTV | Aug-08 |
| Copper | - | - | - | - | $4.00 \mathrm{E}-02$ | HEAST | Jul-97 | - | - | - |
| Manganese | - | - | - | - | $1.40 \mathrm{E}-01$ | IRIS | May-09 | 5.00E-02 | IRIS | May-09 |
| Mercury | - | - | - | - | $3.00 \mathrm{E}-04$ | IRIS; 1 | May-09 | $2.00 \mathrm{E}-01$ | ATSDR | May-09 |
| Nickel | - | 2.60E-04 | OEHHA | May-09 | $2.00 \mathrm{E}-02$ | IRIS | May-09 | $9.00 \mathrm{E}-02$ | ATSDR | May-09 |
| Thallium | - | - | - | - | $8.00 \mathrm{E}-05$ | IRIS;2 | May-09 | - | - | - |
| Vanadium | - | - | - | - | 7.00E-03 | HEAST | Jul-97 | - | - | - |

## Notes:

1-Mercuric chloride used
2 - Thallium (I) sulfate used.
Definitions:
ATSDR
Agency for Toxic Substances and Disease Registry Minimal Risk Levels
Available online at: http://www.atsdr.cdc.gov/mrls/index.html
HEAST
USEPA (1997b) Health Effects Assessment Tables
IRIS
OEHHA
PPRTV
RfC
RfD
SF
SF
URF

Office of Environmental Health Hazard Assessment Toxicity Criteria Database.
Available online at: http://www.oehha.org/risk/chemicalDB/index.asp
USEPA Provisional Peer Reviewed Toxicity Values
Reference concentation
Reference dose
Slope factor
Inhalation unit risk

Table 4-2

## Dermal Toxicity Values

 4835 Glenbrook RoadSpring Valley, Washington, D.C.

| SF $_{\mathrm{d}}$ |  | RfD $_{\mathrm{d}}$ | DAF $^{\mathbf{1}}$ | OAF |  |
| :--- | :---: | :---: | :---: | :---: | :---: |
| COPC | (mg/kg-day) $^{-1}$ | $(\mathrm{mg} / \mathrm{kg}-\mathrm{day})$ | (unitless) $^{2}$ | (unitless) | Source |
| Aluminum | - | $1.00 \mathrm{E}-01$ | - | 0.1 | Bast and Borges (1996) |
| Cobalt | - | $3.00 \mathrm{E}-04$ | - | 1 | USEPA (2004a) |
| Copper | - | $1.20 \mathrm{E}-02$ | - | 0.3 | Bast and Borges (1996) |
| Manganese | - | $5.60 \mathrm{E}-03$ | - | 0.04 | USEPA (2004a) |
| Mercury | - | $2.10 \mathrm{E}-05$ | - | 0.07 | USEPA (2004a) |
| Nickel | - | $8.00 \mathrm{E}-04$ | - | 0.04 | USEPA (2004a) |
| Thallium | - | $8.00 \mathrm{E}-05$ | - | 1 | USEPA (2004a) |
| Vanadium | - | $1.82 \mathrm{E}-04$ | - | 0.026 | USEPA (2004a) |

Notes:
1 - From USEPA (2004a).

Definitions:

| DAF | Dermal absorption fraction from soil |
| :--- | :--- |
| OAF | Oral absorption fraction |
| $\mathrm{RfD}_{\mathrm{d}}$ | Dermal reference dose, which equals $\mathrm{RfD}_{o} \times \mathrm{OAF}$ |
| $\mathrm{RfD}_{o}$ | Oral reference dose |
| $\mathrm{SF}_{\mathrm{d}}$ | Dermal slope factor, which equals $\mathrm{SF}_{o} / \mathrm{OAF}$ |
| $\mathrm{SF}_{\mathrm{o}}$ | Oral slope factor |

4.2.0.2 Toxicity values for chemicals exhibiting noncarcinogenic effects are usually developed using RfDs. In general, the RfD provides an estimate of an average daily exposure to an individual (including sensitive individuals) below which there will not be an appreciable risk of adverse health effects. The RfD is derived using uncertainty factors (e.g., to adjust from animals to humans and to protect sensitive populations) to ensure that it is unlikely to underestimate the potential for adverse noncarcinogenic effects. The purpose of the RfD is to provide a benchmark against which the sum of other doses (i.e., those projected from human exposure to various environmental conditions) might be compared. Doses that are significantly higher than the RfD may indicate that an inadequate margin of safety could exist for exposure to that substance and that an adverse health effect could occur. The RfD is expressed in terms of $\mathrm{mg} / \mathrm{kg}$-d. In addition, a chemical specific Reference Concentration (RfC) was used to evaluate the potential noncarcinogenic effects due to inhalation of COPCs. The RfC is reported in terms $\mathrm{mg} / \mathrm{m}^{3}$.
4.2.0.3 Following USEPA (2003, 2009a,d) guidance, RfDs and RfCs were obtained from the following hierarchy of primary sources:

- USEPA's IRIS (USEPA 2009e)
- USEPA's Provisional Peer Reviewed Toxicity Values (PPRTVs)
- Agency for Toxic Substances and Disease Registry's (ATSDR) Minimal Risk Levels (MRLs) (ATSDR 2009)
- OEHHA's (2009) Toxicity Criteria Database
- USEPA's Health Effects Summary Tables (USEPA 1997b)
4.2.0.4 The RfDs and RfCs used in this evaluation are shown in Tables 4-1 and 4-2.

[^8]
## SECTION 5 <br> RISK CHARACTERIZATION

The final step in the RA process is risk characterization. The purpose of the risk characterization step is to 1) review the results from the exposure and toxicity assessments; 2) quantitatively estimate the potential for cancer (i.e., risk) and noncancer (i.e., hazard) effects; and 3) assess and discuss uncertainties associated with each of the aforementioned steps. To characterize potential noncarcinogenic effects, estimated exposure levels were compared with their respective toxicity values. To characterize potential carcinogenic effects, the incremental probability of an individual developing cancer over a lifetime was calculated from the estimated exposure levels and chemical-specific dose/response information (i.e., carcinogenic toxicity factors). Cancer risk (for carcinogens) and hazard quotient (HQ; for noncarcinogens) estimates were calculated as described below for each COPC.

### 5.1 CARCINOGENIC EFFECTS

5.1.0.1 For carcinogens, risks are estimated as the incremental probability of an individual developing cancer over a lifetime (assumed to be 70 years) as a result of exposure to the potential carcinogen (i.e., incremental or excess individual lifetime cancer risk). For example, an excess lifetime cancer risk of $1 \times 10^{-6}$ indicates an individual has a one-in-one-million probability of developing cancer over a lifetime as a result of exposure to a specific COPC. Carcinogenic risk probabilities were estimated by multiplying the exposure level calculated for each exposure route by the corresponding cancer toxicity value (i.e., SF or URF) (USEPA 1989a, 1996a, 2004a, 2009a) as follows:

$$
\begin{aligned}
& \text { Risk }_{\text {oral }}=\mathrm{CDI}_{\text {oral }} \times \mathrm{SF}_{\mathrm{o}} \\
& \text { Risk }_{\text {dermal }}=\mathrm{CDI}_{\text {dermal }} \times \mathrm{SF}_{\mathrm{d}} \\
& \text { Risk }_{\text {inhalation }}=\frac{\mathrm{C}_{\text {air }} \times \mathrm{EF} \times \mathrm{ED} \times \mathrm{ET} \times \mathrm{URF} \times \mathrm{CF}}{\mathrm{AT} \times 365 \text { days } / \text { year }}
\end{aligned}
$$

where:

| AT | = Averaging time (years) |
| :---: | :---: |
| $\mathrm{Cair}_{\text {air }}$ | $=$ COPC concentration in airborne dust or outdoor air ( $\mathrm{mg} / \mathrm{m}^{3}$ ) |
| $\mathrm{CDI}_{\text {oral,dermal }}$ | $=$ Chronic daily intake for each COPC via pathway indicated ( $\mathrm{mg} / \mathrm{kg}$-day) |
| CF | $=$ Conversion factor ( $1,000 \mu \mathrm{~g} / \mathrm{mg}$ ) |
| ED | $=$ Exposure duration (years) |
| EF | = Exposure frequency (days/year) |
| ET | = Exposure time; i.e., the fraction of the day spent at the site (unitless) |
| OAF | = Oral absorption factor (unitless) |
| Risk | $=$ Incremental or excess individual lifetime cancer risk for each COPC (unitless) |
| $\mathrm{SF}_{\text {o }}$ | $=$ Chemical specific oral slope factor $\left((\mathrm{mg} / \mathrm{kg} \text {-day })^{-1}\right)$ |
| $\mathrm{SF}_{\mathrm{d}}$ | $=\mathrm{SF}_{0} / \mathrm{OAF}$ |
| URF | $=$ Chemical specific inhalation unit risk factor $\left(\left(\mu \mathrm{g} / \mathrm{m}^{3}\right)^{-1}\right)$ |

5.1.0.2 Risk probabilities are assumed to be additive for all COPCs across all exposure pathways to estimate a total excess cancer risk. After summing all of the risks, the total excess cancer risk estimates are then compared to the point of departure of $1 \times 10^{-6}$ (USEPA, 1990). In general, total risks greater than $1 \times 10^{-4}$ require action; risks between $1 \times 10^{-6}$ and $1 \times 10^{-4}$ are in the risk management range and require the stakeholders to discuss and decide whether the risk estimates are acceptable; and risks less than $1 \times 10^{-6}$ are generally considered acceptable.

### 5.2 NONCARCINOGENIC EFFECTS

5.2.0.1 For exposure to noncarcinogens, adverse effects are not assumed to occur below a certain threshold (i.e., the RfD or RfC). The potential for adverse noncarcinogenic effects (i.e., the hazard quotient or HQ) was estimated by dividing the exposure level calculated for each exposure route by the corresponding noncancer toxicity value (i.e., RfD or RfC) (USEPA 1989, 1996, 2004a, 2009) as follows:

$$
\begin{aligned}
& \mathrm{HQ}_{\text {oral }}=\frac{\text { Intake }_{\text {oral }}}{\mathrm{RfD}_{\mathrm{o}}} \\
& \mathrm{HQ}_{\text {dermal }}=\frac{\text { Intake }_{\text {dermal }}}{\mathrm{RfD}_{\mathrm{d}}} \\
& \mathrm{HQ}_{\text {inhalation }}=\frac{\mathrm{C}_{\text {air }} \times \mathrm{EF} \times \mathrm{ED} \times \mathrm{ET} \times \mathrm{CF}}{\mathrm{RfC} \times \mathrm{AT} \times 365 \text { days } / \text { year }}
\end{aligned}
$$

where:

| AT | $=$ Averaging time (years) |
| :---: | :---: |
| $\mathrm{Cair}_{\text {air }}$ | $=$ COPC concentration in airborne dust or outdoor air ( $\mathrm{mg} / \mathrm{m}^{3}$ ) |
| ED | = Exposure duration (years) |
| EF | = Exposure frequency (days/year) |
| ET | $=$ Exposure time; i.e., the fraction of the day spent at the site (unitless) |
| CF | $=$ Conversion factor ( $1,000 \mu \mathrm{~g} / \mathrm{mg}$ ) |
| HQ | = Hazard quotient for each COPC (unitless) |
| Intake $_{\text {oral, dermal }}$ | = Oral and dermal exposure for each COPC ( $\mathrm{mg} / \mathrm{kg}$-day) |
| OAF | = Oral absorption factor (unitless) |
| $\mathrm{RfD}_{\text {o }}$ | $=$ Chemical specific oral reference dose ( $\mathrm{mg} / \mathrm{kg}$-day $)$ |
| $\mathrm{RfD}_{\text {d }}$ | $=\mathrm{RfD}_{0} \times \mathrm{OAF}$ |
| RfC | $=$ inhalation reference concentration $\left(\mu \mathrm{g} / \mathrm{m}^{3}\right)$ |

5.2.0.2 After summing all of the HQs for all COPCs across all exposure pathways, the sum is then compared to the USEPA acceptable hazard level of 1 . An HQ or HI less than 1 indicates a very low threat of adverse health effects, whereas an HQ or HI in excess of 1 indicates the potential for noncancer effects (USEPA, 1989a). It is important to consider that a HQ or HI above unity only indicates a potential for noncarcinogenic adverse health effects for the receptor. It does not predict the incidence, or severity, of effects (USEPA, 1989a).

### 5.3 RISK CHARACTERIZATION RESULTS

Tables 5-1 through 5-5 summarize the human health risk/hazard results for assumed exposures to soils at both $0-2$ and $0-10 \mathrm{ft} \mathrm{bgs}$. Appendix F provides the supporting calculations for the results presented in these tables.

### 5.3.1 Adult Residents

5.3.1.1 Total excess cancer risks for assumed adult residential exposures to soil (through incidental ingestion of soil, ingestion of homegrown vegetables, dermal contact with soil, and the inhalation of outdoor dusts) were estimated using the EPCs shown in Table 2-4. This results in total risk estimates of $6 \times 10^{-10}$ to $3 \times 10^{-9}$, depending on depth interval and whether RME or CT exposures are assumed (Table 5-1). These risk estimates are well below the point of departure of $1 \times 10^{-6}$ and the USEPA (1990) target risk range of $1 \times 10^{-6}$ to $1 \times 10^{-4}$.
5.3.1.2 Assumed adult residential exposures to these COPCs resulted in total HIs of approximately 0.1 to 0.3 , depending on depth interval and whether RME or CT exposures are assumed (Table 5-1). These estimates are all below 1, the benchmark level of concern for noncarcinogenic effects. This indicates that assumed exposures to COPCs at the site are unlikely to result in adverse noncarcinogenic health effects.

### 5.3.2 Child Residents

5.3.2.1 Total excess cancer risks for assumed child residential exposures to soil (through incidental ingestion of soil, ingestion of homegrown vegetables, dermal contact with soil, and the inhalation of outdoor dusts) were estimated using the EPCs shown in Table 2-4. This results in total risk estimates of $5 \times 10^{-10}$ to $7 \times 10^{-10}$, depending on depth interval and whether RME or CT exposures are assumed (Table 5-2). These risk estimates are well below the point of departure of $1 \times 10^{-6}$ and the USEPA (1990) target risk range of $1 \times 10^{-6}$ to $1 \times 10^{-4}$.
5.3.2.2 Assumed child residential exposures to these COPCs resulted in total HIs of approximately 1 (regardless of depth interval) for both RME and CT (regardless of depth interval) (Table 5-2). Since the HIs do not exceed 1, the benchmark level of concern for noncarcinogenic effects, assumed exposures to COPCs at the site are unlikely to result in adverse noncarcinogenic health effects. USEPA (1989a) indicates that the effects of noncarcinogens are not necessarily additive for different target endpoints. Therefore, the combined HQ by target organ will be even less. The target endpoints for the COPCs are summarized in Table 5-3.

### 5.3.3 Outdoor Workers

5.3.3.1 Total excess cancer risks for assumed outdoor worker exposures to soil (through incidental ingestion of soil, dermal contact with soil, and the inhalation of outdoor dusts) were estimated using the EPCs shown in Table 2-4. This results in total risk estimates of $2 \times 10^{-9}$ to 3 $\mathrm{x} 10^{-8}$, depending on depth interval and whether RME or CT exposures are assumed (Table 5-4). These risk estimates are well below the point of departure of $1 \times 10^{-6}$ and the USEPA (1990) target risk range of $1 \times 10^{-6}$ to $1 \times 10^{-4}$.

Table 5-1
Adult Resident Risk Estimates
4835 Glenbrook Road
Spring Valley, Washington, D.C.

| COPC | RME Risk Probabilities |  |  |  |  |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | Surface Soils (0-2 ft bgs) |  |  |  | Summation | Percent Contribution | Mixed Soils (0-10 ft bgs) |  |  |  | Summation | Percent Contribution |
|  | Ingestion | Dermal Contact | Inhalation of VOC/Dust in Outdoor Air | Ingestion of Home-Grown Vegetables |  |  | Ingestion | Dermal Contact | Inhalation of VOC/Dust in Outdoor Air | Ingestion of Home-Grown Vegetables |  |  |
| Aluminum | - | - | - | - | - | - | - | - | - | - | - | - |
| Cobalt | - | - | 3.00E-09 | - | 3.00E-09 | 95\% | - | - | 3.00E-09 | - | 3.00E-09 | 95\% |
| Copper | - | - | - | - | - | - | - | - | - | - | - | - |
| Manganese | - | - | - | - | - | - | - | - | - | - | - | - |
| Mercury | - | - | - | - | - | - | - | - | - | - | - | - |
| Nickel | - | - | $1.52 \mathrm{E}-10$ | - | $1.52 \mathrm{E}-10$ | 5\% | - | - | 1.49E-10 | - | 1.49E-10 | 5\% |
| Thallium | - | - | - | - | - | - | - | - | - | - | - | - |
| Vanadium | - | - | - | - | - | - | - | - | - | - | - | - |
| Summation | - | - | 3E-09 | - | 3E-09 |  | - | - | 3E-09 | - | 3E-09 |  |


| COPC | RME Hazard Index (HI) |  |  |  |  |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | Surface Soils (0-2 ft bgs) |  |  |  | Summation | Percent Contribution | Mixed Soils (0-10 ft bgs) |  |  |  | Summation | Percent Contribution |
|  | Ingestion | Dermal Contact | Inhalation of voc/Dust in Outdoor Air | Ingestion of Home-Grown Vegetables |  |  | Ingestion | Dermal Contact | Inhalation of voc/Dust in Outdoor Air | Ingestion of Home-Grown Vegetables |  |  |
| Aluminum | 3.17E-02 | - | $8.57 \mathrm{E}-05$ | - | 3.18E-02 | 11\% | 3.50E-02 | - | 9.47E-05 | - | $3.51 \mathrm{E}-02$ | 12\% |
| Cobalt | 1.92E-01 | - | $1.30 \mathrm{E}-04$ | - | $1.92 \mathrm{E}-01$ | 67\% | $1.92 \mathrm{E}-01$ | - | $1.30 \mathrm{E}-04$ | - | $1.92 \mathrm{E}-01$ | 63\% |
| Copper | 2.71E-03 | - | - | 8.76E-03 | 1.15E-02 | 4\% | $3.68 \mathrm{E}-03$ | - | - | $9.88 \mathrm{E}-03$ | $1.36 \mathrm{E}-02$ | 4\% |
| Manganese | 5.91E-03 | - | $2.24 \mathrm{E}-04$ | - | $6.13 \mathrm{E}-03$ | 2\% | $7.56 \mathrm{E}-03$ | - | $2.87 \mathrm{E}-04$ | - | $7.85 \mathrm{E}-03$ | 3\% |
| Mercury | $6.66 \mathrm{E}-04$ | - | $1.35 \mathrm{E}-08$ | - | $6.66 \mathrm{E}-04$ | 0\% | 5.49E-04 | - | $1.11 \mathrm{E}-08$ | - | $5.49 \mathrm{E}-04$ | 0\% |
| Nickel | 5.05E-03 | - | $1.52 \mathrm{E}-05$ | 4.33E-03 | $9.40 \mathrm{E}-03$ | 3\% | $4.93 \mathrm{E}-03$ | - | $1.48 \mathrm{E}-05$ | 4.25E-03 | $9.20 \mathrm{E}-03$ | 3\% |
| Thallium | 1.86E-02 | - | - | - | $1.86 \mathrm{E}-02$ | 6\% | $2.31 \mathrm{E}-02$ | - | - | - | $2.31 \mathrm{E}-02$ | 8\% |
| Vanadium | 1.84E-02 | - | - | - | $1.84 \mathrm{E}-02$ | 6\% | 2.12E-02 | - | - | - | 2.12E-02 | 7\% |
| Summation | 3E-01 | - | 5E-04 | 1E-02 | 3E-01 |  | 3E-01 | - | 5E-04 | 1E-02 | 3E-01 |  |


| COPC | CT Risk Probabilities |  |  |  |  |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | Surface Soils (0-2 ft bgs) |  |  |  | Summation | Percent Contribution | Mixed Soils (0-10 ft bgs) |  |  |  | Summation | Percent Contribution |
|  | Ingestion | Dermal Contact | Inhalation of vOC/Dust in Outdoor Air | Ingestion of Home-Grown Vegetables |  |  | Ingestion | Dermal Contact | Inhalation of VOC/Dust in Outdoor Air | Ingestion of Home-Grown Vegetables |  |  |
| Aluminum | - | - | - | - | - | - | - | - | - | - | - | - |
| Cobalt | - | - | $9.01 \mathrm{E}-10$ | - | $9.01 \mathrm{E}-10$ | 96\% | - | - | $6.01 \mathrm{E}-10$ | - | $6.01 \mathrm{E}-10$ | 94\% |
| Copper | - | - | - | - | - | - | - | - | - | - | - | - |
| Manganese | - | - | - | - | - | - | - | - | - | - | - | - |
| Mercury | - | - | - | - | - | - | - | - | - | - | - | - |
| Nickel | - | - | 3.95E-11 | - | 3.95E-11 | 4\% | - | - | 4.09E-11 | - | 4.09E-11 | 6\% |
| Thallium | - | - | - | - | - | - | - | - | - | - | - | - |
| Vanadium | - | - | - | - | - | - | - | - | - | - | - | - |
| Summation | - | - | 9E-10 | - | 9E-10 |  | - | - | 6E-10 | - | 6E-10 |  |


| COPC | CT Hazard Index (HI) |  |  |  |  |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | Surface Soils (0-2 ft bgs) |  |  |  | Summation | Percent Contribution | Mixed Soils (0-10 ft bgs) |  |  |  | Summation | Percent Contribution |
|  | Ingestion | Dermal Contact | Inhalation of vOC/Dust in Outdoor Air | Ingestion of Home-Grown Vegetables |  |  | Ingestion | Dermal Contact | Inhalation of vOC/Dust in Outdoor Air | Ingestion of Home-Grown Vegetables |  |  |
| Aluminum | 1.44E-02 | - | 7.79E-05 | - | 1.45E-02 | 11\% | 1.65E-02 | - | 8.91E-05 | - | 1.65E-02 | 15\% |
| Cobalt | 9.59E-02 | - | $1.30 \mathrm{E}-04$ | - | 9.60E-02 | 71\% | $6.39 \mathrm{E}-02$ | - | 8.65E-05 | - | $6.40 \mathrm{E}-02$ | 59\% |
| Copper | $1.20 \mathrm{E}-03$ | - | - | $1.39 \mathrm{E}-03$ | $2.59 \mathrm{E}-03$ | 2\% | $1.07 \mathrm{E}-03$ | - | - | $1.33 \mathrm{E}-03$ | $2.40 \mathrm{E}-03$ | 2\% |
| Manganese | $2.66 \mathrm{E}-03$ | - | $2.01 \mathrm{E}-04$ | - | $2.86 \mathrm{E}-03$ | 2\% | 3.27E-03 | - | $2.48 \mathrm{E}-04$ | - | $3.52 \mathrm{E}-03$ | 3\% |
| Mercury | $2.45 \mathrm{E}-04$ | - | $9.94 \mathrm{E}-09$ | - | $2.45 \mathrm{E}-04$ | 0\% | $2.28 \mathrm{E}-04$ | - | 9.24E-09 | - | $2.28 \mathrm{E}-04$ | 0\% |
| Nickel | 2.18E-03 | - | $1.31 \mathrm{E}-05$ | 6.46E-04 | $2.84 \mathrm{E}-03$ | 2\% | $2.26 \mathrm{E}-03$ | - | $1.36 \mathrm{E}-05$ | $6.63 \mathrm{E}-04$ | $2.94 \mathrm{E}-03$ | 3\% |
| Thallium | 8.26E-03 | - | - | - | 8.26E-03 | 6\% | $1.00 \mathrm{E}-02$ | - | - | - | $1.00 \mathrm{E}-02$ | 9\% |
| Vanadium | $8.20 \mathrm{E}-03$ | - | - | - | 8.20E-03 | 6\% | 9.17E-03 | - | - | - | $9.17 \mathrm{E}-03$ | 8\% |
| Summation | 1E-01 | - | 4E-04 | 2E-03 | 1E-01 |  | 1E-01 | - | 4E-04 | 2E-03 | 1E-01 |  |

Child Resident Risk Estimates
4835 Glenbrook Road
Spring Valley, Washington, D.C.

| COPC | RME Risk Probabilities |  |  |  |  |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | Surface Soils (0-2 ft bgs) |  |  |  | Summation | Percent Contribution | Mixed Soils (0-10 ft bgs) |  |  |  | Summation | Percent Contribution |
|  | Ingestion | Dermal Contact | Inhalation of VOC/Dust in Outdoor Air | Ingestion of Home-Grown Vegetables |  |  | Ingestion | Dermal Contact | Inhalation of VOC/Dust in Outdoor Air | Ingestion of Home-Grown Vegetables |  |  |
| Aluminum | - | - | - | - | - | - | - | - | - | - | - | - |
| Cobalt | - | - | 7.11E-10 | - | 7.11E-10 | 95\% | - | - | 7.11E-10 | - | 7.11E-10 | 95\% |
| Copper | - | - | - | - | - | - | - | - | - | - | - | - |
| Manganese | - | - | - | - | - | - | - | - | - | - | - | - |
| Mercury | - | - | - | - | - | - | - | - | - | - | - | - |
| Nickel | - | - | $3.61 \mathrm{E}-11$ | - | $3.61 \mathrm{E}-11$ | 5\% | - | - | 3.52E-11 | - | 3.52E-11 | 5\% |
| Thallium | - | - | - | - | - | - | - | - | - | - | - | - |
| Vanadium | - | - | - | - | - | - | - | - | - | - | - | - |
| Summation | - | - | 7E-10 | - | 7E-10 |  | - | - | 7E-10 | - | 7E-10 |  |


| COPC | RME Hazard Index (HI) |  |  |  |  |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | Surface Soils (0-2 ft bgs) |  |  |  | Summation | Percent Contribution | Mixed Soils (0-10 ft bgs) |  |  |  | Summation | Percent Contribution |
|  | Ingestion | Dermal Contact | Inhalation of voc/Dust in Outdoor Air | Ingestion of Home-Grown Vegetables |  |  | Ingestion | Dermal Contact | Inhalation of voc/Dust in Outdoor Air | Ingestion of Home-Grown Vegetables |  |  |
| Aluminum | $1.48 \mathrm{E}-01$ | - | $1.01 \mathrm{E}-04$ | - | $1.48 \mathrm{E}-01$ | 11\% | $1.63 \mathrm{E}-01$ | - | 1.12E-04 | - | $1.63 \mathrm{E}-01$ | 12\% |
| Cobalt | 8.95E-01 | - | $1.54 \mathrm{E}-04$ | - | 8.95E-01 | 69\% | 8.95E-01 | - | $1.54 \mathrm{E}-04$ | - | 8.95E-01 | 66\% |
| Copper | 1.27E-02 | - | - | $8.76 \mathrm{E}-03$ | $2.14 \mathrm{E}-02$ | 2\% | 1.72E-02 | - | - | $9.88 \mathrm{E}-03$ | $2.71 \mathrm{E}-02$ | 2\% |
| Manganese | 2.76E-02 | - | $2.65 \mathrm{E}-04$ | - | $2.78 \mathrm{E}-02$ | 2\% | 3.53E-02 | - | $3.39 \mathrm{E}-04$ | - | $3.56 \mathrm{E}-02$ | 3\% |
| Mercury | 3.11E-03 | - | $1.60 \mathrm{E}-08$ | - | 3.11E-03 | 0\% | 2.56E-03 | - | $1.32 \mathrm{E}-08$ | - | $2.56 \mathrm{E}-03$ | 0\% |
| Nickel | $2.36 \mathrm{E}-02$ | - | $1.80 \mathrm{E}-05$ | $4.33 \mathrm{E}-03$ | $2.79 \mathrm{E}-02$ | 2\% | $2.30 \mathrm{E}-02$ | - | $1.75 \mathrm{E}-05$ | 4.25E-03 | 2.73E-02 | 2\% |
| Thallium | 8.70E-02 | - | - | - | 8.70E-02 | 7\% | $1.08 \mathrm{E}-01$ | - | - | - | $1.08 \mathrm{E}-01$ | 8\% |
| Vanadium | 8.61E-02 | - | - | - | $8.61 \mathrm{E}-02$ | 7\% | 9.91E-02 | - | - | - | 9.91E-02 | 7\% |
| Summation | 1E+00 | - | 5E-04 | 1E-02 | $1 \mathrm{E}+00$ |  | 1E+00 | - | 6E-04 | 1E-02 | $1 \mathrm{E}+00$ |  |


| COPC | CT Risk Probabilities |  |  |  |  |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | Surface Soils (0-2 ft bgs) |  |  |  | Summation | Percent Contribution | Mixed Soils (0-10 ft bgs) |  |  |  | Summation | Percent Contribution |
|  | Ingestion | Dermal Contact | Inhalation of VOC/Dust in Outdoor Air | Ingestion of Home-Grown Vegetables |  |  | Ingestion | Dermal Contact | Inhalation of VOC/Dust in Outdoor Air | Ingestion of Home-Grown Vegetables |  |  |
| Aluminum | - | - | - | - | - | - | - | - | - | - | - | - |
| Cobalt | - | - | 7.11E-10 | - | 7.11E-10 | 96\% | - | - | 4.74E-10 | - | 4.74E-10 | 94\% |
| Copper | - | - | - | - | - | - | - | - | - | - | - | - |
| Manganese | - | - | - | - | - | - | - | - | - | - | - | - |
| Mercury | - | - | - | - | - | - | - | - | - | - | - | - |
| Nickel | - | - | $3.12 \mathrm{E}-11$ | - | $3.12 \mathrm{E}-11$ | 4\% | - | - | 3.23E-11 | - | $3.23 \mathrm{E}-11$ | 6\% |
| Thallium | - | - | - | - | - | - | - | - | - | - | - | - |
| Vanadium | - | - | - | - | - | - | - | - | - | - | - | - |
| Summation | - | - | 7E-10 | - | 7E-10 |  | - | - | $5 \mathrm{E}-10$ | - | 5E-10 |  |


| COPC | CT Hazard Index (HI) |  |  |  |  |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | Surface Soils (0-2 ft bgs) |  |  |  | Summation | Percent Contribution | Mixed Soils (0-10 ft bgs) |  |  |  | Summation | Percent Contribution |
|  | Ingestion | Dermal Contact | Inhalation of voc/Dust in Outdoor Air | Ingestion of Home-Grown Vegetables |  |  | Ingestion | Dermal Contact | Inhalation of vOC/Dust in Outdoor Air | Ingestion of Home-Grown Vegetables |  |  |
| Aluminum | $1.34 \mathrm{E}-01$ | - | 9.22E-05 | - | $1.34 \mathrm{E}-01$ | 11\% | 1.54E-01 | - | 1.05E-04 | - | $1.54 \mathrm{E}-01$ | 15\% |
| Cobalt | 8.95E-01 | - | $1.54 \mathrm{E}-04$ | - | 8.95E-01 | 72\% | 5.97E-01 | - | 1.02E-04 | - | 5.97E-01 | 60\% |
| Copper | 1.12E-02 | - | - | $1.39 \mathrm{E}-03$ | $1.26 \mathrm{E}-02$ | 1\% | $9.99 \mathrm{E}-03$ | - | - | $1.33 \mathrm{E}-03$ | 1.13E-02 | 1\% |
| Manganese | $2.48 \mathrm{E}-02$ | - | $2.38 \mathrm{E}-04$ | - | $2.50 \mathrm{E}-02$ | 2\% | 3.06E-02 | - | $2.94 \mathrm{E}-04$ | - | 3.09E-02 | 3\% |
| Mercury | 2.28E-03 | - | $1.18 \mathrm{E}-08$ | - | $2.28 \mathrm{E}-03$ | 0\% | 2.12E-03 | - | $1.09 \mathrm{E}-08$ | - | 2.12E-03 | 0\% |
| Nickel | 2.04E-02 | - | $1.55 \mathrm{E}-05$ | $6.46 \mathrm{E}-04$ | $2.10 \mathrm{E}-02$ | 2\% | $2.11 \mathrm{E}-02$ | - | $1.61 \mathrm{E}-05$ | $6.63 \mathrm{E}-04$ | $2.18 \mathrm{E}-02$ | 2\% |
| Thallium | 7.71E-02 | - | - | - | 7.71E-02 | 6\% | $9.36 \mathrm{E}-02$ | - | - | - | $9.36 \mathrm{E}-02$ | 9\% |
| Vanadium | 7.65E-02 | - | - | - | 7.65E-02 | 6\% | $8.56 \mathrm{E}-02$ | - | - | - | 8.56E-02 | 9\% |
| Summation | $1 \mathrm{E}+00$ | - | 5E-04 | $2 \mathrm{E}-03$ | $1 \mathrm{E}+00$ |  | $1 \mathrm{E}+00$ | - | 5E-04 | $2 \mathrm{E}-03$ | $1 \mathrm{E}+00$ |  |

Table 5-3
COPC Toxic Endpoints
4835 Glenbrook Road
Spring Valley, Washington, D.C.

| COPC | No Adverse Effects | Autoimmune effects | Respiratory Effects | Developmental Effects | Thyroid Effects | Central Nervous System (CNS) | Decreased <br> Body and Organ Weight | Gastrointestinal Effects | Hematopoietic Effects |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Ingestion |  |  |  |  |  |  |  |  |  |
| Aluminum |  |  |  | X |  | X |  |  |  |
| Cobalt |  |  |  | X | X |  |  |  | X |
| Copper |  |  |  |  |  |  |  | X |  |
| Manganese |  |  |  |  |  | X |  |  |  |
| Mercury |  | X |  |  |  |  |  |  |  |
| Nickel |  |  |  |  |  |  | X |  |  |
| Thallium | X |  |  |  |  |  |  |  |  |
| Vanadium | X |  |  |  |  |  |  |  |  |
| Inhalation |  |  |  |  |  |  |  |  |  |
| Aluminum |  |  |  |  |  | X |  |  |  |
| Cobalt |  |  | X |  |  |  |  |  |  |
| Copper |  |  |  |  |  |  |  |  |  |
| Manganese |  |  |  |  |  | X |  |  |  |
| Mercury |  |  |  |  |  | X |  |  |  |
| Nickel |  |  | X |  |  |  |  |  |  |
| Thallium |  |  |  |  |  |  |  |  |  |
| Vanadium |  |  |  |  |  |  |  |  |  |

Table 5-4
Outdoor Worker Risk Estimates
Surface Soils (0-2 ft bgs) and Mixed Soils (0-10 ft bgs)
4835 Glenbrook Road
Spring Valley, Washington, D.C.

| COPC | RME Risk Probabilities |  |  |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | Surface Soils (0-2 ft bgs) |  |  | Summation | Percent Contribution | Mixed Soils (0-10 ft bgs) |  |  | Summation | Percent Contribution |
|  | Ingestion | Dermal Contact | Inhalation of VOC/Dust in Outdoor Air |  |  | Ingestion | Dermal Contact | Inhalation of VOC/Dust in Outdoor Air |  |  |
| Aluminum | - | - | - | - | - | - | - | - | - | - |
| Cobalt | - | - | $2.86 \mathrm{E}-08$ | $2.86 \mathrm{E}-08$ | 95\% | - | - | $2.86 \mathrm{E}-08$ | $2.86 \mathrm{E}-08$ | 95\% |
| Copper | - | - | - | - | - | - | - | - | - | - |
| Manganese | - | - | - | - | - | - | - | - | - | - |
| Mercury | - | - | - | - | - | - | - | - | - | - |
| Nickel | - | - | 1.45E-09 | 1.45E-09 | 5\% | - | - | 1.42E-09 | 1.42E-09 | 5\% |
| Thallium | - | - | - | - | - | - | - | - | - | - |
| Vanadium | - | - | - | - | - | - | - | - | - | - |
| Summation | - | - | 3E-08 | 3E-08 |  | - | - | 3E-08 | 3E-08 |  |


| COPC | RME Hazard Index (HI) |  |  |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | Surface Soils (0-2 ft bgs) |  |  | Summation | Percent Contribution | Mixed Soils (0-10 ft bgs) |  |  | Summation | Percent Contribution |
|  | Ingestion | Dermal Contact | Inhalation of VOC/Dust in Outdoor Air |  |  | Ingestion | Dermal Contact | Inhalation of VOC/Dust in Outdoor Air |  |  |
| Aluminum | 1.09E-01 | - | 9.79E-04 | $1.10 \mathrm{E}-01$ | 12\% | 1.20E-01 | - | 1.08E-03 | $1.21 \mathrm{E}-01$ | 12\% |
| Cobalt | $6.58 \mathrm{E}-01$ | - | $1.48 \mathrm{E}-03$ | 6.59E-01 | 70\% | 6.58E-01 | - | $1.48 \mathrm{E}-03$ | 6.59E-01 | 66\% |
| Copper | $9.30 \mathrm{E}-03$ | - | - | $9.30 \mathrm{E}-03$ | 1\% | 1.26E-02 | - | - | $1.26 \mathrm{E}-02$ | 1\% |
| Manganese | 2.03E-02 | - | $2.56 \mathrm{E}-03$ | $2.28 \mathrm{E}-02$ | 2\% | 2.59E-02 | - | 3.28E-03 | 2.92E-02 | 3\% |
| Mercury | $2.28 \mathrm{E}-03$ | - | $1.55 \mathrm{E}-07$ | $2.28 \mathrm{E}-03$ | 0\% | 1.88E-03 | - | $1.27 \mathrm{E}-07$ | $1.88 \mathrm{E}-03$ | 0\% |
| Nickel | 1.73E-02 | - | $1.74 \mathrm{E}-04$ | $1.75 \mathrm{E}-02$ | 2\% | 1.69E-02 | - | 1.69E-04 | $1.71 \mathrm{E}-02$ | 2\% |
| Thallium | 6.39E-02 | - | - | $6.39 \mathrm{E}-02$ | 7\% | $7.90 \mathrm{E}-02$ | - | - | $7.90 \mathrm{E}-02$ | 8\% |
| Vanadium | $6.32 \mathrm{E}-02$ | - | - | $6.32 \mathrm{E}-02$ | 7\% | $7.28 \mathrm{E}-02$ | - | - | $7.28 \mathrm{E}-02$ | 7\% |
| Summation | 9E-01 | - | 5E-03 | 9E-01 |  | 1E+00 | - | 6E-03 | 1E+00 |  |


| COPC | CT Risk Probabilities |  |  |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | Surface Soils (0-2 ft bgs) |  |  | Summation | Percent Contribution | Mixed Soils (0-10 ft bgs) |  |  | Summation | Percent Contribution |
|  | Ingestion | Dermal Contact | Inhalation of VOC/Dust in Outdoor Air |  |  | Ingestion | Dermal Contact | Inhalation of VOC/Dust in Outdoor Air |  |  |
| Aluminum | - | - | - | - | - | - | - | - | - | - |
| Cobalt | - | - | $3.00 \mathrm{E}-09$ | $3.00 \mathrm{E}-09$ | 96\% | - | - | $2.00 \mathrm{E}-09$ | $2.00 \mathrm{E}-09$ | 94\% |
| Copper | - | - | - | - | - | - | - | - | - | - |
| Manganese | - | - | - | - | - | - | - | - | - | - |
| Mercury | - | - | - | - | - | - | - | - | - | - |
| Nickel | - | - | $1.32 \mathrm{E}-10$ | $1.32 \mathrm{E}-10$ | 4\% | - | - | $1.36 \mathrm{E}-10$ | $1.36 \mathrm{E}-10$ | 6\% |
| Thallium | - | - | - | - | - | - | - | - | - | - |
| Vanadium | - | - | - | - | - | - | - | - | - | - |
| Summation | - | - | 3E-09 | 3E-09 |  | - | - | 2E-09 | 2E-09 |  |


| COPC | CT Hazard Index (HI) |  |  |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | Surface Soils (0-2 ft bgs) |  |  | Summation | Percent Contribution | Mixed Soils (0-10 ft bgs) |  |  | Summation | Percent Contribution |
|  | Ingestion | Dermal Contact | Inhalation of VOC/Dust in Outdoor Air |  |  | Ingestion | Dermal Contact | Inhalation of VOC/Dust in Outdoor Air |  |  |
| Aluminum | 1.80E-02 | - | $2.60 \mathrm{E}-04$ | 1.83E-02 | 11\% | $2.06 \mathrm{E}-02$ | - | 2.97E-04 | 2.09E-02 | 16\% |
| Cobalt | 1.20E-01 | - | $4.33 \mathrm{E}-04$ | $1.20 \mathrm{E}-01$ | 72\% | 8.00E-02 | - | $2.88 \mathrm{E}-04$ | 8.03E-02 | 60\% |
| Copper | 1.51E-03 | - | - | $1.51 \mathrm{E}-03$ | 1\% | 1.34E-03 | - | - | $1.34 \mathrm{E}-03$ | 1\% |
| Manganese | 3.32E-03 | - | $6.71 \mathrm{E}-04$ | 3.99E-03 | 2\% | 4.10E-03 | - | 8.27E-04 | $4.93 \mathrm{E}-03$ | 4\% |
| Mercury | 3.06E-04 | - | 3.31E-08 | 3.06E-04 | 0\% | 2.85E-04 | - | 3.08E-08 | $2.85 \mathrm{E}-04$ | 0\% |
| Nickel | 2.73E-03 | - | $4.38 \mathrm{E}-05$ | $2.77 \mathrm{E}-03$ | 2\% | 2.83E-03 | - | $4.54 \mathrm{E}-05$ | $2.88 \mathrm{E}-03$ | 2\% |
| Thallium | 1.03E-02 | - | - | $1.03 \mathrm{E}-02$ | 6\% | 1.25E-02 | - | - | $1.25 \mathrm{E}-02$ | 9\% |
| Vanadium | 1.03E-02 | - | - | $1.03 \mathrm{E}-02$ | 6\% | 1.15E-02 | - | - | 1.15E-02 | 9\% |
| Summation | 2E-01 | - | 1E-03 | 2E-01 |  | 1E-01 | - | 1E-03 | 1E-01 |  |

5.3.3.2 Assumed outdoor worker exposures to these COPCs resulted in total HIs of 0.1 to 1 , depending on depth interval and whether RME or CT exposures are assumed (Table 5-5). These estimates do not exceed 1, the benchmark level of concern for noncarcinogenic effects. This indicates that assumed exposures to COPCs at the site are unlikely to result in adverse noncarcinogenic health effects.

### 5.4 UNCERTAINTY ANALYSIS

All RAs involve the use of assumptions, judgments, and imperfect data to varying degrees resulting in uncertainties in the final estimates of risk. These uncertainties are generally associated with the multitude of conditions that characterize each step of the RA process (i.e., data evaluation and identification of COPCs, exposure assessment, toxicity assessment, and risk characterization). These conditions are characteristically conservative and tend to overestimate potential site-related risks. This discussion qualitatively describes the major uncertainties in the RA for 4835 Glenbrook Road.

### 5.4.1 Uncertainty in Data Collection and Evaluation

5.4.1.1 The analysis of uncertainties focuses on determining whether the available data are representative of contaminant concentrations and site conditions, and whether the sampling, analyses, and/or statistical treatment of the data result in an over- or under-estimation of potential risk.
5.4.1.2 Where a chemical was not detected, USEPA's (2009c) ProUCL uses the KaplanMeier method to account for the effect of the non-detects on the estimated UCLs and central tendencies. Nonetheless, since the chemical was not detected, there is still some uncertainty in the true UCL and central tendency.
5.4.1.3 Contaminated soil was removed under interim measures by Apex (1996). While it is assumed that imported clean fill was used to bring the excavation back to grade, this is not mentioned in the report by Apex (1996). Further, the fill used in the excavation was not sampled.
5.4.1.4 The samples used in this RA were collected over a period of approximately 16 years by differing contractors and for different projects with different objectives. Therefore, the same chemicals were not analyzed in all of the samples that were collected. Thus, the sample sizes differ widely among the various analytes (see Tables 2-1, 2-2). Correspondingly, the confidence in the statistical evaluations of the data collected for the site is lower for those chemicals with fewer samples.
5.4.1.5 For some of the older data (e.g., EMS, 1992; APEX, 1996, EPA 1999), detection limits for non-detects were not provided. Thus, there is some uncertainty as to whether the detection limits were adequate for all of the non-detects.
5.4.1.6 Although a Livens round was found at the site while digging a test pit, the shell was empty and did not contain any explosives or CWM. No other MD, CWM, or MEC was found in the numerous test pits or soil borings advanced at the site, only 3 samples were analyzed for explosives and 8 samples were analyzed for agent breakdown products. Thus, it is possible, although unlikely, that MD, CWM, or MEC may be present at the site but was not detected.
5.4.1.7 The analysis of tentatively identified compounds (TICs) at a site can help to determine whether other contaminants were released at the site that were not targeted for analysis. This helps to reduce the uncertainty that chemicals not included in the analytical methods used to analyze the samples collected at the site (e.g., EPA Methods 8260 and 8270) are not overlooked. Three samples were analyzed for TICs at the site; i.e., $4835 \mathrm{~GB}(-190,50) \mathrm{SW}$ -N(5)LW-5, 4835GB(-190,50)SW-N(5)LW-4.5 (removed/excavated), SW-4835GB-04 (assoc w/ TP-40). The following TICs were detected in these three samples:

- (+)-Cycloisosativene
- Cyclotetradecane
- 2-Ethyl hexanoic acid
- 1-Methyl-4-(1-methylethyl)benzene
- E-11,13-Tetradecadien-1-ol
- 1,2,3,4-tetrahydro-1,6-dimethyl-4-(1-methylethyl)-naphthalene
5.4.1.8 1-Methyl-4-(1-methylethyl)benzene is also known as p-isopropyltoluene (or cymene), which is in the analyte list for EPA Method 8260. 2-Ethyl hexanoic acid is an industrial chemical that is used to prepare metal derivatives that are soluble in nonpolar organic solvents and was only detected in a sample that has since been removed as part of the remediation efforts at the site (i.e., SW-4835GB(-190,50)SW-N(5)LW-4.5). The other chemicals are insect pheromones (i.e., E-11,13-tetradecadien-1-ol) and products naturally produced by plants (i.e., cycloisosativene, cyclotetradecane, and 1,2,3,4-tetrahydro-1,6-dimethyl-4-(1-methylethyl)naphthalene). However, none of these chemicals (including p-isopropyltoluene) have either toxicity values or screening levels. Since the site history was used to select the analytical methods used at the site, the samples collected at the site were analyzed for a wide range of chemicals, and the TICs that were detected are either naturally occurring or chemicals that are not known to be toxic, it is assumed that the uncertainty associated with analyzing only three samples collected at the site for TICs is small.
5.4.1.9 Steady-state conditions were assumed for evaluation of potential future exposures which may tend to overestimate long-term exposure and health risk since contaminant concentrations may decrease over time.


### 5.4.2 Uncertainty in Exposure Assessment

5.4.2.1 There is some concern as to how well an exposure scenario approximates the precise conditions to which a receptor may be exposed at a given site. Potential human exposures could deviate from those estimated in this RA through differences in exposure frequency, contact rate, exposure duration, body weight, and life span. However, given that the RME exposure parameter values generally consist of upper bound (i.e., $95^{\text {th }}$ percentile) estimates, it is likely that the RME exposure estimate presented here is an upper-bound estimate that would overestimate exposures (and risks) for the average receptor.
5.4.2.2 Only Parsons' data were used to calculate the exposure point concentrations for COPCs to be consistent with the statistical analysis performed in Appendix B. This will result in an overestimation of the actual risk since the EPCs of COPCs calculated using all data will be lower. For cobalt, a $95 \%$ UCL was not calculated since there are only three data points collected

[^9]by Parsons. If USEPA 1999 data are included, the cobalt EPC for $0-10$ feet mixed soil will be $31.5 \mathrm{mg} / \mathrm{kg}$, which is approximately $25 \%$ lower than the maximum concentration of $42 \mathrm{mg} / \mathrm{kg}$ used in the risk assessment.
5.4.2.3 Generic models were used to estimate the concentrations of COPCs in vegetables grown at the site. However, bioaccumulation from soils to plants is dependent upon multiple factors, including soil pH , metal species present in the soils, plant species, part of the plant measured/consumed, etc. Thus, the predicted concentrations in vegetables presented here are subject to some uncertainty.
5.4.2.4 The soil ingestion rates assumed in this risk assessment are incidental soil ingestion rates; i.e., ingestion of soil or dust particles that adhere to food, cigarettes, objects that are mouthed, or hands. However, some children are known to exhibit a behavior called soil-pica, which "...is the recurrent ingestion of unusually high amounts of soil" (USEPA 2008). Children that exhibit soil-pica behavior have much higher soil ingestion rates than assumed here; i.e., 1,000 to $5,000 \mathrm{mg} /$ day or more (USEPA, 2008) vs. $100 \mathrm{mg} /$ day. Thus, soil-pica children would be expected to have correspondingly higher (i.e., 10 to 50 X ) exposures, and risks, than were estimated here.
5.4.2.5 Due to the numerous test pits and sampling activities that have taken place at the site, some of the samples that were categorized as "surface" may no longer be near the soil surface and may be currently be in deeper soils. However, assumed exposures to both $0-2$ and $0-10 \mathrm{ft}$ bgs were evaluated at the site. Therefore, the uncertainty associated with the current depth of the samples is expected to be small.
5.4.2.6 The incidental soil ingestion rate for outdoor workers assumed here is $480 \mathrm{mg} /$ day (USEPA 1997a). Current USEPA (2002) guidance recommends an incidental soil ingestion rate of $330 \mathrm{mg} /$ day for outdoor workers. Thus, exposures via soil ingestion may have been overestimated for outdoor workers.
5.4.2.7 Outdoor workers were assumed to be exposed to the same level of dust as were residents. It is expected that the outdoor workers at the site would generate and, thus, be exposed to, higher levels of dusts than residents as the outdoor workers at the site are expected to engage in activities that generate dusts, including lawn mowing, leaf blowing, soil tilling, etc. Unfortunately, USEPA (1995c, 1996a, 2002), does not provide guidance on estimating dust emissions from these kinds of activities. Therefore, it is likely outdoor worker exposures to dusts at the site have been underestimated.
5.4.2.8 For residents, adult exposures were evaluated assuming that adults were present at the site for 30 years under an RME exposure scenario. Under USEPA (1989a) guidance, residents are assumed to be present for 30 years at a site, with the first 6 years as a child and the remaining 24 as an adult. Thus, the residential exposures estimated here should be regarded as highly health protective.

### 5.4.3 Uncertainty in Toxicity Assessment

5.4.3.1 Some uncertainty is also inherent in the toxicity values used in the RA. Carcinogenic SFs are route-specific values derived only for compounds that have been shown to cause an increased incidence of tumors in either human or animal studies. Dose-response relationships between tumor incidence and exposure using human epidemiologic or animal studies are used to derive the SF. This dose-response curve is then assumed to be linear at low doses (e.g., those
found in situations of environmental contamination) and is used to predict tumor incidence at low exposure levels. When an animal study is used, the final SF is adjusted to account for extrapolation of animal data to humans. If the studies used to derive the SF were conducted for less than the life span of the test organism, the final SF has also been adjusted to reflect risk associated with lifetime exposure.
5.4.3.2 The SF is generally an upper $95^{\text {th }}$ percentile confidence limit of the probability of a response based on experimental animal data in the multistage model. This means that the sitespecific chemical risk is not likely to exceed the risk estimate derived through the model and is likely to be less than the predicted risk.
5.4.3.3 The chronic RfD for a compound is based on studies where either human or animal populations were exposed to a given compound by a given route of exposure for the major portion of the life span (as a USEPA guideline, seven years to a lifetime) (USEPA, 1989a). RfDs are derived by determining dose-specific effect levels from all available quantitative studies and applying uncertainty factors to the most appropriate effect level to determine a RfD for humans. Uncertainty factors are generally applied as multiples of 10 to represent specific areas of uncertainty in the data. Typically, an uncertainty factor of 100 to 1,000 is used in the adjustments. In addition, USEPA may use a modifying factor of up to 10 that applies to professional judgment of uncertainties. General uncertainties in the derivation of RfDs may be associated with factors such as: (1) variations in the general population (to protect sensitive receptors); (2) extrapolation of animal data to humans; (3) use of a subchronic study versus a chronic study to determine the no-observed-adverse-effect level (NOAEL); or (4) use of a lowest-observed-adverse-effect level (LOAEL) versus a NOAEL. Both the uncertainty and modifying factors are conservative in nature and tend to overestimate risk.
5.4.3.4 As indicated above, toxicity factors are generally route specific (i.e., they are either for inhalation or oral exposure to a given chemical). In this risk assessment, oral RfDs and CSFs were used to evaluate the risk associated with ingestion of a given chemical. RfCs and inhalation URFs were used to evaluate the risk associated with inhalation of chemicals. Due to differences in the exposure pathways, route-to-route extrapolation was not performed between oral and inhalation pathways. In other words, if an inhalation toxicity factor did not exist, the oral RfD or CSF was not used to calculate one. For analytes that are inhaled, are absorbed through the lungs, and have systemic toxic effects, the absence of route-to-route extrapolation will tend to underestimate the risk associated with inhalation exposure to a given chemical. Conversely, for chemicals that have only portal of entry effects, and not systemic effects, the use of route-to-route extrapolation would tend to overestimate the risks.
5.4.3.5 The following chemicals that were detected in soils at the site do not have toxicity values and could not be evaluated quantitatively in this RA:

- (+)-Cycloisosativene
- Iodine pentafluoride
- E-11,13-Tetradecadien-1-ol
- 1,2,3,4-Tetrahydro-1,6-dimethyl-4-(1-methylethyl)-naphthalene
5.4.3.6 However, all of these chemicals, except iodine pentafluoride, are naturally occurring in plants and/or animals. Thus, it cannot be determined, what, if any, contributions former

Department of Defense activities at this site would have had to concentrations of these chemicals.
5.4.3.7 Iodine pentafluoride (as iodate) was detected in both of the soil samples that were analyzed for this chemical. Although the lab reported the detection was iodine pentafluoride, it is more likely that an iodate salt was detected; e.g., sodium iodate $\left(\mathrm{NaIO}_{3}\right)$, silver iodate $\left(\mathrm{AgIO}_{3}\right)$, and calcium iodate $\left(\mathrm{Ca}\left(\mathrm{IO}_{3}\right)_{2}\right)$. In addition to the uncertain identity of the actual iodate present, there are no toxicity values available from the approved sources listed in USEPA (2003) guidance. Thus, the effects from assumed exposures to iodates can not be quantified.
5.4.3.8 Tellurium is a naturally occurring metal in the Earth's crust and it was detected in all three of the soil samples that were analyzed for this metal. However, the maximum detected concentration (i.e., $6.6 \mathrm{mg} / \mathrm{kg}$ ) exceeded the background UTL of $5 \mathrm{mg} / \mathrm{kg}$. At present, it is not possible to quantitatively evaluate exposures to tellurium in a risk assessment, as there are no toxicity values available from the approved sources listed in USEPA (2003) guidance. However, there are reports of adverse effects in humans from occupational exposures to tellurium, which would be expected to be much higher than at the site. The symptoms associated with occupational exposures to high levels of tellurium include garlic odor of the breath and sweat, dryness of the mouth, metallic taste, somnolence, anorexia, occasional nausea, patches of skin that are scaly, itchy, and have lost the ability to sweat function (HSDB, 2009). Thus, the effects from assumed exposures to tellurium can not be quantified.

### 5.4.4 Uncertainty in Estimating Chemical Risk

5.4.4.1 The expression of the potential risk associated with contaminants detected at the site is a result of the combined steps of data evaluation, exposure assessment, and toxicity assessment. This combination provides the potential to magnify the uncertainties present in these steps of the RA process.
5.4.4.2 The chemical risk calculations include the risk associated with exposure to all COPCs evaluated at the site. Whenever carcinogenic and non-carcinogenic toxicity factors are available for a given chemical, the risk and hazard are both calculated. Cumulative risk is calculated using all available analytes. However, the risks are not necessarily additive; e.g., the risks could be synergistic or even antagonistic. When the non-carcinogenic hazard quotient is greater than 1 , potential target organ effects were considered. Only those chemicals that affected the same target organ, as indicated by the critical study for calculating the RfD, were considered to have a cumulative toxicity. This assumption may tend to underestimate the hazard, should a chemical affect multiple target organs that are not represented in the critical study.
5.4.4.3 Arsenic in soils at the site did not exceed the Spring Valley remediation goal of 20 $\mathrm{mg} / \mathrm{kg}$ and was, therefore, not included as a COPC. However, at the request of American University and the DDOE, the risks from assumed exposures to arsenic were evaluated for informational purposes. The results are presented in Appendix H and are summarized here. The soil EPCs for arsenic are $10.55 \mathrm{mg} / \mathrm{kg}$ for $0-2 \mathrm{ft}$ bgs and $11.17 \mathrm{mg} / \mathrm{kg}$ for $0-10 \mathrm{ft} \mathrm{bgs}$. The risks and hazards from assumed exposures to arsenic in soils at the site (through incidental ingestion of soil, ingestion of homegrown vegetables, dermal contact with soil, and the inhalation of outdoor dusts) are calculated for adult residents, child residents and outdoor workers, and included in Appendix H. The calculated RME and CT risks and noncancer hazards from
assumed exposures to arsenic are presented in Table H.6. The arsenic RME risks and hazards are summarized below:

|  | Risks |  |  | Hazards |  |
| :--- | :---: | :---: | :--- | :--- | :--- | :--- |
| Depth (ft bgs) | $0-2$ | $0-10$ |  | $0-2$ | $0-10$ |
| Adult resident | $2 \times 10^{-5}$ | $2 \times 10^{-5}$ |  | 0.1 | 0.1 |
| Child resident | $1 \times 10^{-5}$ | $1 \times 10^{-5}$ |  | 0.3 | 0.3 |
| Outdoor worker | $3 \times 10^{-5}$ | $3 \times 10^{-5}$ |  | 0.2 | 0.2 |

The COPCs and arsenic combined risk and hazards for adult residents, child residents and outdoor workers are presented in Tables H.7.1, H.7.2, and H.7.3. The combined RME risk and hazards are summarized below:

|  | Risks |  |  | Hazards |  |
| :--- | :---: | :---: | :---: | :---: | :---: |
| Depth (ft bgs) | $0-2$ | $0-10$ |  | $0-2$ | $0-10$ |
| Adult resident | $2 \times 10^{-5}$ | $2 \times 10^{-5}$ |  | 0.4 | 0.4 |
| Child resident | $1 \times 10^{-5}$ | $1 \times 10^{-5}$ |  | 2 | 2 |
| Outdoor worker | $3 \times 10^{-5}$ | $3 \times 10^{-5}$ |  | 1 | 1 |

As shown in the above summary, the risk estimates including arsenic are within the USEPA (1990) target risk range of $1 \times 10^{-6}$ to $1 \times 10^{-4}$ and the noncancer hazards are below the threshold value of 1 except for HIs for child residents. HIs for child residents are 2, which are greater than 1. USEPA (1989a) indicates that the effects of noncarcinogens are not necessarily additive. Therefore, when the HI from assumed exposures to more than one COPC are greater than 1 , the HQs should be separated by toxic endpoint (or target organ); i.e., separate HIs are calculated for each toxic endpoint (USEPA 1989a). The toxic endpoints for the COPCs at the site are summarized in Table H.7.4. For the ingestion pathway, the arsenic target organs are skin and cardiovascular system. For the inhalation pathway, the arsenic target organs are respiratory, developmental, central nervous system, and cardiovascular system.
When COPCs are summed including arsenic by toxic endpoint (Table H.7.5), none of the HIs exceed 1. This indicates that assumed exposures to COPCs and arsenic at the site are unlikely to result in cumulative adverse noncarcinogenic health effects. Although the HI for developmental effects is 1 and the HIs for thyroid and hematopoietic effects are each 0.9 , none of these exceeded the benchmark level of concern of 1 . The other toxic endpoints were all well below 1 . Since adequate safety margins have been built into both the exposure assessment process and the toxicity criteria used to estimate potential risks, multiple HIs at or below 1 are unlikely to result in adverse noncarcinogenic health effects.

### 5.5 RISK SUMMARY

5.5.0.1 The carcinogenic risks estimated for the three receptor groups assumed to be exposed to COPCs in soils (via ingestion, dermal contact, and the inhalation of dusts, as well as homegrown vegetable ingestion for residents) at the site are summarized in Tables 5-1, 5-2, and 5-4. The carcinogenic risks estimated for adult residents, child residents, and outdoor workers are all well below the point of departure of $1 \times 10^{-6}$, regardless of depth interval (i.e., $0-2 \mathrm{vs} .0-10 \mathrm{ft}$ bgs) and exposure scenario (i.e., RME vs. CT).
5.5.0.2 Tables 5-1, 5-2, and 5-4 show that the noncarcinogenic HIs estimated for adult residents, child residents, and outdoor workers assumed to be exposed to COPCs in soils (via ingestion, dermal contact, and the inhalation of dusts, as well as homegrown vegetable ingestion for residents) at the site do not exceed USEPA's benchmark level of concern for noncarcinogenic effects of 1 .
5.5.0.3 The combined RME risk and hazards from arsenic and the identified COPCs show that the risk estimates including arsenic are within the USEPA (1990) target risk range of 1 x $10^{-6}$ to $1 \times 10^{-4}$ and the noncancer hazards do not exceed the threshold value of 1 (when summed by toxic endpoint for children). This indicates that assumed exposures to COPCs and arsenic at the site are unlikely to result in adverse noncarcinogenic health effects.

[^10]
## SECTION 6 <br> CONCLUSIONS

6.0.0.1 The primary objective of this RA was to quantitatively characterize the human health risk associated with current and reasonably expected future exposure to contaminated soils at 4835 Glenbrook Road. The potential receptors at the site include outdoor workers and future residents. The exposure pathways evaluated here include incidental soil ingestion, dermal contact with soils, and the inhalation of particulates for all receptors. In addition, the ingestion of homegrown vegetables was evaluated for residents (Figure 3-1). Tables 5-1 through 5-4 provide a summary of the human health risk for each COPC for each receptor.
6.0.0.2 The cumulative cancer risk estimates for child residents, adult residents, and outdoor workers are all below the USEPA point of departure of $1 \times 10^{-6}$. Thus, unacceptable cancer risks to the receptors resident are not expected from assumed exposures to COPCs in soils at the site. Additionally, the hazard indexes (HIs) estimated for all receptors at the site do not exceed the benchmark level of concern of 1 . This indicates that unacceptable adverse noncarcinogenic health effects are not expected from assumed exposures to COPCs in soils at the site. Overall, this indicates that the risks and hazards from assumed exposures to soils at 4835 Glenbrook Road are acceptable and that further action is not warranted.
6.0.0.3 The combined RME risk and hazards from arsenic and the identified COPCs show that the risk estimates including arsenic are within the USEPA (1990) target risk range of 1 x $10^{-6}$ to $1 \times 10^{-4}$ and the noncancer hazards do not exceed the threshold value of 1 (when summed by toxic endpoint for children). This indicates that assumed exposures to COPCs and arsenic at the site are unlikely to result in adverse noncarcinogenic health effects.

[^11]
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[^12]
## APPENDIX A

DATA SUMMARY

P:\ISEH $746040($ NewDA01) 105 _Suppl RA \& MEC Haz Assess 44835 RISK ASSESSMENT\Final\Revision 1\Final 4835 Glenbrook Rd HHRA9-11-09.doc Rev 2 Contract No. DACA87-02-D-0005
Delivery Order No. DA01

TABLE A.1--SUMMARY OF SAMPLING EVENTS AT 4835 GLENBROOK ROAD

| TYPE | SAMPLE NAME | No. of Samples | Arsenic | Combination-12-suite, Prior Pollutant, or TAL METALS | VOCs | SVOCs | $\begin{aligned} & \text { Pest/ } \\ & \text { PCBs } \end{aligned}$ | Explosives | HD or L Agent | $\begin{gathered} \mathrm{HD} \\ \text { ABPs } \end{gathered}$ | $\stackrel{\mathrm{L}}{\mathrm{ABPs}}$ | SOURCE |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 4835 SOIL SAMPLING |  |  |  |  |  |  |  |  |  |  |  |  |
| 1992 Construction worker incident sampling. No specific locations. Table A. 2 | $\begin{aligned} & 050992 \\ & 052692 \end{aligned}$ | 1 | 1 | 1 | 1 | 1 | 1 |  |  |  |  | EMS 1992 <br> Report |
| 1996 Initial tree removal area samples. No specific locations. Table A. 3 | $\begin{aligned} & 1002, \\ & 9000 \end{aligned}$ | 2 | 2 | 2 | 2 | 2 | 2 |  |  |  |  | $1996 \text { APEX }$ <br> Report |
| 1996 soil samples for tree removal. Pre and post removal, surf and sub. Table A. 3 | $\begin{aligned} & 9005 \text { thru } \\ & 9013 \end{aligned}$ | 9 | 9 | 9 | 9 |  |  |  |  |  |  | $\begin{array}{\|l} \hline 1996 \text { APEX } \\ \text { Report } \\ \hline \end{array}$ |
| 1996 sub-surf soil samples-Based on high PID readings in the 4 ft deep probes. Table A. 3 | $\begin{aligned} & 9014 \text { thru } \\ & 9019 \end{aligned}$ | 6 | 6 | 6 | 6 |  |  |  |  |  |  | $1996 \text { APEX }$ <br> Report |
| 1999 surface samples. G-3 has a subsurface component, which is SB-02 for metals only (assoc with XRF arsenic. Table A. 4 | $\begin{gathered} \text { G-01, G- } \\ \text { 02, and } \\ \text { G-03 } \end{gathered}$ | 4 | 4 | 4 | 3 | 3 | 3 |  |  |  |  | Interim Trip <br> Report \#1, <br> EPA Aug <br> 10, 1999 |
| Oct 2000 Quad sx, 4 surf, 3 sub-surf intervals. Oct 2000 arsenic grid sx. Tables A. 5 and A. 6 | MTL-1 or SB | 67 | 60 |  |  |  |  |  |  | 7 |  | Parsons |
| 2007-2008 samples Assoc w/arsenic grids. Table A. 7 | 4835 (x,y) | 113 | 113 | 113 |  |  |  |  |  |  |  | Parsons |
| 2007-2008 sub-surface grabs assoc w/test pits. 2 northernmost got full list or Metals+Explosives. <br> Tables A. 8 and A. 9 | $\begin{gathered} 4835 \\ (\mathrm{~GB}-01) \end{gathered}$ | 6 | 6 | 2 | 1 | 1 |  | 2 | 6 | 6 | 6 | Parsons |
| April 20082 high arsenic related 'full suite' samples. One Waste Profile sx. Table A. 10 | $(-190,50)$ | 3 |  | 3 | 3 | 3 |  | 3 |  | 2 |  | Parsons |
| TOTALS |  |  | 300+ | 130+ | 25 | 10 | 6 | 5 | 6 | 15 | 6 |  |

4835 OTHER SAMPLING and/or MONITORING
--Near real time air monitoring (MINICAMS) for HD, L, Arsine, Cyanogen Chloride, Phosgene, Chloropicrin--with no confirmed detections at 76 test pits (Parsons).
--PID readings in approx 90 soil probes in backyard, with no significant readings (data unavailable), where highest had soil sampled--see above (APEX 1996)
Note: Six additional soil samples were collected during the November and December 2008 arsenic removal activities. Two samples were excavated due to elevated arsenic concentrations. The analytical results of the remaining four samples are included in Table A.7.

Table A.2--EMS 1992
Glenbrook Road Soil Sampling
Spring Valley FUDS

| SAMPLE NUMBER--> | RSL | 052692-1CM |
| :---: | :---: | :---: |
| VOCs (ug/kg) - 8240 | Non-Carcs adjusted downward |  |
| Acetone |  | $<2$ |
| Benzene |  | <1 |
| Bromodichloromethane |  | <1 |
| Bromoform |  | <5 |
| Bromomethane |  | $<10$ |
| 2-Butanone |  | <5 |
| Carbon Disulfide |  | <2 |
| Carbon tetrachloride |  | $<1$ |
| Chlorobenzene |  | $<1$ |
| Chloroethane |  | <5 |
| Chloroform |  | <1 |
| Chloromethane |  | <10 |
| Dibromochloromethane |  | <5 |
| 1,2-Dichloroethane |  | <1 |
| 1,1-Dichloroethane |  | <1 |
| 1,1-Dichloroethene |  | $<1$ |
| 1,2-Dichloroethene |  | <1 |
| 1,2-Dichloropropane |  | $<1$ |
| Trans-1,3-Dichloropropene |  | $<1$ |
| Cis-1,3-Dichloropropene |  | $<1$ |
| Ethylbenzene |  | <1 |
| 2-Hexanone |  | <5 |
| 4-Methyl-2-Pentanone |  | <1 |
| Methylene chloride | 11,000 | 74 |
| Styrene |  | <1 |
| 1,1,2,2-Tetrachloroethane |  | $<1$ |
| Tetrachloroethene |  | $<1$ |
| Toluene | 500,000 | 2 |
| 1,1,1-Trichloroethane |  | $<1$ |
| 1,1,2-Trichloroethane |  | <1 |
| Trichloroethene |  | $<1$ |
| Vinyl Acetate |  | $<1$ |
| Vinyl chloride |  | <5 |
| Xylenes, total |  | <5 |
| SVOCs (ug/kg) -8270 |  |  |
| Acenaphthalene |  | $<100$ |
| Acenaphthalyene |  | <100 |
| Anthracene |  | <100 |
| Benz(a)anthracene |  | <100 |
| Benzo(b)fluoranthene |  | <100 |
| Benzo(k)fluoranthene |  | <100 |
| Benzoic acid |  | <100 |
| Benzo(g,h,i)perylene |  | <100 |
| Benzo(a)pyrene |  | <100 |
| Benzyl alcohol |  | <100 |
| Bis(2-chloroethoxy)methane |  | <100 |
| Bis(2-chloroethoxy)ether |  | <100 |
| Bix(2-chloroisopropyl)ether |  | $<100$ |
| 4-bromophenyl phenyl ether |  | <100 |
| Butyl benzyl phthalate |  | <100 |

Table A.2--EMS 1992 Glenbrook Road Soil Sampling Spring Valley FUDS

| SAMPLE NUMBER--> | RSL | 052692-1CM |
| :---: | :---: | :---: |
| 4-Chloroaniline |  | <100 |
| 4-Chloro-3-methylphenol |  | $<100$ |
| 2-Chloronaphthalene |  | <100 |
| 2-Chlorophenol |  | <100 |
| 4-Chlorophenyl phenyl ether |  | <100 |
| Chrysene |  | $<100$ |
| Dibenz(a,h)anthracene |  | $<100$ |
| Di-n-butyl-phthalate |  | <100 |
| Dibenzofuran |  | <100 |
| 1,2-Dichlorobenzene |  | <100 |
| 1,3-Dichlorobenzene |  | <100 |
| 1.4-Dichlorobenzene |  | $<100$ |
| 3,3-Dichlorobenzidine |  | <100 |
| 2,4-Dichlorophenol |  | $<100$ |
| Diethyl phthalate |  | $<100$ |
| 2,4-Dimethylphenol |  | <100 |
| Dimethyl phthalate |  | <100 |
| 4.6-Dinitro-2-methylphenol |  | $<100$ |
| 2,4-Dinitrophenol |  | <100 |
| 2,4-Dinitrotoluene |  | <100 |
| 2,6-Dinitrotoluene |  | <100 |
| Di-n-octyl phthalate |  | $<100$ |
| Bis(2-ethylhexyl)phthalate |  | $<100$ |
| Fluoranthene |  | <100 |
| Hexachlorobenzene |  | $<100$ |
| Hexachlorobutadine |  | $<100$ |
| Hexachlorocyclopentadiene |  | <100 |
| Hexachloroethane |  | <100 |
| Indeno(1,2,3-cd)pyrene |  | <100 |
| Isophorone |  | $<100$ |
| 2-Methylnaphthalene |  | $<100$ |
| 2-Methylphenol |  | <100 |
| 4-Methylphenol |  | <100 |
| Naphthalene |  | <100 |
| 2-Nitroaniline |  | <100 |
| 3-Nitroaniline |  | $<100$ |
| 4-Nitroaniline |  | $<100$ |
| Nitrobenzene |  | $<100$ |
| 2-Nitrophenol |  | <100 |
| 4-Nitrophenol |  | <100 |
| N-Nitrosodiphenylamine |  | <100 |
| N-Nitroso-di-n-propylamine |  | $<100$ |
| Pentachlorophenol |  | $<100$ |
| Phenantrene |  | $<100$ |
| Phenol |  | <100 |
| Pyrene |  | <100 |
| 1,2,4-Trichlorobenzene |  | <100 |
| 2,4,5-Trichlorophenol |  | <100 |
| 2,4,6-Trichlorophenol |  | <100 |

Table A.2--EMS 1992
Glenbrook Road Soil Sampling
Spring Valley FUDS

| SAMPLE NUMBER--> | RSL | 052692-1CM |
| :---: | :---: | :---: |
| Metals (mg/kg) - 200.7 |  |  |
| Arsenic |  | <10 |
| Barium | 1,500 | 145 |
| Cadmium |  | <0.5 |
| Chromium | 12,000 | 54 |
| Lead | 400 | 100 |
| Mercury |  | <0.1 |
| Selenium |  | <10 |
| Silver |  | <1 |
| PCB Type (mg/kg) |  |  |
| 1016 |  | <0.1 |
| 1221 |  | <0.1 |
| 1232 |  | <0.1 |
| 1242 |  | $<0.1$ |
| 1248 |  | <0.1 |
| 1254 |  | $<0.1$ |
| 1260 |  | <0.1 |
| Pesticide Type (ug/kg) |  |  |
| Aldrin |  | <100 |
| A-BHC |  | $<100$ |
| b-BHC |  | <100 |
| g-BHC (Lindane) |  | $<100$ |
| d-BHC |  | $<100$ |
| Chlordane (total) |  | <100 |
| 4, $4^{1}$-DDD |  | <100 |
| 4,4 ${ }^{1}$-DDE |  | <100 |
| 4,4 ${ }^{1}$-DDT |  | <100 |
| Dieldrin |  | <110 |
| Endosulfan I |  | <100 |
| Endosulfan II |  | <100 |
| Endosulfan sulfate |  | <100 |
| Endrin |  | <100 |
| Endrin Ketone |  | <100 |
| Heptachlor |  | <100 |
| Heptachlor epoxide |  | <100 |
| Herbicide Type (ug/kg) |  |  |
| 2,4-D |  | $<10$ |
| 2,4,5-TP (silvex) | 63,000 | 13 |
| < = Not detected at this reporting limit. |  |  |
| Detections are bolded. |  |  |
| RSLs are EPA Regional Screening Levels (12 September 2008). |  |  |
| RSLs are shown only for detected chemicals. |  |  |
| SOURCE: Environmental Management Systems, Inc., May and June 1992 |  |  |
| Letter Reports. |  |  |

Table A.3--APEX 1996
4835 Glenbrook Road Soil Sampling
Spring Valley FUDS

| SAMPLE NUMBER--> | RSL | 1002 | 9000 | 9005 | 9006 | 9007 | 9008 | 9009 | 9010 | 9011 | 9012 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| VOCs (ug/kg) - 8260 | Non-Carcs | adjusted downward |  |  |  |  |  |  |  |  |  |
| Benzene | 1,100 | 280 | $<5$ | $<5$ | $<5$ | $<5$ | $<5$ | $<5$ | $<5$ | $<5$ | $<5$ |
| Bromobenzene |  | $<25$ | $<5$ | $<5$ | <5 | $<5$ | <5 | <5 | $<5$ | $<5$ | $<5$ |
| Bromochloromethane |  | $<25$ | $<5$ | $<5$ | $<5$ | $<5$ | <5 | $<5$ | $<5$ | $<5$ | $<5$ |
| Bromodichloromethane |  | $<25$ | $<5$ | <5 | <5 | <5 | <5 | <5 | <5 | $<5$ | $<5$ |
| Bromoform |  | $<25$ | $<5$ | $<5$ | <5 | $<5$ | $<5$ | <5 | $<5$ | e1 | $<5$ |
| Bromomethane |  | $<25$ | $<5$ | $<5$ | $<5$ | $<5$ | $<5$ | $<5$ | $<5$ | <5 | $<5$ |
| n-Butylbenzene |  | $<25$ | $<5$ | $<5$ | <5 | $<5$ | <5 | <5 | $<5$ | $<5$ | $<5$ |
| sec-Butylbenzene |  | $<25$ | $<5$ | $<5$ | $<5$ | $<5$ | $<5$ | $<5$ | $<5$ | $<5$ | $<5$ |
| tert-Butylbenzene |  | $<25$ | $<5$ | $<5$ | $<5$ | $<5$ | $<5$ | $<5$ | $<5$ | $<5$ | $<5$ |
| Carbon tetrachloride | 250 | e12 | $<5$ | $<5$ | $<5$ | $<5$ | $<5$ | $<5$ | $<5$ | $<5$ | $<5$ |
| Chlorobenzene |  | <25 | $<5$ | $<5$ | $<5$ | $<5$ | $<5$ | $<5$ | $<5$ | $<5$ | $<5$ |
| Chloroethane | 1,500,000 | e7 | <5 | $<5$ | $<5$ | $<5$ | $<5$ | $<5$ | $<5$ | $<5$ | $<5$ |
| 2-Chorotoluene |  | $<25$ | <5 | $<5$ | <5 | <5 | <5 | <5 | <5 | <5 | <5 |
| 4-Chlorotoluene |  | $<25$ | $<5$ | $<5$ | $<5$ | <5 | $<5$ | $<5$ | $<5$ | $<5$ | $<5$ |
| Dibromochloromethane |  | $<25$ | $<5$ | $<5$ | <5 | <5 | $<5$ | <5 | <5 | $<5$ | $<5$ |
| 1,2-Dichlorobenzene |  | $<25$ | $<5$ | $<5$ | <5 | <5 | <5 | <5 | <5 | $<5$ | $<5$ |
| 1,3-Dichlorobenzene |  | $<25$ | $<5$ | $<5$ | <5 | $<5$ | $<5$ | $<5$ | $<5$ | $<5$ | $<5$ |
| 1,4-Dichlorobenzene |  | $<25$ | <5 | $<5$ | <5 | <5 | <5 | <5 | $<5$ | <5 | <5 |
| Dichlorodifluoromethane |  | <25 | <5 | <5 | < 5 | <5 | <5 | <5 | <5 | <5 | <5 |
| 1,1-Dichloromethane |  | 39 | $<5$ | $<5$ | $<5$ | <5 | $<5$ | $<5$ | $<5$ | $<5$ | $<5$ |
| 1,2-Dichloroethane | 450 | e24 | $<5$ | $<5$ | <5 | <5 | $<5$ | <5 | $<5$ | $<5$ | $<5$ |
| 1,1-Dichloroehtene |  | <25 | $<5$ | $<5$ | $<5$ | $<5$ | $<5$ | $<5$ | $<5$ | $<5$ | $<5$ |
| cis-1,2-Dichloroethene | 78,000 | e24 | <5 | <5 | <5 | <5 | <5 | <5 | <5 | <5 | <5 |
| trans-1,2-Dichloroethene |  | $<25$ | <5 | $<5$ | <5 | <5 | <5 | <5 | <5 | <5 | <5 |
| 1,2-Dichloropropane |  | $<25$ | $<5$ | $<5$ | $<5$ | $<5$ | $<5$ | $<5$ | $<5$ | $<5$ | $<5$ |
| 1,3-Dichloropropane |  | $<25$ | $<5$ | $<5$ | $<5$ | <5 | $<5$ | $<5$ | $<5$ | $<5$ | $<5$ |
| 2,2-Dichloropropane |  | $<25$ | $<5$ | $<5$ | $<5$ | <5 | $<5$ | $<5$ | $<5$ | $<5$ | $<5$ |
| 1,1-Dichloropropene |  | $<25$ | $<5$ | $<5$ | $<5$ | $<5$ | $<5$ | $<5$ | $<5$ | $<5$ | $<5$ |
| Ethylbenzene |  | $<25$ | $<5$ | $<5$ | $<5$ | <5 | <5 | $<5$ | $<5$ | <5 | $<5$ |
| Hexachlorobutadiene |  | $<25$ | $<5$ | $<5$ | <5 | <5 | <5 | <5 | $<5$ | $<5$ | $<5$ |
| Isopropylbenzene |  | $<25$ | <5 | $<5$ | <5 | <5 | <5 | <5 | $<5$ | <5 | $<5$ |
| p-Isopropyltoluene |  | $<25$ | $<5$ | $<5$ | $<5$ | <5 | <5 | <5 | <5 | <5 | $<5$ |
| Methylene chloride | 11,000 | b25 ${ }^{2}$ | b7 | eb1 ${ }^{1,2}$ | eb2 | eb1 | eb1 | eb2 ${ }^{2}$ | eb1 | eb1 | eb2 |
| Naphthalene | 3,900 | $<25$ | e2 | $<5$ | <5 | <5 | <5 | <5 | <5 | <5 | <5 |
| n-Propylbenzene |  | $<25$ | $<5$ | $<5$ | $<5$ | $<5$ | $<5$ | $<5$ | $<5$ | $<5$ | $<5$ |
| Styrene |  | $<25$ | $<5$ | $<5$ | $<5$ | <5 | $<5$ | $<5$ | $<5$ | <5 | <5 |
| 1,1,1,2-Tetrachloroethane | 2,000 | 310 | <5 | <5 | <5 | <5 | <5 | <5 | <5 | <5 | <5 |

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4835 Glenbrook Road Soil Sampling
Spring Valley FUDS

| SAMPLE NUMBER--> | RSL | 1002 | 9000 | 9005 | 9006 | 9007 | 9008 | 9009 | 9010 | 9011 | 9012 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1,1,2,2-Tetrachloroethane | 590 | <25 | <5 | <5 | <5 | 380 | <5 | <5 | <5 | <5 | <5 |
| Tetrachloroethene |  | $<25$ | <5 | <5 | <5 | <5 | <5 | $<5$ | <5 | <5 | <5 |
| Toluene | 500,000 | e13 | <5 | <5 | <5 | <5 | <5 | <5 | <5 | <5 | <5 |
| 1,2,3-Trichlorobenzene |  | $<25$ | <5 | <5 | <5 | <5 | <5 | <5 | <5 | <5 | <5 |
| 1,2,4-Trichlorobenzene |  | $<25$ | <5 | <5 | <5 | <5 | <5 | <5 | <5 | <5 | <5 |
| 1,1,1-Trichloroethane |  | <25 | <5 | <5 | <5 | <5 | <5 | <5 | <5 | <5 | <5 |
| 1,1,2-Trichloroethane | 1,100 | 1,300 | $<5$ | $<5$ | $<5$ | 320 | <5 | $<5$ | $<5$ | $<5$ | $<5$ |
| Trichloroethene | 2,800 | e17 | <5 | <5 | <5 | e2 | <5 | $<5$ | $<5$ | <5 | <5 |
| Trichlorofluoromethane |  | $<25$ | $<5$ | $<5$ | $<5$ | $<5$ | $<5$ | <5 | $<5$ | $<5$ | $<5$ |
| 1,2,3-Trichloropropane |  | $<25$ | $<5$ | <5 | <5 | $<5$ | <5 | <5 | <5 | <5 | $<5$ |
| 1,2,4-Trimethylbenzene | NL | $<25$ | <5 | $<5$ | <5 | $<5$ | <5 | eb2 | eb4 | eb3 | eb4 |
| 1,3,5-Trimethylbenzene | NL | $<25$ | $<5$ | <5 | <5 | $<5$ | <5 | <5 | e1 | e1 | e2 |
| Vinyl chloride |  | $<25$ | $<5$ | $<5$ | $<5$ | $<5$ | $<5$ | $<5$ | <5 | <5 | <5 |
| Xylenes, total |  | $<75$ | $<15$ | <15 | $<15$ | $<15$ | $<15$ | $<15$ | $<15$ | $<15$ | $<15$ |
| SVOCs (ug/kg) - 8270 |  |  |  |  |  |  |  |  |  |  |  |
| Acenaphthene |  | <660 | <3,300 | NS | NS | NS | NS | NS | NS | NS | NS |
| Acenaphthylene |  | <660 | $<3,300$ | NS | NS | NS | NS | NS | NS | NS | NS |
| Anthracene |  | $<660$ | $<3,300$ | NS | NS | NS | NS | NS | NS | NS | NS |
| Benz(a)anthracene |  | <660 | $<3,300$ | NS | NS | NS | NS | NS | NS | NS | NS |
| Benzo(b)fluoranthene |  | <660 | <3,300 | NS | NS | NS | NS | NS | NS | NS | NS |
| Benzo(k)fluoranthene |  | $<660$ | $<3,300$ | NS | NS | NS | NS | NS | NS | NS | NS |
| Benzoic acid |  | <3,300 | <16500 | NS | NS | NS | NS | NS | NS | NS | NS |
| Benzo(g,h,i)perylene |  | $<660$ | $<3,300$ | NS | NS | NS | NS | NS | NS | NS | NS |
| Benzo(a)pyrene |  | $<660$ | $<3,300$ | NS | NS | NS | NS | NS | NS | NS | NS |
| Benzyl alcohol |  | <1,300 | $<3,300$ | NS | NS | NS | NS | NS | NS | NS | NS |
| Bis(2-chloroethoxy)methane | NL | e1041 | $<3,300$ | NS | NS | NS | NS | NS | NS | NS | NS |
| Bis(2-chloroethyl)ether |  | <660 | <3,300 | NS | NS | NS | NS | NS | NS | NS | NS |
| Bix(2-chloroisopropyl)ether |  | <660 | $<3,300$ | NS | NS | NS | NS | NS | NS | NS | NS |
| 4-bromophenyl phenyl ether |  | <660 | <3,300 | NS | NS | NS | NS | NS | NS | NS | NS |
| Butyl benzyl phthalate |  | $<660$ | $<3,300$ | NS | NS | NS | NS | NS | NS | NS | NS |
| 4-Chloroaniline |  | <1,300 | $<6,500$ | NS | NS | NS | NS | NS | NS | NS | NS |
| 4-Chloro-3-methylphenol |  | <1,300 | <6,500 | NS | NS | NS | NS | NS | NS | NS | NS |
| 2-Chloronaphthalene |  | <660 | $<3,300$ | NS | NS | NS | NS | NS | NS | NS | NS |
| 2-Chlorophenol |  | <660 | <3,300 | NS | NS | NS | NS | NS | NS | NS | NS |
| 4-Chlorophenyl phenyl ether |  | <660 | <3,300 | NS | NS | NS | NS | NS | NS | NS | NS |
| Chrysene |  | $<660$ | $<3,300$ | NS | NS | NS | NS | NS | NS | NS | NS |
| Dibenz(a,h)anthracene |  | <660 | $<3,300$ | NS | NS | NS | NS | NS | NS | NS | NS |
| Di-n-butyl-phthalate |  | <660 | <3,300 | NS | NS | NS | NS | NS | NS | NS | NS |

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Spring Valley FUDS

| SAMPLE NUMBER--> | RSL | 1002 | 9000 | 9005 | 9006 | 9007 | 9008 | 9009 | 9010 | 9011 | 9012 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Dibenzofuran |  | <660 | <3,300 | NS | NS | NS | NS | NS | NS | NS | NS |
| 1,2-Dichlorobenzene |  | <660 | <3,300 | NS | NS | NS | NS | NS | NS | NS | NS |
| 1,3-Dichlorobenzene |  | <660 | <3,300 | NS | NS | NS | NS | NS | NS | NS | NS |
| 1.4-Dichlorobenzene |  | <660 | <3,300 | NS | NS | NS | NS | NS | NS | NS | NS |
| 3,3-Dichlorobenzidine |  | <1,300 | <6,500 | NS | NS | NS | NS | NS | NS | NS | NS |
| 2,4-Dichlorophenol |  | <660 | <3,300 | NS | NS | NS | NS | NS | NS | NS | NS |
| Diethyl phthalate |  | $<660$ | $<3,300$ | NS | NS | NS | NS | NS | NS | NS | NS |
| 2,4-Dimethylphenol |  | <660 | <3,300 | NS | NS | NS | NS | NS | NS | NS | NS |
| Dimethyl phthalate |  | <660 | $<3,300$ | NS | NS | NS | NS | NS | NS | NS | NS |
| 4.6-Dinitro-2-methylphenol |  | $<3,300$ | $<16,500$ | NS | NS | NS | NS | NS | NS | NS | NS |
| 2,4-Dinitrophenol |  | <3,300 | <16,500 | NS | NS | NS | NS | NS | NS | NS | NS |
| 2,4-Dinitrotoluene |  | <660 | $<3,300$ | NS | NS | NS | NS | NS | NS | NS | NS |
| 2,6-Dinitrotoluene |  | <660 | $<3,300$ | NS | NS | NS | NS | NS | NS | NS | NS |
| Di-n-octyl phthalate |  | $<660$ | $<3,300$ | NS | NS | NS | NS | NS | NS | NS | NS |
| Bis(2-ethylhexyl)phthalate |  | <660 | $<3,300$ | NS | NS | NS | NS | NS | NS | NS | NS |
| Fluoranthene |  | <660 | <3,300 | NS | NS | NS | NS | NS | NS | NS | NS |
| Fluorene |  | $<660$ | <3,300 | NS | NS | NS | NS | NS | NS | NS | NS |
| Hexachlorobenzene | 300 | <660 | 12,000 | NS | NS | NS | NS | NS | NS | NS | NS |
| Hexachlorobutadine |  | <660 | <3,300 | NS | NS | NS | NS | NS | NS | NS | NS |
| Hexachlorocyclopentadiene |  | <660 | <3,300 | NS | NS | NS | NS | NS | NS | NS | NS |
| Hexachloroethane |  | <660 | $<3,300$ | NS | NS | NS | NS | NS | NS | NS | NS |
| Indeno(1,2,3-cd)pyrene |  | <660 | $<3,300$ | NS | NS | NS | NS | NS | NS | NS | NS |
| Isophorone |  | <660 | <3,300 | NS | NS | NS | NS | NS | NS | NS | NS |
| 2-Methylnaphthalene |  | <660 | <3,300 | NS | NS | NS | NS | NS | NS | NS | NS |
| 2-Mehylphenol |  | <660 | <3,300 | NS | NS | NS | NS | NS | NS | NS | NS |
| 4-Methylphenol |  | $<660$ | $<3,300$ | NS | NS | NS | NS | NS | NS | NS | NS |
| 4-Methylphenol |  | <660 | $<3,300$ | NS | NS | NS | NS | NS | NS | NS | NS |
| Naphthalene |  | $<660$ | $<3,300$ | NS | NS | NS | NS | NS | NS | NS | NS |
| 2-Nitroaniline |  | $<3,300$ | $<16,500$ | NS | NS | NS | NS | NS | NS | NS | NS |
| 4-Nitrophenol |  | <3,300 | $<16,500$ | NS | NS | NS | NS | NS | NS | NS | NS |
| N-Nitrosodiphenylamine |  | <660 | <3,300 | NS | NS | NS | NS | NS | NS | NS | NS |
| N-Nitroso-di-n-propylamine |  | <660 | <3,300 | NS | NS | NS | NS | NS | NS | NS | NS |
| Pentachlorophenol |  | <3,300 | $<16,500$ | NS | NS | NS | NS | NS | NS | NS | NS |
| Phenanthrene |  | <660 | $<3,300$ | NS | NS | NS | NS | NS | NS | NS | NS |
| Phenol |  | <660 | $<3,300$ | NS | NS | NS | NS | NS | NS | NS | NS |
| Pyrene |  | <660 | $<3,300$ | NS | NS | NS | NS | NS | NS | NS | NS |
| 1,2,4-Trichlorobenzene |  | <660 | $<3,300$ | NS | NS | NS | NS | NS | NS | NS | NS |
| 2,4,5-Trichlorophenol |  | <660 | $<3,300$ | NS | NS | NS | NS | NS | NS | NS | NS |
| 2,4,6-Trichlorophenol |  | <660 | <3,300 | NS | NS | NS | NS | NS | NS | NS | NS |

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4835 Glenbrook Road Soil Sampling
Spring Valley FUDS

| SAMPLE NUMBER--> | RSL | 1002 | 9000 | 9005 | 9006 | 9007 | 9008 | 9009 | 9010 | 9011 | 9012 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Metals (mg/kg) - PP |  |  |  |  |  |  |  |  |  |  |  |
| Antimony | 5.2(BG) | 40.0 | ND | ND | ND | ND | ND | ND | ND | ND | ND |
| Arsenic | 20(SV) | 1,200 | 9.5 | 0.6 | 2.1 | 5.3 | 25.0 | 15.0 | 2.6 | 6.7 | 2.7 |
| Beryllium | 16 | 1.3 | ND | 0.6 | 0.5 | ND | 0.5 | 0.9 | ND | 0.8 | 0.9 |
| Cadmium |  | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND |
| Chromium | 12,000 | 70.0 | 2.1 | 2.5 | 6.2 | 3.4 | 4.9 | 4.1 | ND | 4.6 | 1.5 |
| Copper | 310 | 35.5 | 10.0 | 5.1 | 9.1 | 4.3 | 8.3 | 5.9 | 1.6 | 6.4 | 3.9 |
| Lead | 400 | 102.0 | 7.6 | 3.9 | 5.3 | 3.8 | 42.0 | 4.1 | 35.0 | 5.9 | 0.9 |
| Mercury |  | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND |
| Nickel | 160 | 27.0 | 0.9 | 3.7 | 4.7 | 2.2 | 2.9 | 4.2 | 0.8 | 3.7 | 1.1 |
| Selenium |  | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND |
| Silver |  | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND |
| Thallium |  | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND |
| Zinc | 2,300 | 180.0 | 11.0 | 7.7 | 11.0 | 6.7 | 24.0 | 11.0 | 5.6 | 9.8 | 3.9 |
| PCB Type (mg/kg) |  |  |  |  |  |  |  |  |  |  |  |
| 1016 |  | ND | ND | NS | NS | NS | NS | NS | NS | NS | NS |
| 1221 |  | ND | ND | NS | NS | NS | NS | NS | NS | NS | NS |
| 1232 |  | ND | ND | NS | NS | NS | NS | NS | NS | NS | NS |
| 1242 |  | ND | ND | NS | NS | NS | NS | NS | NS | NS | NS |
| 1248 |  | ND | ND | NS | NS | NS | NS | NS | NS | NS | NS |
| 1254 |  | ND | ND | NS | NS | NS | NS | NS | NS | NS | NS |
| 1260 |  | ND | ND | NS | NS | NS | NS | NS | NS | NS | NS |
| Pesticide Type (ug/kg) |  |  |  |  |  |  |  |  |  |  |  |
| Aldrin |  | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND |
| A-BHC |  | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND |
| b-BHC |  | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND |
| g-BHC (Lindane) |  | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND |
| d-BHC |  | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND |
| Chlordane (total) |  | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND |
| 4,4 ${ }^{1}$-DDD |  | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND |
| $4,4^{1}-\mathrm{DDE}$ |  | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND |
| 4,4 ${ }^{1}$-DDT |  | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND |
| Dieldrin |  | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND |
| Endosulfan I |  | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND |
| Endosulfan II |  | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND |
| Endosulfan sulfate |  | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND |
| Endrin |  | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND |
| Heptachlor |  | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND |

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Spring Valley FUDS

| SAMPLE NUMBER--> | RSL | 9013 | 9014 | 9015 | 9016 | 9017 | 9018 | 9019 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| VOCs (ug/kg) - 8260 | Non-Carcs | adjusted downward |  |  |  |  |  |  |
| Benzene | 1,100 | <5 | <5 | < 5 | $<5$ | <5 | <5 | <5 |
| Bromobenzene |  | <5 | <5 | <5 | <5 | <5 | <5 | <5 |
| Bromochloromethane |  | $<5$ | $<5$ | $<5$ | $<5$ | $<5$ | $<5$ | $<5$ |
| Bromodichloromethane |  | $<5$ | $<5$ | $<5$ | $<5$ | <5 | $<5$ | <5 |
| Bromoform |  | <5 | <5 | <5 | <5 | <5 | <5 | <5 |
| Bromomethane |  | <5 | $<5$ | <5 | <5 | <5 | $<5$ | <5 |
| n-Butylbenzene |  | <5 | <5 | <5 | <5 | <5 | <5 | <5 |
| sec-Butylbenzene |  | <5 | <5 | <5 | $<5$ | <5 | $<5$ | <5 |
| tert-Butylbenzene |  | $<5$ | $<5$ | $<5$ | $<5$ | $<5$ | $<5$ | $<5$ |
| Carbon tetrachloride | 250 | $<5$ | $<5$ | $<5$ | $<5$ | $<5$ | $<5$ | $<5$ |
| Chlorobenzene |  | <5 | <5 | $<5$ | <5 | <5 | <5 | <5 |
| Chloroethane | 1,500,000 | <5 | <5 | $<5$ | <5 | <5 | $<5$ | <5 |
| 2-Chorotoluene |  | <5 | <5 | <5 | <5 | <5 | <5 | <5 |
| 4-Chlorotoluene |  | <5 | <5 | $<5$ | $<5$ | $<5$ | $<5$ | $<5$ |
| Dibromochloromethane |  | $<5$ | $<5$ | $<5$ | $<5$ | <5 | <5 | $<5$ |
| 1,2-Dichlorobenzene |  | <5 | $<5$ | $<5$ | $<5$ | <5 | <5 | $<5$ |
| 1,3-Dichlorobenzene |  | <5 | <5 | <5 | <5 | <5 | <5 | <5 |
| 1,4-Dichlorobenzene |  | <5 | <5 | <5 | <5 | <5 | <5 | <5 |
| Dichlorodifluoromethane |  | <5 | <5 | <5 | <5 | <5 | <5 | <5 |
| 1,1-Dichloromethane |  | <5 | <5 | <5 | <5 | <5 | <5 | <5 |
| 1,2-Dichloroethane | 450 | <5 | <5 | <5 | <5 | <5 | <5 | <5 |
| 1,1-Dichloroehtene |  | <5 | <5 | <5 | <5 | <5 | <5 | <5 |
| cis-1,2-Dichloroethene | 78,000 | $<5$ | $<5$ | $<5$ | $<5$ | $<5$ | $<5$ | $<5$ |
| trans-1,2-Dichloroethene |  | $<5$ | $<5$ | $<5$ | $<5$ | $<5$ | $<5$ | $<5$ |
| 1,2-Dichloropropane |  | $<5$ | $<5$ | $<5$ | $<5$ | $<5$ | $<5$ | $<5$ |
| 1,3-Dichloropropane |  | $<5$ | $<5$ | $<5$ | $<5$ | <5 | $<5$ | $<5$ |
| 2,2-Dichloropropane |  | $<5$ | $<5$ | <5 | <5 | <5 | $<5$ | <5 |
| 1,1-Dichloropropene |  | <5 | <5 | <5 | <5 | <5 | <5 | <5 |
| Ethylbenzene |  | $<5$ | $<5$ | $<5$ | $<5$ | $<5$ | $<5$ | <5 |
| Hexachlorobutadiene |  | $<5$ | $<5$ | $<5$ | $<5$ | $<5$ | $<5$ | $<5$ |
| Isopropylbenzene |  | $<5$ | $<5$ | $<5$ | $<5$ | $<5$ | $<5$ | $<5$ |
| p-Isopropyltoluene |  | <5 | <5 | <5 | <5 | <5 | <5 | <5 |
| Methylene chloride | 11,000 | eb5 | eb3 | eb2 | eb3 | eb2 | eb2 | eb1 |
| Naphthalene | 3,900 | <5 | $<5$ | <5 | $<5$ | <5 | <5 | <5 |
| n-Propylbenzene |  | <5 | <5 | <5 | <5 | <5 | <5 | <5 |
| Styrene |  | <5 | <5 | <5 | <5 | <5 | <5 | <5 |
| 1,1,1,2-Tetrachloroethane | 2,000 | <5 | <5 | <5 | <5 | <5 | <5 | <5 |

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4835 Glenbrook Road Soil Sampling
Spring Valley FUDS

| SAMPLE NUMBER--> | RSL | 9013 | 9014 | 9015 | 9016 | 9017 | 9018 | 9019 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1,1,2,2-Tetrachloroethane | 590 | <5 | <5 | <5 | <5 | <5 | <5 | <5 |
| Tetrachloroethene |  | $<5$ | <5 | <5 | <5 | <5 | <5 | <5 |
| Toluene | 500,000 | <5 | e1 | <5 | e1 | <5 | e1 | e3 |
| 1,2,3-Trichlorobenzene |  | $<5$ | <5 | <5 | <5 | $<5$ | <5 | <5 |
| 1,2,4-Trichlorobenzene |  | <5 | <5 | <5 | <5 | <5 | <5 | <5 |
| 1,1,1-Trichloroethane |  | <5 | <5 | <5 | <5 | <5 | <5 | <5 |
| 1,1,2-Trichloroethane | 1,100 | $<5$ | $<5$ | $<5$ | <5 | $<5$ | $<5$ | <5 |
| Trichloroethene | 2,800 | <5 | <5 | <5 | <5 | <5 | <5 | <5 |
| Trichlorofluoromethane |  | $<5$ | $<5$ | $<5$ | $<5$ | $<5$ | < | $<5$ |
| 1,2,3-Trichloropropane |  | $<5$ | $<5$ | <5 | <5 | <5 | $<5$ | $<5$ |
| 1,2,4-Trimethylbenzene | NL | <5 | <5 | <5 | <5 | <5 | <5 | <5 |
| 1,3,5-Trimethylbenzene | NL | <5 | <5 | <5 | <5 | $<5$ | $<5$ | <5 |
| Vinyl chloride |  | $<5$ | $<5$ | $<5$ | $<5$ | $<5$ | $<5$ | $<5$ |
| Xylenes, total |  | <15 | <15 | $<15$ | <15 | $<15$ | <15 | <15 |
| SVOCs (ug/kg) - 8270 |  |  |  |  |  |  |  |  |
| Acenaphthene |  | NS | NS | NS | NS | NS | NS | NS |
| Acenaphthylene |  | NS | NS | NS | NS | NS | NS | NS |
| Anthracene |  | NS | NS | NS | NS | NS | NS | NS |
| Benz(a)anthracene |  | NS | NS | NS | NS | NS | NS | NS |
| Benzo(b)fluoranthene |  | NS | NS | NS | NS | NS | NS | NS |
| Benzo(k)fluoranthene |  | NS | NS | NS | NS | NS | NS | NS |
| Benzoic acid |  | NS | NS | NS | NS | NS | NS | NS |
| Benzo(g,h,i)perylene |  | NS | NS | NS | NS | NS | NS | NS |
| Benzo(a)pyrene |  | NS | NS | NS | NS | NS | NS | NS |
| Benzyl alcohol |  | NS | NS | NS | NS | NS | NS | NS |
| Bis(2-chloroethoxy)methane | NL | NS | NS | NS | NS | NS | NS | NS |
| Bis(2-chloroethyl)ether |  | NS | NS | NS | NS | NS | NS | NS |
| Bix(2-chloroisopropyl)ether |  | NS | NS | NS | NS | NS | NS | NS |
| 4-bromophenyl phenyl ether |  | NS | NS | NS | NS | NS | NS | NS |
| Butyl benzyl phthalate |  | NS | NS | NS | NS | NS | NS | NS |
| 4-Chloroaniline |  | NS | NS | NS | NS | NS | NS | NS |
| 4-Chloro-3-methylphenol |  | NS | NS | NS | NS | NS | NS | NS |
| 2-Chloronaphthalene |  | NS | NS | NS | NS | NS | NS | NS |
| 2-Chlorophenol |  | NS | NS | NS | NS | NS | NS | NS |
| 4-Chlorophenyl phenyl ether |  | NS | NS | NS | NS | NS | NS | NS |
| Chrysene |  | NS | NS | NS | NS | NS | NS | NS |
| Dibenz(a,h)anthracene |  | NS | NS | NS | NS | NS | NS | NS |
| Di-n-butyl-phthalate |  | NS | NS | NS | NS | NS | NS | NS |

Table A.3--APEX 1996
4835 Glenbrook Road Soil Sampling
Spring Valley FUDS

| SAMPLE NUMBER--> | RSL | 9013 | 9014 | 9015 | 9016 | 9017 | 9018 | 9019 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Dibenzofuran |  | NS | NS | NS | NS | NS | NS | NS |
| 1,2-Dichlorobenzene |  | NS | NS | NS | NS | NS | NS | NS |
| 1,3-Dichlorobenzene |  | NS | NS | NS | NS | NS | NS | NS |
| 1.4-Dichlorobenzene |  | NS | NS | NS | NS | NS | NS | NS |
| 3,3-Dichlorobenzidine |  | NS | NS | NS | NS | NS | NS | NS |
| 2,4-Dichlorophenol |  | NS | NS | NS | NS | NS | NS | NS |
| Diethyl phthalate |  | NS | NS | NS | NS | NS | NS | NS |
| 2,4-Dimethylphenol |  | NS | NS | NS | NS | NS | NS | NS |
| Dimethyl phthalate |  | NS | NS | NS | NS | NS | NS | NS |
| 4.6-Dinitro-2-methylphenol |  | NS | NS | NS | NS | NS | NS | NS |
| 2,4-Dinitrophenol |  | NS | NS | NS | NS | NS | NS | NS |
| 2,4-Dinitrotoluene |  | NS | NS | NS | NS | NS | NS | NS |
| 2,6-Dinitrotoluene |  | NS | NS | NS | NS | NS | NS | NS |
| Di-n-octyl phthalate |  | NS | NS | NS | NS | NS | NS | NS |
| Bis(2-ethylhexyl)phthalate |  | NS | NS | NS | NS | NS | NS | NS |
| Fluoranthene |  | NS | NS | NS | NS | NS | NS | NS |
| Fluorene |  | NS | NS | NS | NS | NS | NS | NS |
| Hexachlorobenzene | 300 | NS | NS | NS | NS | NS | NS | NS |
| Hexachlorobutadine |  | NS | NS | NS | NS | NS | NS | NS |
| Hexachlorocyclopentadiene |  | NS | NS | NS | NS | NS | NS | NS |
| Hexachloroethane |  | NS | NS | NS | NS | NS | NS | NS |
| Indeno(1,2,3-cd)pyrene |  | NS | NS | NS | NS | NS | NS | NS |
| Isophorone |  | NS | NS | NS | NS | NS | NS | NS |
| 2-Methylnaphthalene |  | NS | NS | NS | NS | NS | NS | NS |
| 2-Mehylphenol |  | NS | NS | NS | NS | NS | NS | NS |
| 4-Methylphenol |  | NS | NS | NS | NS | NS | NS | NS |
| 4-Methylphenol |  | NS | NS | NS | NS | NS | NS | NS |
| Naphthalene |  | NS | NS | NS | NS | NS | NS | NS |
| 2-Nitroaniline |  | NS | NS | NS | NS | NS | NS | NS |
| 4-Nitrophenol |  | NS | NS | NS | NS | NS | NS | NS |
| N-Nitrosodiphenylamine |  | NS | NS | NS | NS | NS | NS | NS |
| N-Nitroso-di-n-propylamine |  | NS | NS | NS | NS | NS | NS | NS |
| Pentachlorophenol |  | NS | NS | NS | NS | NS | NS | NS |
| Phenanthrene |  | NS | NS | NS | NS | NS | NS | NS |
| Phenol |  | NS | NS | NS | NS | NS | NS | NS |
| Pyrene |  | NS | NS | NS | NS | NS | NS | NS |
| 1,2,4-Trichlorobenzene |  | NS | NS | NS | NS | NS | NS | NS |
| 2,4,5-Trichlorophenol |  | NS | NS | NS | NS | NS | NS | NS |
| 2,4,6-Trichlorophenol |  | NS | NS | NS | NS | NS | NS | NS |

Table A.3--APEX 1996
4835 Glenbrook Road Soil Sampling
Spring Valley FUDS

| SAMPLE NUMBER--> | RSL | 9013 | 9014 | 9015 | 9016 | 9017 | 9018 | 9019 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Metals (mg/kg) - PP |  |  |  |  |  |  |  |  |
| Antimony | 5.2(BG) | ND | ND | ND | ND | ND | ND | ND |
| Arsenic | 20(SV) | 20.0 | ND | ND | ND | ND | ND | 2.9 |
| Beryllium | 16 | 1.0 | ND | ND | ND | 0.6 | ND | 0.6 |
| Cadmium |  | ND | ND | ND | ND | ND | ND | ND |
| Chromium | 12,000 | 0.9 | ND | 1.3 | 1.6 | 5.7 | 1.1 | 3.1 |
| Copper | 310 | 3.2 | 0.8 | 4.1 | 1.3 | 9.6 | 4.3 | 6.9 |
| Lead | 400 | 7.8 | 2.3 | 14.0 | 6.4 | 23.0 | 26.0 | 19.0 |
| Mercury |  | ND | ND | ND | ND | ND | ND | ND |
| Nickel | 160 | 1.3 | 1.1 | 2.9 | 1.0 | 12.0 | 1.2 | 4.1 |
| Selenium |  | ND | ND | ND | ND | ND | ND | ND |
| Silver |  | ND | ND | ND | ND | ND | ND | ND |
| Thallium |  | ND | ND | ND | ND | ND | ND | ND |
| Zinc | 2,300 | 28.0 | 6.5 | 19 | 4.3 | 42 | 9.4 | 19 |
| PCB Type (mg/kg) |  |  |  |  |  |  |  |  |
| 1016 |  | NS | NS | NS | NS | NS | NS | NS |
| 1221 |  | NS | NS | NS | NS | NS | NS | NS |
| 1232 |  | NS | NS | NS | NS | NS | NS | NS |
| 1242 |  | NS | NS | NS | NS | NS | NS | NS |
| 1248 |  | NS | NS | NS | NS | NS | NS | NS |
| 1254 |  | NS | NS | NS | NS | NS | NS | NS |
| 1260 |  | NS | NS | NS | NS | NS | NS | NS |
| Pesticide Type (ug/kg) |  |  |  |  |  |  |  |  |
| Aldrin |  | ND | NS | NS | NS | NS | NS | NS |
| A-BHC |  | ND | NS | NS | NS | NS | NS | NS |
| b-BHC |  | ND | NS | NS | NS | NS | NS | NS |
| g-BHC (Lindane) |  | ND | NS | NS | NS | NS | NS | NS |
| d-BHC |  | ND | NS | NS | NS | NS | NS | NS |
| Chlordane (total) |  | ND | NS | NS | NS | NS | NS | NS |
| 4,4 ${ }^{1}$-DDD |  | ND | NS | NS | NS | NS | NS | NS |
| 4,4 ${ }^{1}$-DDE |  | ND | NS | NS | NS | NS | NS | NS |
| 4,4 ${ }^{1}$-DDT |  | ND | NS | NS | NS | NS | NS | NS |
| Dieldrin |  | ND | NS | NS | NS | NS | NS | NS |
| Endosulfan I |  | ND | NS | NS | NS | NS | NS | NS |
| Endosulfan II |  | ND | NS | NS | NS | NS | NS | NS |
| Endosulfan sulfate |  | ND | NS | NS | NS | NS | NS | NS |
| Endrin |  | ND | NS | NS | NS | NS | NS | NS |
| Heptachlor |  | ND | NS | NS | NS | NS | NS | NS |

Table A.3--APEX 1996
4835 Glenbrook Road Soil Sampling
Spring Valley FUDS

| SAMPLE NUMBER--> | RSL | 9013 | 9014 | 9015 | 9016 | 9017 | 9018 | 9019 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Heptachlor epoxide |  | ND | NS | NS | NS | NS | NS | NS |
| Methoxychlor |  | ND | NS | NS | NS | NS | NS | NS |
| Toxaphene |  | ND | NS | NS | NS | NS | NS | NS |
| Herbicide Type (ug/kg) |  |  |  |  |  |  |  |  |
| Dalapon |  | ND | NS | NS | NS | NS | NS | NS |
| Dicamba |  | ND | NS | NS | NS | NS | NS | NS |
| MCPP |  | ND | NS | NS | NS | NS | NS | NS |
| MCPA |  | ND | NS | NS | NS | NS | NS | NS |
| Dichloroprop |  | ND | NS | NS | NS | NS | NS | NS |
| 2,4-D |  | ND | NS | NS | NS | NS | NS | NS |
| 2,4,5-TP (silvex) |  | ND | NS | NS | NS | NS | NS | NS |
| 2,4,5-T |  | ND | NS | NS | NS | NS | NS | NS |
| Dinoseb |  | ND | NS | NS | NS | NS | NS | NS |
| 2,4-DB |  | ND | NS | NS | NS | NS | NS | NS |
|  |  |  |  |  |  |  |  |  |
| * Sample represents soil that was removed in1996 . |  |  | $\mathrm{b}=$ Not found substantially above level in blank. |  |  |  |  |  |
| ND = Not detected |  |  | $e=$ estimated value below reporting limit. "J" flag equivalent. |  |  |  |  |  |
| NS = Not sampled for this parameter.NL = Not Listed. |  |  | RSLs are EPA Regional Screening Levels (12 September 2008). |  |  |  |  |  |
| NL $=$ Not Listed. |  |  | RSLs are shown only for detected chemicals. |  |  |  |  |  |
| SV - Spring Valley Remediation Goal. |  |  | BG - 95th percentile of the 2007 Background Study. |  |  |  |  |  |
| Exceeds the standard. |  |  | This used when it is higher than the RSL. |  |  |  |  |  |
| Sample has been removed |  |  | SOURCE: APEX Environmental Inc., 4835 Glenbrook Road |  |  |  |  |  |
|  |  |  | FINAL Report, August 6, 1996. |  |  |  |  |  |
|  |  |  |  |  |  |  |  |  |
|  |  |  |  |  |  |  |  |  |
|  |  |  |  |  |  |  |  |  |

Table A.4--USEPA 1999

## 4835 Glenbrook Road Soil Sampling

Spring Valley FUDS

|  | RSL (mg/kg) | Background (mg/kg) | G-01 | G-02 | G-03 | OU3-SB02 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| VOCs - CLP | Non-Carc RSLs <br> Adjusted <br> Down |  |  |  |  |  |
| Chloromethane | 1.7 |  | ND | ND | ND | NS |
| Bromomethane | 0.79 |  | ND | ND | ND | NS |
| Vinyl chloride | 0.06 |  | ND | ND | ND | NS |
| Chloroethane | 1500 |  | ND | ND | ND | NS |
| Methylene Chloride | 11 |  | ND | ND | ND | NS |
| Acetone | 6100 |  | ND | ND | ND | NS |
| Carbon disulfide | 67 |  | ND | ND | ND | NS |
| 1,1-Dichloroethene | 25 |  | ND | ND | ND | NS |
| 1,1-Dichloroethane | 3.4 |  | ND | ND | ND | NS |
| 1,2-Dichloroethene (total) | 0.7 |  | ND | ND | ND | NS |
| Chloroform | 0.3 |  | ND | ND | ND | NS |
| 1,2-Dichloroethane | 0.45 |  | ND | ND | ND | NS |
| 2-Butanone (MEK) | 2,800 |  | ND | ND | ND | NS |
| 1,1,1-Trichloroethane | 900 |  | ND | ND | ND | NS |
| Carbon tetrachloride | 0.25 |  | ND | ND | ND | NS |
| Bromodichloromethane | 10 |  | ND | ND | ND | NS |
| 1,2-Dichloropropane | 9.3 |  | ND | ND | ND | NS |
| 1,3-Dichloropropene | 78 |  | ND | ND | ND | NS |
| Trichloroethene | 2.8 |  | ND | ND | ND | NS |
| Dibromochloromethane | 5.8 |  | ND | ND | ND | NS |
| 1,1,2-Trichloroethane | 1.1 |  | ND | ND | ND | NS |
| Benzene | 1.1 |  | ND | ND | ND | NS |
| Bromoform | 61 |  | ND | ND | ND | NS |
| 4-Methyl-2-pentatone | 5300 |  | ND | ND | ND | NS |
| 2-Hexanone | NL |  | ND | ND | ND | NS |
| Tetrachloroethene | 0.57 |  | ND | ND | ND | NS |
| 1,1,2,2-Tetrachloroethane | 0.59 |  | ND | ND | ND | NS |
| Toluene | 500 |  | ND | ND | ND | NS |
| Chlorobenzene | 31 |  | ND | ND | ND | NS |
| Ethylbenzene | 5.7 |  | ND | ND | ND | NS |
| Styrene | 650 |  | ND | ND | ND | NS |
| Xylene (total) | 60 |  | ND | ND | ND | NS |
| SVOCs - CLP |  |  |  |  |  |  |
| Phenol | 1,800 |  | ND | ND | ND | NS |
| bis(2-Chloroethyl) ether | 0.19 |  | ND | ND | ND | NS |
| 2-Chlorophenol | 39 |  | 0.380 UJ | ND | ND | NS |
| 1.3-Dichlorobenzene | 230 |  | 0.380 UJ | ND | ND | NS |
| 1,4-Dichlorobenzene | 2.6 |  | 0.380 UJ | ND | ND | NS |
| 1,2-Dichlorobenzene | 200 |  | 0.380 UJ | ND | ND | NS |
| 2-Methylphenol | 310 |  | 0.380 UJ | ND | ND | NS |
| 2,2'-oxybis(1-Chloropropane) | 3100 |  | 0.380 UJ | ND | ND | NS |
| 4-Methylphenol | 31 |  | 0.380 UJ | ND | ND | NS |
| N-nitroso-di-n-propylamine | 0.069 |  | 0.380 UJ | ND | ND | NS |
| Hexachloroethane | 61 |  | 0.380 UJ | ND | ND | NS |
| Nitrobenzene | 31 |  | 0.380 UJ | ND | ND | NS |
| Isophorone | 510 |  | 0.380 UJ | ND | ND | NS |
| 2-Nitrophenol | NL |  | 0.380 UJ | ND | ND | NS |
| 2,4-Dimethylphenol | 120 |  | 0.380 UJ | ND | ND | NS |
| bis(2-chloroethoxy) methane | NL |  | 0.380 UJ | ND | ND | NS |
| 2,4-Dichlorophenol | 18 |  | 0.380 UJ | ND | ND | NS |
| 1,2,4-Trichlorobenzene | 8.7 |  | 0.380 UJ | ND | ND | NS |
| Naphthalene | 3.9 |  | 0.380 UJ | ND | ND | NS |
| 4-Chloroaniline | 9 |  | 0.380 UJ | ND | ND | NS |
| Hexachlorobutadiene | 6.2 |  | 0.380 UJ | ND | ND | NS |
| 4-Chloro-3-methylphenol | NL |  | 0.380 UJ | ND | ND | NS |

Table A.4--USEPA 1999

## 4835 Glenbrook Road Soil Sampling

Spring Valley FUDS

|  | RSL (mg/kg) | Background (mg/kg) | G-01 | G-02 | G-03 | OU3-SB02 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 2-Methylnaphthalene | 31 |  | 0.380 UJ | ND | ND | NS |
| Hexachlorocyclopentadiene | 37 |  | 0.380 UJ | ND | ND | NS |
| 2,4,6-Trichlorophenol | 44 |  | 0.380 UJ | ND | ND | NS |
| 2,4,5-Trichlorophenol | 610 |  | 0.940 UJ | ND | ND | NS |
| 2-Chloronaphthalene | 630 |  | 0.380 UJ | ND | ND | NS |
| 2-Nitroaniline | 23 |  | 0.940 UJ | ND | ND | NS |
| Dimethylphthalate | 78,000 |  | 0.380 UJ | ND | ND | NS |
| Acenaphthalene | NL |  | 0.380 UJ | ND | ND | NS |
| 2,6-Dinitrotoluene | 6.1 |  | 0.380 UJ | ND | ND | NS |
| 3-Nitroaniline | 1.8 |  | 0.940 UJ | ND | ND | NS |
| Acenaphthene | 340 |  | 0.380 UJ | ND | ND | NS |
| 2.4-Dinitrophenol | 12 |  | 0.940 UJ | 0.990 UJ | 1.100 UJ | NS |
| 4-Nitrophenol | 63 |  | 0.940 UJ | ND | ND | NS |
| Dibenzofuran | 7.8 |  | 0.380 UJ | ND | ND | NS |
| 2,4-Dinitrotoluene | 12 |  | 0.380 UJ | ND | ND | NS |
| Diethylphthalate | 4,900 |  | 0.380 UJ | ND | ND | NS |
| 4-Chlorophenyl-phenylether | NL |  | 0.380 UJ | ND | ND | NS |
| Fluorene | 230 |  | 0.380 UJ | ND | ND | NS |
| 4-Nitroaniline | 23 |  | 0.940 UJ | ND | ND | NS |
| 4,6-Dinitro-2-methylphenol | 0.61 |  | 0.940 UJ | 0.990 UJ | 1.100 UJ | NS |
| N-Nitrosodiphenylamine | 99 |  | 0.380 UJ | ND | ND | NS |
| 4-Bromophenyl-phenylether | NL |  | 0.380 UJ | ND | ND | NS |
| Hexachlorobenzene | 0.3 |  | 0.380 UJ | ND | ND | NS |
| Pentachlorophenol | 3 |  | 0.940 UJ | ND | ND | NS |
| Phenanthrene | NL |  | 0.380 UJ | ND | ND | NS |
| Anthracene | 1,700 |  | 0.380 UJ | ND | ND | NS |
| Carbazole | 32 |  | 0.380 UJ | ND | ND | NS |
| Di-n-butylphthalate | 610 |  | 0.380 UJ | 0.061 B | ND | NS |
| Fluoranthene | 230 |  | 0.055 J | ND | 0.005 J | NS |
| Pyrene | 170 |  | 0.048 J | ND | ND | NS |
| Butylbenzylphthalate | 260 |  | 0.380 UJ | ND | ND | NS |
| 3,3'-Dichlorobenzidine | 1.1 |  | 0.380 UJ | ND | ND | NS |
| Benzo(a)anthracene | 0.15 |  | 0.380 UJ | ND | ND | NS |
| Chrysene | 15 |  | 0.380 UJ | ND | ND | NS |
| bis(2-Ethylhexyl)phthalate | 35 |  | 0.045 J | 0.044 J | ND | NS |
| Di-n-octylphthalate | 310 |  | 0.380 UJ | ND | ND | NS |
| Benzo(b)fluoranthene | 0.15 |  | 0.380 UJ | ND | ND | NS |
| Benzo(k)fluoranthene | 1.5 |  | 0.380 UJ | ND | ND | NS |
| Benzo(a)pyrene | 0.015 |  | 0.380 UJ | ND | ND | NS |
| Indeno(1,2,3-cd)pyrene | 0.15 |  | 0.380 UJ | ND | ND | NS |
| Dibenz(a,h)anthracene | 0.015 |  | 0.380 UJ | ND | ND | NS |
| Benzo(g,h,i)perylene | NL |  | 0.380 UJ | ND | ND | NS |
| Pest/PCBs |  |  |  |  |  |  |
| Alpha-BHC | 0.077 |  | ND | ND | ND | NS |
| Beta-BHC | 0.27 |  | ND | ND | ND | NS |
| Delta-BHC | NL |  | ND | ND | ND | NS |
| Gamma-BHC (Lindane) | 21 |  | ND | ND | ND | NS |
| Heptachlor | 31 |  | ND | ND | ND | NS |
| Aldrin | 1.8 |  | ND | ND | ND | NS |
| Heptachlor-epoxide | 0.79 |  | ND | ND | 0.0023 J | NS |
| Endosulfan I | NL |  | ND | ND | ND | NS |
| Dieldrin | 3.1 |  | ND | ND | ND | NS |
| 4,4'-DDE | 1.4 |  | ND | ND | ND | NS |
| Endrin | 18 |  | ND | ND | ND | NS |
| Endosulfan II | NL |  | ND | ND | ND | NS |
| 4,4'-DDD | 2 |  | ND | ND | ND | NS |
| Endosulfan sulfate | NL |  | ND | ND | ND | NS |
| 4,4'-DDT | 36 |  | ND | ND | 0.0031 J | NS |
| Methoxychlor | 310 |  | ND | ND | ND | NS |

Table A.4--USEPA 1999

## 4835 Glenbrook Road Soil Sampling Spring Valley FUDS

|  | RSL (mg/kg) | Background (mg/kg) | G-01 | G-02 | G-03 | OU3-SB02 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Endrin ketone | NL |  | ND | ND | ND | NS |
| Endrin aldehyde | NL |  | ND | ND | ND | NS |
| Alpha-Chlordame | 35.00 |  | 0.0018 J | ND | 0.0079 J | NS |
| gamma-Chlordane | 35.00 |  | 0.0019 J | ND | 0.0084 | NS |
| Toxaphene | 0.44 |  | ND | ND | ND | NS |
| Aroclor-1016 | 6.3 |  | ND | ND | ND | NS |
| Aroclor-1221 | 0.17 |  | ND | ND | ND | NS |
| Aroclor-1232 | 0.17 |  | ND | ND | ND | NS |
| Aroclor-1242 | 0.22 |  | ND | ND | ND | NS |
| Aroclor-1248 | 0.22 |  | ND | ND | ND | NS |
| Aroclor-1254 | 0.22 |  | ND | ND | ND | NS |
| Aroclor-1260 | 0.22 |  | ND | ND | ND | NS |
| Metals - TAL |  |  |  |  |  |  |
| Aluminum | 19,100 | 19,100 | 18,700 | 17,200 | 14,200 | 22,900 |
| Antimony | 5.2(BG) |  | UL | UL | UL | UL |
| Arsenic | 20(SV) |  | 26.7 | 9.1 | 12.6 | 10.6 |
| Barium | 1,500 |  | 78.0 | 54.0 | 85.3 | 76.3 |
| Beryllium | 16 |  | 0.97 | 1.0 | 0.96 | 1.7 |
| Cadmium | 7 |  | 0.80 | 0.90 | 0.82 | 1.1 |
| Calcium | 4207 |  | 4,120 | [636] | 4,690 | 961 |
| Chromium | 12,000 |  | 112 | 226 | 56.8 | 54.1 |
| Cobalt | 17.8 |  | 16.7 | 30.6 | 19.5 | 26.2 |
| Copper | 310 |  | 62.4 | 37.6 | 53.3 | 67.3 |
| Hexavalent Chromium | 23 |  | NS | NS | ND | NS |
| Iron | 32,400 | 32,400 | 34,400 | 42,400 | 29,300 | 45,100 |
| Lead | 400 |  | 17.9 | 19.4 | 28.3 | 21.1 |
| Magnesium | 6950 | 6,950 | 9,390 | 5,380 | 6,900 | 7,140 |
| Manganese | 1800.00 | 968 | 437 | 441 | 681 | 668 |
| Mercury | 0.78 |  | 0.32 | 0.19 B | [0.13] B | [0.10] B |
| Nickel | 160 |  | 45.8 | 35.4 | 30.7 | 32.9 |
| Potassium | 4945 |  | 3,540 J | 2,940 J | 3,320 J | 3,850 J |
| Selenium | 39 |  | R | R | [0.59] L | R |
| Silver | 39 |  | ND | ND | ND | ND |
| Sodium | 55.8 |  | [74.8] | ND | [42.3] | [36.8] |
| Thallium | 2.2 | 2.20 | [0.26] L | [0.39] L | [0.34] L | [0.31] L |
| Vanadium | 390 | 75.5 | 74.7 | 76.9 | 67.1 | 121 |
| Zinc | 2,300 |  | 72.2 | 68.3 | 81.6 | 74.9 |
| Cyanide | 160 |  | ND | ND | ND | ND |
| ND = No analyte detected. |  |  |  |  |  |  |
| NS = Not sampled for this parameter. |  |  |  |  |  |  |
| $\mathrm{R}=$ Rejected. Unusable result. |  |  |  |  |  |  |
| $B$ = Not found substantially above level in field blank. |  |  |  |  |  |  |
| NA = Not applicable. |  |  |  |  |  |  |
| NL = Not Listed. |  |  |  |  |  |  |
| [ ] Present but estimated. "J" flag equivalent for metals. |  |  |  |  |  |  |
| BG - 95th percentile of 2007 Background Study. This is used when it is higher than the RSL. |  |  |  |  |  |  |
| SV - Spring Valley Remediation Goal. |  |  |  |  |  |  |
| Exceeds higher of Adjusted RSL or BG (2007 Study) |  |  |  |  |  |  |
| RSLs are EPA Regional Screening Levels (12 September 2008). Otherwise, original April 1999 values shown. |  |  |  |  |  |  |
|  |  |  |  |  |  |  |
| SOURCE: USEPA Interim Trip Report \#1, August 10, 1999. |  |  |  |  |  |  |

Table A.5--PARSONS 2000
4835 GLENBROOK ROAD
VALIDATED QUADRANT MUSTARD AGENT BREAKDOWN PRODUCT RESULTS

| Lot <br> Number | Sample Type | SampleID | Date <br> Collected | Depth Inches/feet | PARAMETER |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  |  |  |  |  | ABP (Dithiane) |  | ABP (Oxathiane) |  | ABP (Thiodiglycol) |  |
|  |  |  |  |  | ug/kg | Qualifier | ug/kg | Qualifier | ug/kg | Qualifier |
|  | Surface | OU3 MTL-4835-1 | 10/31/2000 | 0-6" | 132 | U | 264 | U | 934 | J |
|  |  | OU3 MTL-4835-2 | 10/31/2000 | 0-6" | 121 | U | 241 | U | 792 | J |
| 4835 |  | OU3 MTL-4835-3 | 10/31/2000 | 0-6" | 132 | U | 264 | U | 1190 | J |
| Glenbrook |  | OU3 MTL-4835-4 | 10/31/2000 | 0-6" | 145 | U | 289 | U | 905 | J |
|  | Subsurface | OU3 MTL-4835-SB-(0-2) | 10/31/2000 | 0-2' | 117 | U | 234 | U | 586 | U |
|  |  | OU3 MTL-4835-SB-(2-4) | 10/31/2000 | 2-4' | 115 | U | 230 | U | 575 | U |
|  |  | OU3 MTL-4835-SB-(4-6) | 10/31/2000 | 4-6' | 113 | U | 225 | U | 579 | U |

J - Result is estimated due to a minor QA/QC problem.
$U$ - Analyte not detected at the quantitation limit.

Table A.6--PARSONS 2000
4835 GLENBROOK ROAD
VALIDATED ARSENIC GRID RESULTS

| SAMPLE ID | SAMPLE DEPTH (FT) | DATE | ARSENIC (mg/kg) | DVQUAL |
| :---: | :---: | :---: | :---: | :---: |
| OU3-MTL-4835(-100,0) | 0.5 | 10/30/2000 | 6.6 | DJ |
| OU3-MTL-4835(-100,100) | 0.5 | 10/31/2000 | 4.66 | DJ |
| OU3-MTL-4835(-100,120) | 0.5 | 10/31/2000 | 5.43 | DJ |
| OU3-MTL-4835(-100,140) | 0.5 | 10/31/2000 | 14.9 | D |
| OU3-MTL-4835(-100,20) | 0.5 | 10/31/2000 | 18.4 | DJ |
| OU3-MTL-4835(-100,40) | 0.5 | 10/31/2000 | 36.4 | DJ |
| OU3-MTL-4835(-100,60) | 0.5 | 10/31/2000 | 27.4 | DJ |
| OU3-MTL-4835(-100,80) | 0.5 | 10/31/2000 | 9.23 | DJ |
| OU3-MTL-4835(-120,100) | 0.5 | 10/31/2000 | 6.01 | D |
| OU3-MTL-4835(-120,120) | 0.5 | 11/1/2000 | 5.26 | D |
| OU3-MTL-4835(-120,140) | 0.5 | 11/2/2000 | 7.15 | D |
| OU3-MTL-4835(-140,100) | 0.5 | 11/3/2000 | 5.96 | D |
| OU3-MTL-4835(-140,120) | 0.5 | 11/4/2000 | 5.51 | D |
| OU3-MTL-4835(-140,140) | 0.5 | 11/5/2000 | 6.46 | D |
| OU3-MTL-4835(-160,100) | 0.5 | 11/6/2000 | 8.19 | D |
| OU3-MTL-4835(-160,120) | 0.5 | 11/7/2000 | 8.71 | D |
| OU3-MTL-4835(-160,140) | 0.5 | 11/8/2000 | 8.3 | D |
| OU3-MTL-4835(-180,120) | 0.5 | 11/9/2000 | 15.5 | DJ |
| OU3-MTL-4835(-180,140) | 0.5 | 11/10/2000 | 11.6 | D |
| OU3-MTL-4835(-200,100) | 0.5 | 11/11/2000 | 21.7 | DJ |
| OU3-MTL-4835(-200,120) | 0.5 | 11/12/2000 | 17.9 | DJ |
| OU3-MTL-4835(-200,140) | 0.5 | 11/13/2000 | 4.73 | DJ |
| OU3-MTL-4835(-220,120) | 0.5 | 11/14/2000 | 3.95 | DJ |
| OU3-MTL-4835(-220,140) | 0.5 | 11/15/2000 | 4.97 | DJ |
| OU3-MTL-4835(-240,120) | 0.5 | 11/16/2000 | 2.71 | DJ |
| OU3-MTL-4835(-240,140) | 0.5 | 11/17/2000 | 3.13 | DJ |
| OU3-MTL-4835(-260,120) | 0.5 | 11/18/2000 | 3.4 | DJ |
| OU3-MTL-4835(-260,140) | 0.5 | 11/19/2000 | 1.93 | DJ |
| OU3-MTL-4835(-280,120) | 0.5 | 11/20/2000 | 7.23 | DJ |
| OU3-MTL-4835(-320,0) | 0.5 | 11/21/2000 | 2.53 | DJ |
| OU3-MTL-4835(-340,0) | 0.5 | 11/22/2000 | 5.78 | D |
| OU3-MTL-4835(280,140) | 0.5 | 11/23/2000 | 3.27 | D |
| OU3-MTL-4835-(-100,-20) | 0.5 | 11/24/2000 | 7.55 | D |
| OU3-MTL-4835-(-100,-40) | 0.5 | 11/25/2000 | 13.9 | D |
| OU3-MTL-4835-(-120,-20) | 0.5 | 11/26/2000 | 14.5 | D |
| OU3-MTL-4835-(-120,-40) | 0.5 | 11/27/2000 | 11.5 | D |
| OU3-MTL-4835-(-120,0) | 0.5 | 11/28/2000 | 14.1 | D |
| OU3-MTL-4835-(-140,-20) | 0.5 | 11/29/2000 | 40.7 | D |
| OU3-MTL-4835-(-140,-40) | 0.5 | 11/30/2000 | 14.9 | D |
| OU3-MTL-4835-(-140,0) | 0.5 | 12/1/2000 | 12.4 | D |
| OU3-MTL-4835-(-160,60) | 0.5 | 12/2/2000 | 28.0 | D |
| OU3-MTL-4835-(-160,80) | 0.5 | 12/3/2000 | 14.9 | D |

Table A.6--PARSONS 2000
4835 GLENBROOK ROAD
VALIDATED ARSENIC GRID RESULTS

| SAMPLE ID | SAMPLE DEPTH (FT) | DATE | ARSENIC (mg/kg) | DVQUAL |
| :---: | :---: | :---: | :---: | :---: |
| OU3-MTL-4835-(-180,100) | 0.5 | 12/4/2000 | 12.1 | DJ |
| OU3-MTL-4835-(-180,20) | 0.5 | 12/5/2000 | 52.9 | DJ |
| OU3-MTL-4835-(-180,60) | 0.5 | 12/6/2000 | 16.5 | D |
| OU3-MTL-4835-(-180,80) | 0.5 | 12/7/2000 | 15.1 | D |
| OU3-MTL-4835-(-200,20) | 0.5 | 12/8/2000 | 20.6 | DJ |
| OU3-MTL-4835-(-200,60) | 0.5 | 12/9/2000 | 16.9 | D |
| OU3-MTL-4835-(-200,80) | 0.5 | 12/10/2000 | 13.9 | D |
| OU3-MTL-4835-(-220,100) | 0.5 | 12/11/2000 | 13.6 | DJ |
| OU3-MTL-4835-(-220,40) | 0.5 | 12/12/2000 | 8.41 | D |
| OU3-MTL-4835-(-220,60) | 0.5 | 12/13/2000 | 12.9 | DJ |
| OU3-MTL-4835-(-220,80) | 0.5 | 12/14/2000 | 12.5 | D |
| OU3-MTL-4835-(-240,100) | 0.5 | 12/15/2000 | 2.2 | DJ |
| OU3-MTL-4835-(-240,60) | 0.5 | 12/16/2000 | 9.43 | D |
| OU3-MTL-4835-(-240,80) | 0.5 | 12/17/2000 | 9.4 | D |
| OU3-MTL-4835-(-260,100) | 0.5 | 12/18/2000 | 9.9 | DJ |
| OU3-MTL-4835-(-260,80) | 0.5 | 12/19/2000 | 39.7 | D |
| OU3-MTL-4835-(-280,100) | 0.5 | 12/20/2000 | 4.5 | DJ |
| OU3-MTL-4835-(-300,0) | 0.5 | 12/21/2000 | 7.28 | D |
|  |  |  |  |  |
| J - Result is estimated due to a minor QA/QC problem. |  |  |  |  |
| D - Sample was diluted due to matrix interferences. |  |  |  |  |
| JD - Value reported is from a diluted sample and is estimated due to minor QA/QC problem. |  |  |  |  |
| Sample has been removed |  |  |  |  |
|  |  |  |  |  |

Table A.7--PARSONS 2007-2008
ANALYTICAL RESULTS FOR 4835 TEST PIT SAMPLES
(12 Metals-Suite)

| SAMPLE ID: |  |  | $\begin{aligned} & \text { SW-4835GB- } \\ & (-170,10)-2 \end{aligned}$ |  | SW-4835GB-$(-170,10)-3$ |  | $\begin{aligned} & \text { SW-4835GB- } \\ & (-170,10) \text { SW-E } \end{aligned}$ |  | $\begin{aligned} & \text { SW-4835GB- } \\ & (-170,10) \text { SW-E(5) } \end{aligned}$ |  | $\begin{aligned} & \text { SW-4835GB- } \\ & (-170,10) \text { SW-S } \end{aligned}$ |  | $\begin{aligned} & \text { SW-4835GB- } \\ & (-170,10) \text { SW-W } \end{aligned}$ |  | SW-4835GB-$(-190,10)-2$ |  | SW-4835GB-$(-190,10)-\mathrm{N}$ |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| DATE SAMPLED: |  |  | 08/23/07 |  | 08/23/07 |  | 08/23/07 |  | 08/23/07 |  | 08/23/07 |  | 08/23/07 |  | 08/23/07 |  | 08/23/07 |  |
| LAB SAMPLE ID: |  |  | 708188-001 |  | 708188-002 |  | 708188-005 |  | 708188-006 |  | 708188-003 |  | 708188-007 |  | 708188-015 |  | 708188-017 |  |
| SAMPLE DEPTH(FT) |  |  | 2-2.5 |  | 3-3.5 |  | 0.5 |  | 0.5 |  | 0.5 |  | 0.5 |  | 2-2.5 |  | 0.5 |  |
|  | Units | Regional Screening Level |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| Total Metals - ILM05.4 |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| Aluminum | $\mathrm{mg} / \mathrm{kg}$ | 19,100 | 17900 |  | 21000 |  | 12500 |  | 8960 |  | 12500 |  | 15100 |  | 15600 |  | 21000 |  |
| Antimony | $\mathrm{mg} / \mathrm{kg}$ | 5.2(BG) | 1.2 | J | 1 | J | 0.8 | J | 0.51 | J | 4.5 | UL | 0.4 | J | 0.78 | J | 0.75 | J |
| Arsenic | $\mathrm{mg} / \mathrm{kg}$ | 20(SV) | 80.2 |  | 77 | J | 43.4 |  | 8 | J | 1.9 | K | 7.8 | K | 8.3 | K | 11.4 | K |
| Barium | $\mathrm{mg} / \mathrm{kg}$ | 1,500 | 63.6 |  | 45.7 |  | 17.9 |  | 18.2 |  | 26.3 |  | 41.8 |  | 71.8 |  | 54.8 |  |
| Cadmium | $\mathrm{mg} / \mathrm{kg}$ | 7.0 | 0.19 | J | 0.45 | U | 0.45 | U | 0.4 | U | 0.38 | U | 0.41 | U | 0.49 | U | 0.43 | U |
| Copper | $\mathrm{mg} / \mathrm{kg}$ | 310 | 41.2 | J | 49.8 |  | 116 | J | 85.1 |  | 59.7 | J | 78.2 | J | 38.4 | J | 42 | J |
| Lead | $\mathrm{mg} / \mathrm{kg}$ | 400 | 38 |  | 22.8 | J | 3.5 |  | 15.2 | J | 4.8 |  | 9.1 |  | 36.6 |  | 27 |  |
| Manganese | $\mathrm{mg} / \mathrm{kg}$ | 1,800 | 503 | J | 272 |  | 505 | J | 391 |  | 227 | J | 344 | J | 790 | J | 398 | J |
| Mercury | $\mathrm{mg} / \mathrm{kg}$ | 0.78* | 0.85 |  | 0.11 | J | 0.058 | J | 0.17 | J | 0.077 | U | 0.089 | U | 0.17 |  | 0.11 | U |
| Nickel | mg/kg | 160 | 50.1 |  | 61.8 |  | 647 | + | 345 |  | 18.8 |  | 25.7 |  | 72.8 |  | 82.4 |  |
| Thallium | mg/kg | 2.20 | 1 | J | 2.3 | U | 2.2 | U | 2 | U | 1.9 | U | 2 | U | 1.3 | J | 0.98 | J |
| Vanadium | mg/kg | 390.00 | 86.6 |  | 80.6 |  | 82.3 |  | 130 |  | 62.9 |  | 102 |  | 96.1 |  | 83.3 |  |
| Zinc | $\mathrm{mg} / \mathrm{kg}$ | 2,300 | 244 |  | 101 |  | 39.9 |  | 31.9 |  | 34.9 |  | 57.1 |  | 63.4 |  | 61.8 |  |
|  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| Comparison value based on the higher of the adjusted Sept 2008 Regional Screening Level (RSL) (if non-carcinogenic) or the 2007 Background value. |  |  |  |  | U - Analyte was analyzed for but not detected above the adjusted practical quantitation limit (PQL). |  |  |  |  |  |  |  |  |  |  |  |  |  |
| NA - Not available. |  |  |  |  | UJ - Analyte not detected, reported PQL may be inaccurate or imprecise. |  |  |  |  |  |  |  |  |  |  |  |  |  |
| BG - Background Value (2007 Study). |  |  |  |  | UL - Analyte not detected, reported PQL is biased low, actual PQL is expected to be higher. |  |  |  |  |  |  |  |  |  |  |  |  |  |
| SV - Spring Valley Remediation Goal. |  |  |  |  | $J$ - Analyte detected, estimated concentration. |  |  |  |  |  |  |  |  |  |  |  |  |  |
| (NO CODE) - Confirmed identification. |  |  |  |  | L - Analyte detected, reported result is biased low, actual value is expected to be higher. |  |  |  |  |  |  |  |  |  |  |  |  |  |
| + - Result reported from diluted sample. |  |  |  |  | K - Analyte detected, reported result is biased high, actual value is expected to be lower. |  |  |  |  |  |  |  |  |  |  |  |  |  |
| Detections are bolded. |  |  |  |  | Detections exceeding the comparison level are shown shaded and bolded. |  |  |  |  |  |  |  |  |  |  |  |  |  |
| Sample has been removed |  |  |  |  | * RSL for methyl mercury since methyl mercury is on the AUES list. |  |  |  |  |  |  |  |  |  |  |  |  |  |
|  |  |  |  |  | Sample has been removed |  |  |  |  |  |  |  |  |  |  |  |  |  |
|  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
|  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |

Table A.7--PARSONS 2007-2008
ANALYTICAL RESULTS FOR 4835 TEST PIT SAMPLES
(12 Metals-Suite)

| SAMPLE ID: |  |  | $\begin{aligned} & \text { SW-4835GB- } \\ & (-130,-30)-1.5 \end{aligned}$ |  | $\begin{aligned} & \text { SW-4835GB- } \\ & (-130,-30) S W-N \end{aligned}$ |  | $\begin{aligned} & \text { SW-4835GB- } \\ & (-130,-30) \text { SW-S } \end{aligned}$ |  | $\begin{aligned} & \text { SW-4835GB- } \\ & (-130,-30) \text { SW-W } \end{aligned}$ |  | SW-4835GB-$(-190,90)-2$ |  | SW-4835GB- <br> (-190,90)SW-E |  | $\begin{aligned} & \text { SW-4835GB- } \\ & (-190,90) \text { SW-E(5) } \end{aligned}$ |  | $\begin{aligned} & \text { SW-4835GB- } \\ & (-190,90) \text { SW-N } \end{aligned}$ |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| DATE SAMPLED: |  |  | 08/23/07 |  | 08/23/07 |  | 08/23/07 |  | 08/23/07 |  | 08/27/07 |  | 08/27/07 |  | 08/27/07 |  | 08/27/07 |  |
| LAB SAMPLE ID: |  |  | 708188-009 |  | 708188-010 |  | 708188-012 |  | 708188-013 |  | 708203-006 |  | 708203-012 |  | 708203-013 |  | 708203-008 |  |
| SAMPLE DEPTH(FT) |  |  | 1.5 |  | 0.5 |  | 0.5 |  | 0.5 |  | 2-2.5 |  | 0.5 |  | 0.5 |  | 0.5 |  |
|  | Units | Regional Screening Level |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| Total Metals - ILM05.4 |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| Aluminum | mg/kg | 19,100 | 19800 |  | 10700 |  | 24100 |  | 13900 |  | 23100 | + | 20600 |  | 24800 | + | 19500 |  |
| Antimony | mg/kg | 5.2(BG) | 6.2 | UL | 0.34 | J | 0.63 | J | 0.42 | J | 0.63 | J+ | 0.41 | J | 1.6 | J+ | 1 | J |
| Arsenic | mg/kg | 20(SV) | 18.3 | K | 5.1 | K | 20.3 | K | 2.4 | K | 8.8 | L+ | 23.9 | L | 16.1 | J | 22.8 | L+ |
| Barium | mg/kg | 1,500 | 90.3 |  | 44 |  | 90.1 |  | 23.3 |  | 115 | + | 86.8 |  | 113 | + | 84.5 |  |
| Cadmium | mg/kg | 7.0 | 0.037 | J | 0.41 | U | 0.48 | U | 0.12 | J | 0.85 | U+ | 0.44 | U | 0.85 | U+ | 0.85 | U+ |
| Copper | mg/kg | 310 | 52.8 | J | 58 | J | 75 | J | 102 | J | 91.7 | + | 60.7 |  | 88.5 | + | 65.3 |  |
| Lead | mg/kg | 400 | 12.3 |  | 7.4 |  | 12.7 |  | 13.3 |  | 27.1 | + | 20.2 |  | 22.4 | J+ | 18.4 |  |
| Manganese | mg/kg | 1,800 | 436 | J | 302 | J | 441 | J | 385 | J | 954 | + | 858 |  | 781 | + | 610 |  |
| Mercury | mg/kg | 0.78* | 0.033 | J | 0.076 | U | 0.046 | J | 0.096 | U | 0.06 | J | 0.099 |  | 0.17 | J | 0.19 |  |
| Nickel | mg/kg | 160 | 32 |  | 24.9 |  | 45.9 |  | 25.8 |  | 64.3 |  | 49.3 |  | 49.2 |  | 42.8 |  |
| Thallium | mg/kg | 2.20 | 2.6 | U | 2 | U | 1.1 | J | 0.56 | J | 4.3 | U+ | 1.4 | J | 4.3 | U+ | 0.96 | J |
| Vanadium | mg/kg | 390.00 | 61.9 |  | 98.5 |  | 79.4 |  | 131 |  | 119 | + | 73.6 |  | 107 | + | 72.9 |  |
| Zinc | mg/kg | 2,300 | 63.8 |  | 47.7 |  | 73 |  | 77.4 |  | 109 | + | 88.3 |  | 115 | + | 84.2 |  |
|  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| Comparison value based on the higher of the adjusted Sept 2008 Regional Screening Level (RSL) (if non-carcinogenic) or the 2007 Background value. |  |  |  |  | U - Analyte was analyzed for but not detected above the adjusted practical quantitation limit (PQL). |  |  |  |  |  |  |  |  |  |  |  |  |  |
| NA - Not available. |  |  |  |  | UJ - Analyte not detected, reported PQL may be inaccurate or imprecise. |  |  |  |  |  |  |  |  |  |  |  |  |  |
| BG - Background Value (2007 Study). |  |  |  |  | UL - Analyte not detected, reported PQL is biased low, actual PQL is expected to be higher. |  |  |  |  |  |  |  |  |  |  |  |  |  |
| SV - Spring Valley Remediation Goal. |  |  |  |  | J - Analyte detected, estimated concentration. |  |  |  |  |  |  |  |  |  |  |  |  |  |
| (NO CODE) - Confirmed identification. |  |  |  |  | L - Analyte detected, reported result is biased low, actual value is expected to be higher. |  |  |  |  |  |  |  |  |  |  |  |  |  |
| +- Result reported from diluted sample. |  |  |  |  | K - Analyte detected, reported result is biased high, actual value is expected to be lower. |  |  |  |  |  |  |  |  |  |  |  |  |  |
| Detections are bolded. |  |  |  |  | Detections exceeding the comparison level are shown shaded and bolded. |  |  |  |  |  |  |  |  |  |  |  |  |  |
| Sample has been removed |  |  |  |  | * RSL for methyl mercury since methyl mercury is on the AUES list. |  |  |  |  |  |  |  |  |  |  |  |  |  |
|  |  |  |  |  | Sample has been removed |  |  |  |  |  |  |  |  |  |  |  |  |  |
|  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
|  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |

Table A.7--PARSONS 2007-2008
ANALYTICAL RESULTS FOR 4835 TEST PIT SAMPLES
(12 Metals-Suite)

| SAMPLE ID: |  |  | SW-4835GB-$(-190,90) S W-N(5)$ |  | SW-4835GB-(-190,90)SW-S |  | SW-4835GB-$(-250,70)-2$ |  | SW-4835GB-(-250,70)SW-E |  | $\begin{aligned} & \text { SW-4835GB- } \\ & (-250,70) \text { SW-S } \end{aligned}$ |  | SW-4835GB-$(-150,50)-2$ |  | SW-4835GB( $-150,50$ )SW-E |  | $\begin{aligned} & \text { SW-4835GB- } \\ & (-150,50) \text { SW-N } \end{aligned}$ |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| DATE SAMPLED: |  |  | 08/27/07 |  | 08/27/07 |  | 08/27/07 |  | 08/27/07 |  | 08/27/07 |  | 08/27/07 |  | 08/27/07 |  | 08/27/07 |  |
| LAB SAMPLE ID: |  |  | 708203-009 |  | 708203-010 |  | 708203-001 |  | 708203-004 |  | 708203-003 |  | 708204-001 |  | 708204-005 |  | 708204-003 |  |
| SAMPLE DEPTH(FT) |  |  | 0.5 |  | 0.5 |  | 2-2.5 |  | 0.5 |  | 0.5 |  | 2-2.5 |  | 0.5 |  | 0.5 |  |
|  | Units | Regional Screening Level |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| Total Metals - ILM05.4 |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| Aluminum | mg/kg | 19,100 | 23200 |  | 17700 |  | 15000 |  | 19000 |  | 14700 |  | 25800 |  | 16700 |  | 14500 |  |
| Antimony | mg/kg | 5.2(BG) | 0.56 | J | 0.86 | J | 5.4 | UL | 0.52 | J | 0.53 | J | 0.43 | J | 0.83 | J | 0.58 | J |
| Arsenic | mg/kg | 20(SV) | 24.8 | J | 10.8 | L | 3 | L | 11.9 | L+ | 8 | L | 16.7 | J+ | 8.2 | J+ | 9.3 | J |
| Barium | mg/kg | 1,500 | 100 |  | 106 |  | 50.9 |  | 73.9 |  | 87.1 |  | 108 |  | 62 |  | 94.6 |  |
| Cadmium | mg/kg | 7.0 | 0.45 | U | 0.39 | U | 2.3 | U+ | 0.92 | U+ | 0.46 | U | 0.92 | U+ | 0.86 | U+ | 0.46 | U |
| Copper | mg/kg | 310 | 70.5 |  | 53.6 |  | 22.6 |  | 49.4 |  | 38.4 |  | 63.3 | J | 52.4 | J | 32.7 | J |
| Lead | mg/kg | 400 | 25.5 | J | 53 |  | 10 |  | 19.6 |  | 19.7 |  | 9.3 |  | 20.5 |  | 18.8 |  |
| Manganese | mg/kg | 1,800 | 651 |  | 596 |  | 651 |  | 497 |  | 626 |  | 455 |  | 451 |  | 931 |  |
| Mercury | mg/kg | 0.78* | 0.2 | J | 0.09 | J | 0.013 | J | 0.14 |  | 0.096 | J | 0.03 | J | 0.084 | J | 0.018 | J |
| Nickel | mg/kg | 160 | 59.7 |  | 52.2 |  | 47.6 |  | 44.9 |  | 54.3 |  | 45.3 |  | 43.3 |  | 38.3 |  |
| Thallium | mg/kg | 2.20 | 0.95 | J | 0.66 | J | 1.2 | J | 0.91 | J | 2.3 | U | 0.93 | J | 0.91 | J | 0.83 | J |
| Vanadium | $\mathrm{mg} / \mathrm{kg}$ | 390.00 | 80.6 |  | 74.8 |  | 75.9 |  | 71.5 |  | 56.7 |  | 82 |  | 80.1 |  | 48.8 |  |
| Zinc | $\mathrm{mg} / \mathrm{kg}$ | 2,300 | 80.3 |  | 119 |  | 44.1 |  | 79.5 |  | 66.6 |  | 69.8 |  | 70.2 |  | 84.1 |  |
|  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| Comparison value based on the higher of the adjusted Sept 2008 Regional Screening Level (RSL) (if non-carcinogenic) or the 2007 Background value. |  |  |  |  | U - Analyte was analyzed for but not detected above the adjusted practical quantitation limit (PQL). |  |  |  |  |  |  |  |  |  |  |  |  |  |
| NA - Not available. |  |  |  |  | UJ - Analyte not detected, reported PQL may be inaccurate or imprecise. |  |  |  |  |  |  |  |  |  |  |  |  |  |
| BG - Background Value (2007 Study). |  |  |  |  | UL - Analyte not detected, reported PQL is biased low, actual PQL is expected to be higher. |  |  |  |  |  |  |  |  |  |  |  |  |  |
| SV - Spring Valley Remediation Goal. |  |  |  |  | $J$ - Analyte detected, estimated concentration. |  |  |  |  |  |  |  |  |  |  |  |  |  |
| (NO CODE) - Confirmed identification. |  |  |  |  | L - Analyte detected, reported result is biased low, actual value is expected to be higher. |  |  |  |  |  |  |  |  |  |  |  |  |  |
| + - Result reported from diluted sample. |  |  |  |  | K - Analyte detected, reported result is biased high, actual value is expected to be lower. |  |  |  |  |  |  |  |  |  |  |  |  |  |
| Detections are bolded. |  |  |  |  | Detections exceeding the comparison level are shown shaded and bolded. |  |  |  |  |  |  |  |  |  |  |  |  |  |
| Sample has been removed |  |  |  |  | * RSL for methyl mercury since methyl mercury is on the AUES list. |  |  |  |  |  |  |  |  |  |  |  |  |  |
|  |  |  |  |  | Sample has been removed |  |  |  |  |  |  |  |  |  |  |  |  |  |
|  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
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Table A.7--PARSONS 2007-2008
ANALYTICAL RESULTS FOR 4835 TEST PIT SAMPLES
(12 Metals-Suite)

| SAMPLE ID: |  |  | $\begin{gathered} \text { SW-4835GB- } \\ (-90,50)-2 \end{gathered}$ |  | SW-4835GB- <br> (-90,50)SW-N |  | SW-4835GB-$(-90,50) S W-E$ |  | $\begin{aligned} & \text { SW-4835GB- } \\ & (-190,10) \text { SW-E } \end{aligned}$ |  | $\begin{aligned} & \text { SW-4835GB- } \\ & (-190,10) S W-E(5) \end{aligned}$ |  | $\begin{gathered} \text { SW-4835GB- } \\ (-90,30)-2 \end{gathered}$ |  | $\begin{gathered} \text { SW-4835GB- } \\ (-90,30)-3 \end{gathered}$ |  | SW-4835GB- <br> (-90,30)SW-W |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| DATE SAMPLED: |  |  | 08/27/07 |  | 08/27/07 |  | 08/27/07 |  | 08/27/07 |  | 08/27/07 |  | 08/27/07 |  | 08/27/07 |  | 08/27/07 |  |
| LAB SAMPLE ID: |  |  | 708204-007 |  | 708204-009 |  | 708204-010 |  | 708205-001 |  | 708205-002 |  | 708205-003 |  | 708205-004 |  | 708205-005 |  |
| SAMPLE DEPTH(FT) |  |  | 2-2.5 |  | 0.5 |  | 0.5 |  | 0.5 |  | 0.5 |  | 2-2.5 |  | 3-3.5 |  | 0.5 |  |
|  | Units | Regional Screening Level |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| Total Metals - ILM05.4 |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| Aluminum | $\mathrm{mg} / \mathrm{kg}$ | 19,100 | 25400 |  | 31700 | + | 26200 |  | 21700 | + | 18500 |  | 26200 |  | 30300 |  | 29600 |  |
| Antimony | mg/kg | 5.2(BG) | 0.46 | J | 10.2 | UL+ | 5.7 | UL | 10.3 | UL+ | 0.76 | J | 0.25 | J | 0.55 | J | 0.48 | J |
| Arsenic | $\mathrm{mg} / \mathrm{kg}$ | 20(SV) | 17.6 | J+ | 24.2 | J+ | 11.7 | J | 23.4 | J | 54 | J | 20.2 | J | 22.6 | J | 23.8 | J |
| Barium | $\mathrm{mg} / \mathrm{kg}$ | 1,500 | 101 |  | 123 | + | 126 |  | 83.3 | + | 78.1 |  | 102 |  | 113 |  | 114 |  |
| Cadmium | $\mathrm{mg} / \mathrm{kg}$ | 7.0 | 0.78 | U+ | 0.85 | U+ | 0.48 | U | 0.86 | U+ | 0.12 | J | 0.41 | U | 0.41 | U | 0.44 | U |
| Copper | mg/kg | 310 | 65.1 | J | 140 | J+ | 72.8 | J | 51.2 | + | 46.9 |  | 69.2 |  | 65.3 |  | 118 |  |
| Lead | $\mathrm{mg} / \mathrm{kg}$ | 400 | 10.3 |  | 12.6 | + | 16.1 |  | 20.4 | + | 17.9 |  | 10.5 |  | 8.8 |  | 10 |  |
| Manganese | $\mathrm{mg} / \mathrm{kg}$ | 1,800 | 489 |  | 645 | + | 695 |  | 611 | + | 525 |  | 501 |  | 429 |  | 612 |  |
| Mercury | $\mathrm{mg} / \mathrm{kg}$ | 0.78* | 0.2 |  | 0.15 |  | 0.077 | J | 0.04 | J | 0.27 | J | 0.014 | J | 0.32 | J | 0.15 |  |
| Nickel | $\mathrm{mg} / \mathrm{kg}$ | 160 | 41.5 |  | 88.7 |  | 80.5 |  | 69.6 | L | 62 |  | 42.9 | L | 45.1 |  | 70.1 | L |
| Thallium | $\mathrm{mg} / \mathrm{kg}$ | 2.20 | 1.6 | J | 4.3 | U+ | 1.4 | J | 1.2 | J+ | 0.56 | U | 0.7 | J | 0.81 | J | 2.2 | U |
| Vanadium | $\mathrm{mg} / \mathrm{kg}$ | 390.00 | 80.5 |  | 137 | + | 82.4 |  | 77.8 | J+ | 74.9 |  | 86.1 | J | 94.4 |  | 115 | J |
| Zinc | $\mathrm{mg} / \mathrm{kg}$ | 2,300 | 73 |  | 85.9 | + | 73.4 |  | 71 | + | 71.9 |  | 72.6 |  | 70.4 |  | 89.3 |  |
|  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| Comparison value based on the higher of the adjusted Sept 2008 Regional Screening Level (RSL) (if non-carcinogenic) or the 2007 Background value. |  |  |  |  | U - Analyte was analyzed for but not detected above the adjusted practical quantitation limit (PQL). |  |  |  |  |  |  |  |  |  |  |  |  |  |
| NA - Not available. |  |  |  |  | UJ - Analyte not detected, reported PQL may be inaccurate or imprecise. |  |  |  |  |  |  |  |  |  |  |  |  |  |
| BG - Background Value (2007 Study). |  |  |  |  | UL - Analyte not detected, reported PQL is biased low, actual PQL is expected to be higher. |  |  |  |  |  |  |  |  |  |  |  |  |  |
| SV - Spring Valley Remediation Goal. |  |  |  |  | $J$ - Analyte detected, estimated concentration. |  |  |  |  |  |  |  |  |  |  |  |  |  |
| (NO CODE) - Confirmed identification. |  |  |  |  | L - Analyte detected, reported result is biased low, actual value is expected to be higher. |  |  |  |  |  |  |  |  |  |  |  |  |  |
| +- Result reported from diluted sample. |  |  |  |  | K - Analyte detected, reported result is biased high, actual value is expected to be lower. |  |  |  |  |  |  |  |  |  |  |  |  |  |
| Detections are bolded. |  |  |  |  | Detections exceeding the comparison level are shown shaded and bolded. |  |  |  |  |  |  |  |  |  |  |  |  |  |
| Sample has been removed |  |  |  |  | * RSL for methyl mercury since methyl mercury is on the AUES list. |  |  |  |  |  |  |  |  |  |  |  |  |  |
|  |  |  |  |  | Sample has been removed |  |  |  |  |  |  |  |  |  |  |  |  |  |
|  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
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Table A.7--PARSONS 2007-2008
ANALYTICAL RESULTS FOR 4835 TEST PIT SAMPLES
(12 Metals-Suite)

| SAMPLE ID: |  |  | $\begin{aligned} & \text { SW-4835GB- } \\ & (-90,30) \text { SW-W(5) } \end{aligned}$ |  | $\begin{gathered} \text { SW-4835GB- } \\ (-130,-30) \text { SW-S(2.5) } \end{gathered}$ |  | SW-4835GB- <br> (190,90)SW-E(5)LC |  | SW-4835GB- <br> $(-190,90)$ SW-E(5)LN |  | SW-4835GB- <br> (190,90)SW-E(5)LS |  | $\begin{gathered} \text { SW-4835GB- } \\ (-190,90) \text { SW-N(6) } \end{gathered}$ |  | SW-4835GB-(-90,50)SW-N(5) |  | SW-4835GB-$(-170,10)-4$ |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| DATE SAMPLED: |  |  | 08/27/07 |  | 10/03/07 |  | 10/03/07 |  | 10/03/07 |  | 10/03/07 |  | 10/03/07 |  | 10/04/07 |  | 10/09/07 |  |
| LAB SAMPLE ID: |  |  | 708205-006 |  | 710027-009 |  | 710027-008 |  | 710027-006 |  | 710027-007 |  | 710027-001 |  | 710033-003 |  | 710060-018 |  |
| SAMPLE DEPTH(FT) |  |  | 0.5 |  | 0.5 |  | 2 |  | 0.5 |  | 0.5 |  | 0.5 |  | 0.5 |  | 4-4.5 |  |
|  | Units | Regional Screening Level |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| Total Metals - ILM05.4 |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| Aluminum | mg/kg | 19,100 | 31400 |  | 15500 |  | 21900 |  | 22100 |  | 21100 |  | 21400 |  | 29100 |  | 25900 | + |
| Antimony | mg/kg | 5.2(BG) | 0.8 | J | 8.5 | UL | 0.74 | J | 0.87 | J | 2.1 | J | 0.96 | J | 0.42 | J | 31.5 | UL+ |
| Arsenic | $\mathrm{mg} / \mathrm{kg}$ | 20(SV) | 18 | J | 8.3 |  | 19.6 |  | 10.7 |  | 9.9 |  | 11.4 |  | 14.4 | K | 6.1 | J+ |
| Barium | $\mathrm{mg} / \mathrm{kg}$ | 1,500 | 128 |  | 128 |  | 84.8 |  | 110 |  | 89.7 |  | 108 |  | 96.3 |  | 39.1 | J+ |
| Cadmium | $\mathrm{mg} / \mathrm{kg}$ | 7.0 | 0.11 | J | 0.12 | J | 0.51 | U | 0.51 | U | 0.49 | U | 0.49 | U | 0.46 | U | 2.6 | U+ |
| Copper | mg/kg | 310 | 105 |  | 51.6 |  | 66.1 |  | 84.1 |  | 61.8 |  | 71.0 |  | 107 |  | 62.5 | + |
| Lead | $\mathrm{mg} / \mathrm{kg}$ | 400 | 12.8 |  | 18.2 |  | 22.5 |  | 27.6 |  | 20.5 |  | 23.3 |  | 9.3 | J | 8.9 | U |
| Manganese | $\mathrm{mg} / \mathrm{kg}$ | 1,800 | 751 |  | 756 |  | 634 |  | 736 |  | 643 |  | 649 |  | 570 | J | 174 | + |
| Mercury | mg/kg | 0.78* | 0.12 | J | 0.098 | B | 0.25 | J | 0.15 | J | 0.12 | J | 0.083 | B | 0.088 | B | 0.058 | B |
| Nickel | mg/kg | 160 | 70.5 |  | 29.2 |  | 50.1 |  | 46.7 |  | 47.1 |  | 43.5 |  | 139 |  | 81.6 | + |
| Thallium | mg/kg | 2.20 | 0.6 | U | 3.6 | U | 2.6 | U | 1.1 | J | 1.0 | J | 1.0 | J | 1.3 | J | 13.1 | U+ |
| Vanadium | $\mathrm{mg} / \mathrm{kg}$ | 390.00 | 121 |  | 52.0 |  | 81.6 |  | 106 |  | 77.3 |  | 85.9 |  | 108 |  | 105 | + |
| Zinc | $\mathrm{mg} / \mathrm{kg}$ | 2,300 | 87.9 |  | 80.3 |  | 93.7 |  | 106 |  | 96.7 |  | 99.6 |  | 66.4 |  | 65.2 | + |
|  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| Comparison value based on the higher of the adjusted Sept 2008 Regional Screening Level (RSL) (if non-carcinogenic) or the 2007 Background value. |  |  |  |  | U - Analyte was analyzed for but not detected above the adjusted practical quantitation limit (PQL). |  |  |  |  |  |  |  |  |  |  |  |  |  |
| NA - Not available. |  |  |  |  | UJ - Analyte not detected, reported PQL may be inaccurate or imprecise. |  |  |  |  |  |  |  |  |  |  |  |  |  |
| BG - Background Value (2007 Study). |  |  |  |  | UL - Analyte not detected, reported PQL is biased low, actual PQL is expected to be higher. |  |  |  |  |  |  |  |  |  |  |  |  |  |
| SV - Spring Valley Remediation Goal. |  |  |  |  | J - Analyte detected, estimated concentration. |  |  |  |  |  |  |  |  |  |  |  |  |  |
| (NO CODE) - Confirmed identification. |  |  |  |  | L - Analyte detected, reported result is biased low, actual value is expected to be higher. |  |  |  |  |  |  |  |  |  |  |  |  |  |
| + - Result reported from diluted sample. |  |  |  |  | K - Analyte detected, reported result is biased high, actual value is expected to be lower. |  |  |  |  |  |  |  |  |  |  |  |  |  |
| Detections are bolded. |  |  |  |  | Detections exceeding the comparison level are shown shaded and bolded. |  |  |  |  |  |  |  |  |  |  |  |  |  |
| Sample has been removed |  |  |  |  | * RSL for methyl mercury since methyl mercury is on the AUES list. |  |  |  |  |  |  |  |  |  |  |  |  |  |
|  |  |  |  |  | Sample has been removed |  |  |  |  |  |  |  |  |  |  |  |  |  |
|  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
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Table A.7--PARSONS 2007-2008
ANALYTICAL RESULTS FOR 4835 TEST PIT SAMPLES
(12 Metals-Suite)

| SAMPLE ID: |  |  | SW-4835GB- <br> (-170,10)SW-E(5)LC-4 |  | SW-4835GB-(-170,10)SW-E(5)LS |  | $\begin{gathered} \text { SW-4835GB- } \\ (-170,10) \text { SW-S3.5 } \end{gathered}$ |  | $\begin{gathered} \text { SW-4835GB- } \\ (-170,10)-\text { SW-W3. } \end{gathered}$ |  | $\begin{aligned} & \text { SW-4835GB- } \\ & (-190,10) \text { SW-E(7) } \end{aligned}$ |  | SW-4835GB- <br> (-190,90)SW-N(6)LC |  | SW-4835GB-(-190,90)SW-N(6)LE |  | $\begin{aligned} & \text { SW-4835GB- } \\ & (-150,50) \text { SW-S(8) } \end{aligned}$ |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| DATE SAMPLED: |  |  | 10/09/07 |  | 10/09/07 |  | 10/09/07 |  | 10/09/07 |  | 10/09/07 |  | 11/21/07 |  | 11/21/07 |  | 12/11/07 |  |
| LAB SAMPLE ID: |  |  | 710060-008 |  | 710060-011 |  | 710060-015 |  | 710060-021 |  | 710060-004 |  | 711122-001 |  | 711122-002 |  | 712054-003 |  |
| SAMPLE DEPTH(FT) |  |  | 4-4.5 |  | 0.5 |  | 3.5 |  | 3.5 |  | 0.5 |  | 2-2.5 |  | 0.5 |  | 0.5 |  |
|  | Units | Regional Screening Level |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| Total Metals - ILM05.4 |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| Aluminum | $\mathrm{mg} / \mathrm{kg}$ | 19,100 | 45200 |  | 15900 |  | 27900 |  | 22200 |  | 10500 |  | 21000 |  | 29000 | + | 23300 | + |
| Antimony | $\mathrm{mg} / \mathrm{kg}$ | 5.2(BG) | 0.97 | B | 0.42 | B | 0.75 | B | 6.3 | UL | 1.0 | B | 1.8 | J | 0.57 | J+ | 10.8 | UL+ |
| Arsenic | $\mathrm{mg} / \mathrm{kg}$ | 20(SV) | 6.6 | J | 9.1 | J | 5.0 | J | 9.5 |  | 15.8 | J | 15.1 |  | 2.6 | J | 11.2 | J+ |
| Barium | $\mathrm{mg} / \mathrm{kg}$ | 1,500 | 142 |  | 31.3 |  | 43.9 |  | 81.1 |  | 60.3 |  | 84.6 |  | 71.3 | + | 140 | + |
| Cadmium | $\mathrm{mg} / \mathrm{kg}$ | 7.0 | 0.56 | U | 0.46 | U | 0.16 | J+ | 5.2 | U+ | 0.49 | U | 0.31 | J | 0.32 | J+ | 0.81 | J+ |
| Copper | mg/kg | 310 | 137 |  | 71.4 |  | 65.0 |  | 42.9 |  | 34.2 |  | 52.3 |  | 48.2 | + | 83.1 | + |
| Lead | mg/kg | 400 | 6.1 | K | 8.8 | K | 9.6 | K | 9.5 | J | 11.8 | K | 25.5 | K | 10.7 | K+ | 10.8 | + |
| Manganese | mg/kg | 1,800 | 982 |  | 199 |  | 193 |  | 823 |  | 383 |  | 844 | J | 590 | J+ | 560 | + |
| Mercury | mg/kg | 0.78* | 0.026 | B | 0.027 | B | 0.068 | B | 0.014 | B | 0.27 |  | 0.17 | J | 0.026 | B | 0.13 |  |
| Nickel | $\mathrm{mg} / \mathrm{kg}$ | 160 | 129 |  | 37.2 |  | 73.2 |  | 53.4 |  | 16.3 |  | 51.9 |  | 44.9 |  | 61.1 | + |
| Thallium | $\mathrm{mg} / \mathrm{kg}$ | 2.20 | 0.85 | J | 2.3 | U | 1.3 | J | 0.74 | J | 2.5 | U | 1.0 | J | 1.1 | J+ | 4.5 | U+ |
| Vanadium | $\mathrm{mg} / \mathrm{kg}$ | 390.00 | 105 |  | 99.3 |  | 114 |  | 83.6 |  | 38.4 |  | 77.1 |  | 69.1 | + | 96.9 | + |
| Zinc | $\mathrm{mg} / \mathrm{kg}$ | 2,300 | 89.6 |  | 44.1 |  | 62.8 |  | 66.6 |  | 63.4 |  | 83.3 |  | 81.1 | + | 62.6 | + |
|  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| Comparison value based on the higher of the adjusted Sept 2008 Regional Screening Level (RSL) (if non-carcinogenic) or the 2007 Background value. |  |  |  |  | U - Analyte was analyzed for but not detected above the adjusted practical quantitation limit (PQL). |  |  |  |  |  |  |  |  |  |  |  |  |  |
| NA - Not available. |  |  |  |  | UJ - Analyte not detected, reported PQL may be inaccurate or imprecise. |  |  |  |  |  |  |  |  |  |  |  |  |  |
| BG - Background Value (2007 Study). |  |  |  |  | UL - Analyte not detected, reported PQL is biased low, actual PQL is expected to be higher. |  |  |  |  |  |  |  |  |  |  |  |  |  |
| SV - Spring Valley Remediation Goal. |  |  |  |  | $J$ - Analyte detected, estimated concentration. |  |  |  |  |  |  |  |  |  |  |  |  |  |
| (NO CODE) - Confirmed identification. |  |  |  |  | L - Analyte detected, reported result is biased low, actual value is expected to be higher. |  |  |  |  |  |  |  |  |  |  |  |  |  |
| + - Result reported from diluted sample. |  |  |  |  | K - Analyte detected, reported result is biased high, actual value is expected to be lower. |  |  |  |  |  |  |  |  |  |  |  |  |  |
| Detections are bolded. |  |  |  |  | Detections exceeding the comparison level are shown shaded and bolded. |  |  |  |  |  |  |  |  |  |  |  |  |  |
| Sample has been removed |  |  |  |  | * RSL for methyl mercury since methyl mercury is on the AUES list. |  |  |  |  |  |  |  |  |  |  |  |  |  |
|  |  |  |  |  | Sample has been removed |  |  |  |  |  |  |  |  |  |  |  |  |  |
|  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
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Table A.7--PARSONS 2007-2008
ANALYTICAL RESULTS FOR 4835 TEST PIT SAMPLES
(12 Metals-Suite)

| SAMPLE ID: |  |  | SW-4835GB-(-90,50)SW-N(5)LE |  | SW-4835GB <br> (-90,50)-SW-N(5) LE2.5 |  | SW-4835GB <br> (-90,50)-SW-N(5)LC-3 |  | $\begin{gathered} \text { SW-4835GB } \\ (-150,50)-\text { SW-S(8)LE } \end{gathered}$ |  | $\begin{gathered} \text { SW-4835GB } \\ (-150,50)-\text { SW-S(8)LC-3 } \end{gathered}$ |  | $\begin{gathered} \text { SW-4835GB } \\ (-190,10) S W-E(7) L N \end{gathered}$ |  | $\begin{gathered} \text { SW-4835GB } \\ (-190,10) \text { SW-E(7)LC } \end{gathered}$ |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| DATE SAMPLED: |  |  | 12/11/07 |  | 01/11/08 |  | 01/11/08 |  | 01/11/08 |  | 01/11/08 |  | 01/16/08 |  | 01/16/08 |  |
| LAB SAMPLE ID: |  |  | 712054-006 |  | 801057-001 |  | 801057-002 |  | 801057-004 |  | 801057-006 |  | 801087-001 |  | 801087-002 |  |
| SAMPLE DEPTH(FT) |  |  | 0.5 |  | 2.5-3 |  | 3-3.5 |  | 0.5 |  | 3-3.5 |  | 0.5 |  | 2-2.5 |  |
|  | Units | Regional Screening Level |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| Total Metals - ILM05.4 |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| Aluminum | mg/kg | 19,100 | 30700 | + | 45900 | + | 31600 | + | 25000 |  | 29300 | + | 19700 |  | 15900 |  |
| Antimony | $\mathrm{mg} / \mathrm{kg}$ | 5.2(BG) | 9.6 | UL+ | 11.7 | UL+ | 10 | UL+ | 0.31 | J | 0.86 | J+ | 0.34 | J | 0.4 | J |
| Arsenic | $\mathrm{mg} / \mathrm{kg}$ | 20(SV) | 2.1 | J | 2 |  | 0.89 |  | 18.8 |  | 19.9 |  | 13 | J | 11.5 | J |
| Barium | $\mathrm{mg} / \mathrm{kg}$ | 1,500 | 116 | + | 139 | + | 139 | + | 101 |  | 107 | + | 82.4 |  | 58.7 |  |
| Cadmium | $\mathrm{mg} / \mathrm{kg}$ | 7.0 | 0.71 | J+ | 0.97 | U+ | 0.83 | U+ | 0.47 | U | 0.88 | U+ | 0.47 | U | 0.42 | U |
| Copper | $\mathrm{mg} / \mathrm{kg}$ | 310 | 314 | + | 236 | + | 444 | + | 90.9 |  | 75.9 | + | 56 | K | 48.6 | K |
| Lead | $\mathrm{mg} / \mathrm{kg}$ | 400 | 5.0 | J+ | 19.4 | J+ | 2.9 | B+ | 9.3 | J | 13.7 | J+ | 46 |  | 33.7 |  |
| Manganese | $\mathrm{mg} / \mathrm{kg}$ | 1,800 | 949 | + | 1800 | J+ | 691 | J+ | 442 | J | 516 | J+ | 523 |  | 472 |  |
| Mercury | $\mathrm{mg} / \mathrm{kg}$ | 0.78* | 0.099 | U | 0.063 | B | 0.055 | B | 0.15 | B | 0.23 | B | 0.14 | B | 0.72 |  |
| Nickel | $\mathrm{mg} / \mathrm{kg}$ | 160 | 92.1 |  | 160 | J+ | 93.5 | J | 61 | J | 64.2 | J+ | 48.3 | J | 44.5 | J |
| Thallium | $\mathrm{mg} / \mathrm{kg}$ | 2.20 | 2.6 | J+ | 2.6 | J+ | 2.4 | J+ | 0.93 | J | 1.5 | J+ | 0.9 | J | 0.97 | J |
| Vanadium | $\mathrm{mg} / \mathrm{kg}$ | 390.00 | 142 | + | 231 | + | 143 | + | 100 |  | 103 | + | 78.1 |  | 66.7 |  |
| Zinc | $\mathrm{mg} / \mathrm{kg}$ | 2,300 | 71.6 | + | 83.9 | + | 76.5 | + | 65.3 |  | 78.9 | + | 83.5 |  | 66.9 |  |
|  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| Comparison value based on the higher of the adjusted Sept 2008 Regional Screening Level (RSL) (if non-carcinogenic) or the 2007 Background value. |  |  |  |  | U - Analyte was analyzed for but not detected above the adjusted practical quantitation limit (PQL). |  |  |  |  |  |  |  |  |  |  |  |
| NA - Not available. |  |  |  |  | UJ - Analyte not detected, reported PQL may be inaccurate or imprecise. |  |  |  |  |  |  |  |  |  |  |  |
| BG - Background Value (2007 Study). |  |  |  |  | UL - Analyte not detected, reported PQL is biased low, actual PQL is expected to be higher. |  |  |  |  |  |  |  |  |  |  |  |
| SV - Spring Valley Remediation Goal. |  |  |  |  | $J$ - Analyte detected, estimated concentration. |  |  |  |  |  |  |  |  |  |  |  |
| (NO CODE) - Confirmed identification. |  |  |  |  | L - Analyte detected, reported result is biased low, actual value is expected to be higher. |  |  |  |  |  |  |  |  |  |  |  |
| + - Result reported from diluted sample. |  |  |  |  | K - Analyte detected, reported result is biased high, actual value is expected to be lower. |  |  |  |  |  |  |  |  |  |  |  |
| Detections are bolded. |  |  |  |  | Detections exceeding the comparison level are shown shaded and bolded. |  |  |  |  |  |  |  |  |  |  |  |
| Sample has been removed |  |  |  |  | * RSL for methyl mercury since methyl mercury is on the AUES list. |  |  |  |  |  |  |  |  |  |  |  |
|  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
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Table A.7--PARSONS 2007-2008
ANALYTICAL RESULTS FOR 4835 TEST PIT SAMPLES
(12 Metals-Suite)

| SAMPLE ID: |  |  | $\begin{gathered} \text { SW-4835GB } \\ (-150,50) \text { SWS(8)2.5 } \end{gathered}$ |  | $\begin{gathered} \text { SW-4835GB } \\ (-150,50) \text { SWS(8)LE2.5 } \end{gathered}$ |  | SW-4835GB (-90,50)SWN(5)2.5 |  | $\begin{gathered} \text { SW-4835GB- } \\ (-170,50) \end{gathered}$ |  | SW-4835GB-$(-150,10)$ |  | SW-4835GB-$(-150,30)-2$ |  | SW-4835GB-(-150,-10)SW-E |  | SW-4835GB-$(-170,30)-4$ |  | SW-4835GB-$(-150,-10)-2$ |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| DATE SAMPLED: |  |  | 01/29/08 |  | 01/29/08 |  | 01/29/08 |  | 02/27/08 |  | 02/27/08 |  | 03/06/08 |  | 03/05/08 |  | 03/10/08 |  | 03/10/08 |  |
| LAB SAMPLE ID: |  |  | 801136-002 |  | 801136-001 |  | 801136-003 |  | 802139-004 |  | 802139-002 |  | 803060-003 |  | 803061-002 |  | 803063-001 |  | 803063-002 |  |
| FAMPLE DEPTH(FT) |  |  | 2.5 |  | 2.5 |  | 2.5-3 |  | 0.5 |  | 0.5 |  | 2 |  | 0.5 |  | 4 |  | 2 |  |
|  | Units | Regional Screening Level |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| Total Metals - ILM05.4 |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| Aluminum | $\mathrm{mg} / \mathrm{kg}$ | 19,100 | 32400 |  | 25300 |  | 55900 | + | 22800 |  | 22400 |  | 32600 |  | 20900 | + | 19000 |  | 24700 |  |
| Antimony | $\mathrm{mg} / \mathrm{kg}$ | 5.2(BG) | 5.3 | UL | 5.2 | UL | 26.1 | UL+ | 5.7 | UL | 0.56 | J | 0.39 | J | 24.9 | U+ | 5.1 | U | 0.35 | J |
| Arsenic | $\mathrm{mg} / \mathrm{kg}$ | 20(SV) | 16.6 |  | 17.9 |  | 3.9 | J | 2.0 | B | 14.6 | K | 2.5 | J | 9.7 | K | 10.2 |  | 19.9 |  |
| Barium | $\mathrm{mg} / \mathrm{kg}$ | 1,500 | 125 |  | 99.3 |  | 236 | + | 40.0 |  | 58.8 |  | 119 |  | 55.3 | J+ | 69.2 |  | 126 |  |
| Cadmium | $\mathrm{mg} / \mathrm{kg}$ | 7.0 | 0.41 | J | 0.13 | B | 0.38 | B+ | 0.025 | B | 0.19 | B | 0.33 | J | 0.3 | B+ | 0.35 | J | 0.92 |  |
| Copper | $\mathrm{mg} / \mathrm{kg}$ | 310 | 123 | K | 63.0 | K | 231 | K+ | 57.5 | K | 66.5 | K | 56.3 |  | 52.9 | + | 48 |  | 65.6 |  |
| Lead | $\mathrm{mg} / \mathrm{kg}$ | 400 | 10.7 |  | 14.1 |  | 14.0 | + | 6.3 | K | 11.5 | K | 15.1 |  | 22.8 | + | 6 |  | 7.7 |  |
| Manganese | $\mathrm{mg} / \mathrm{kg}$ | 1,800 | 656 |  | 446 |  | 1290 | + | 235 |  | 734 |  | 365 |  | 388 | + | 611 |  | 530 |  |
| Mercury | $\mathrm{mg} / \mathrm{kg}$ | 0.78* | 0.075 | J | 0.1 | U | 0.11 | U | 0.1 | U | 0.039 | J | 0.11 | U | 0.11 | U | 0.1 | U | 0.1 | UJ |
| Nickel | $\mathrm{mg} / \mathrm{kg}$ | 160 | 79.5 |  | 39.8 |  | 145 |  | 83.8 |  | 69.0 |  | 58.5 |  | 95.6 |  | 75.9 |  | 40.9 |  |
| Thallium | $\mathrm{mg} / \mathrm{kg}$ | 2.20 | 2.8 |  | 2.2 | U | 8.7 | J+ | 2.4 | U | 2.5 | U | 2.3 | U | 10.4 | U+ | 2.1 | $u$ | 2.1 | $\cup$ |
| Vanadium | $\mathrm{mg} / \mathrm{kg}$ | 390.00 | 130 |  | 79.1 |  | 232 | + | 67.3 |  | 93.7 |  | 39.4 |  | 103 | + | 96.6 |  | 83.8 |  |
| Zinc | $\mathrm{mg} / \mathrm{kg}$ | 2,300 | 83.3 |  | 73.0 |  | 139 | + | 53.5 |  | 62.9 |  | 33.7 |  | 96.8 | + | 41 |  | 72.1 |  |
|  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| Comparison value based on the higher of the adjusted Sept 2008 Regional Screening Level (RSL) (if non-carcinogenic) or the 2007 Background value. |  |  |  |  | U - Analyte was analyzed for but not detected above the adjusted practical quantitation limit (PQL). |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| NA - Not available. |  |  |  |  | UJ - Analyte not detected, reported PQL may be inaccurate or imprecise. |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| BG - Background Value (2007 Study). |  |  |  |  | UL - Analyte not detected, reported PQL is biased low, actual PQL is expected to be higher. |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| SV - Spring Valley Remediation Goal. |  |  |  |  | $J$ - Analyte detected, estimated concentration. |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| (NO CODE) - Confirmed identification. |  |  |  |  | L - Analyte detected, reported result is biased low, actual value is expected to be higher. |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| + - Result reported from diluted sample. |  |  |  |  | K - Analyte detected, reported result is biased high, actual value is expected to be lower. |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| Detections are bolded. |  |  |  |  | Detections exceeding the comparison level are shown shaded and bolded. |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| Sample has been removed |  |  |  |  | * RSL for methyl mercury since methyl mercury is on the AUES list. |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
|  |  |  |  |  | Sample has been removed |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
|  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
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Table A.7--PARSONS 2007-2008

## ANALYTICAL RESULTS FOR 4835 TEST PIT SAMPLES

(12 Metals-Suite)

| SAMPLE ID: |  |  | $\begin{array}{r} \text { SW-483 } \\ (-150, \end{array}$ |  | $\begin{gathered} \text { SW-483 } \\ (-190,5 \end{gathered}$ |  | $\begin{aligned} & \text { SW-48: } \\ & (-170,- \end{aligned}$ |  | $\begin{array}{r} \text { SW-48 } \\ (-190,50) \end{array}$ | $\begin{aligned} & \text { GB- } \\ & I-N(5) \end{aligned}$ | $\begin{array}{r} \text { SW-483 } \\ (-190,50) S \end{array}$ |  | $\begin{gathered} \text { SW-48: } \\ (-170,30 \end{gathered}$ |  | $\begin{array}{r} \text { SW-483 } \\ (-170,30) S \end{array}$ | B- E-3.5 | $\begin{array}{r} \text { SW-48 } \\ (-150,30) S \end{array}$ | B- $E(5) L N$ | $\begin{array}{r} \text { SW-483 } \\ (-150,30) S \end{array}$ | $\begin{aligned} & \text { W(5) } \\ & \hline \end{aligned}$ |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| DATE SAMPLED: |  |  | 03/10 |  | 03/13 |  | 03/1 |  | 03/1 |  | 03/13 |  | 03/13 |  | 03/13 |  | 03/1 |  | 03/13 |  |
| LAB SAMPLE ID: |  |  | 803063 |  | 803089 |  | 80308 |  | 80309 |  | 803090 |  | 80309 |  | 803090 |  | 80309 |  | 803090 |  |
| SAMPLE DEPTH(FT) |  |  | 0.5 |  | 5 |  | 3-3.5 |  | 0.5 |  | 0.5 |  | 0.5 |  | 3.5 |  | 0.5 |  | 0.5 |  |
|  | Units | Regional Screening Level |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| Total Metals - ILM05 |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| Aluminum | mg/kg | 19,100 | 31700 | + | 26400 | + | 39600 | + | 16300 |  | 17500 |  | 29900 | + | 25100 | + | 32400 | + | 26100 |  |
| Antimony | $\mathrm{mg} / \mathrm{kg}$ | 5.2(BG) | 25.1 | U+ | 56.3 | U+ | 51.9 | U+ | 0.63 | L | 5.2 | UL | 10.9 | UL+ | 50.9 | UL+ | 53 | UL+ | 0.25 | L |
| Arsenic | mg/kg | 20(SV) | 17.1 |  | 7.8 |  | 10.5 |  | 10.4 | J | 1.1 | J | 2.8 | J | 3.3 | J | 5.8 | J | 19.5 | J |
| Barium | mg/kg | 1,500 | 122 | + | 55.2 | J+ | 70.4 | J+ | 77.8 |  | 38.6 |  | 37.9 | + | 39.6 | J+ | 54.5 | J+ | 46.6 |  |
| Cadmium | mg/kg | 7.0 | 2.1 | U+ | 0.49 | B+ | 0.39 | B+ | 0.18 | J | 0.048 | J | 0.62 | J+ | 4.2 | UJ+ | 4.4 | UJ+ | 0.27 | J |
| Copper | $\mathrm{mg} / \mathrm{kg}$ | 310 | 192 | + | 123 | + | 128 | + | 37.4 | K | 34.7 | K | 67.8 | K+ | 99 | K+ | 85.3 | K+ | 66.4 | K |
| Lead | mg/kg | 400 | 4.3 | B+ | 7.9 | B+ | 9.8 | B+ | 12.9 | J | 4.8 | J | 8.5 | J+ | 6.3 | B+ | 9.8 | B+ | 10.3 | J |
| Manganese | mg/kg | 1,800 | 773 | + | 486 | + | 249 | + | 1200 | L+ | 281 | L | 256 | L+ | 363 | L+ | 241 | L+ | 144 | L |
| Mercury | mg/kg | 0.78* | 0.034 | J | 0.052 | J | 0.035 | J | 0.17 | J | 0.1 | UJ | 0.097 | UJ | 0.096 | UJ | 0.099 | UJ | 0.026 | J |
| Nickel | $\mathrm{mg} / \mathrm{kg}$ | 160 | 45.4 |  | 87.7 |  | 46.7 |  | 52.8 |  | 65.3 |  | 94.7 |  | 88.2 |  | 96.8 |  | 70.6 |  |
| Thallium | $\mathrm{mg} / \mathrm{kg}$ | 2.20 | 10.4 | U+ | 23.4 | U+ | 21.6 | U+ | 0.55 | J | 2.2 | UJ | 4.5 | UJ+ | 21.2 | UJ+ | 22.1 | UJ+ | 2.2 | UJ |
| Vanadium | mg/kg | 390.00 | 198 | + | 139 | + | 202 | + | 70 |  | 49.9 |  | 105 | + | 123 | + | 114 | + | 114 |  |
| Zinc | $\mathrm{mg} / \mathrm{kg}$ | 2,300 | 65.4 | + | 57.7 | + | 81.2 | + | 49.6 |  | 43.7 |  | 60.4 | + | 54.5 | + | 67.9 | + | 61.8 |  |
|  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| Comparison value based on the higher of the adjusted Sept 2008 Regional Screening Level (RSL) (if non-carcinogenic) or the 2007 Background value. |  |  |  |  | U - Analyte was analyzed for but not detected above the adjusted practical quantitation limit (PQL). |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| NA - Not available. |  |  |  |  | UJ - Analyte not detected, reported PQL may be inaccurate or imprecise. |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| BG - Background Value (2007 Study). |  |  |  |  | UL - Analyte not detected, reported PQL is biased low, actual PQL is expected to be higher. |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| SV - Spring Valley Remediation Goal. |  |  |  |  | J - Analyte detected, estimated concentration. |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| (NO CODE) - Confirmed identification. |  |  |  |  | L - Analyte detected, reported result is biased low, actual value is expected to be higher. |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| + - Result reported from diluted sample. |  |  |  |  | K - Analyte detected, reported result is biased high, actual value is expected to be lower. |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| Detections are bolded. |  |  |  |  | Detections exceeding the comparison level are shown shaded and bolded. |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| Sample has been removed |  |  |  |  | * RSL for methyl mercury since methyl mercury is on the AUES list. |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
|  |  |  |  |  | Sample has been removed |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
|  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
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Table A.7--PARSONS 2007-2008
ANALYTICAL RESULTS FOR 4835 TEST PIT SAMPLES
(12 Metals-Suite)

| SAMPLE ID: |  |  | SW-4835GB-$(-170,10)$ SW-N |  | $\begin{aligned} & \text { 4835GB- } \\ & (-190,30)-5 \end{aligned}$ |  | $\begin{gathered} \text { 4835GB- } \\ (-190,30)-\text { SW-N(4.5) } \end{gathered}$ |  | $\begin{gathered} \text { 4835GB- } \\ (-190,30) \text {-SW-N } \end{gathered}$ |  | 4835GB- <br> (-170,30)SW-S(5)-3.5 |  | 4835GB- <br> (-170,30)SW-S(5)LW |  | 4835GB- <br> (-150,30)SW-W(5)LC |  | SW-4835GB- <br> $(190,50)$ SW-S(5)LC |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| DATE SAMPLED: |  |  | 03/13/08 |  | 03/27/08 |  | 03/27/08 |  | $03 / 27 / 08$ |  | 03/28/08 |  | 03/28/08 |  | 03/28/08 |  | 03/31/08 |  |
| LAB SAMPLE ID: |  |  | 803090-008 |  | 803143-005 |  | 803144-001 |  | 803144-002 |  | 803143-001 |  | 803143-003 |  | 803143-004 |  | 804002-001 |  |
| SAMPLE DEPTH(FT) |  |  | 0.5 |  |  |  |  |  |  |  | 3.5 |  | 3.5 |  | 2 |  | 5 |  |
|  | Units | Regional Screening Level |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| Total Metals - ILM05.4 |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| Aluminum | $\mathrm{mg} / \mathrm{kg}$ | 19,100 | 35300 | + | 23000 |  | 12900 |  | 14400 |  | 35500 |  | 25300 |  | 34700 | + | 39600 | + |
| Antimony | mg/kg | 5.2(BG) | 52.2 | UL+ | 3.8 | L | 1.4 | J | 1.9 | J | 1.9 | L | 2.1 | L | 2.5 | L+ | 2.6 | J+ |
| Arsenic | mg/kg | 20(SV) | 19.2 | J | 13.3 | J | 1.9 |  | 11.6 |  | 1.2 | J | 3.4 | J | 2.4 | J | 1.6 |  |
| Barium | mg/kg | 1,500 | 125 | J+ | 52.4 |  | 51.6 |  | 60.4 |  | 164 |  | 58.5 |  | 82.4 | + | 119 | + |
| Cadmium | mg/kg | 7.0 | 4.3 | UJ+ | 0.15 | B | 0.24 | B | 0.092 | B | 0.25 | B | 0.1 | B | 0.19 | B+ | 0.28 | B+ |
| Copper | mg/kg | 310 | 134 | K+ | 69.1 | L | 18.1 |  | 24.5 |  | 108 | L | 82.3 | L | 74.9 | L+ | 158 | + |
| Lead | mg/kg | 400 | 13.8 | B+ | 4.3 | J | 5.1 |  | 10.2 |  | 7.8 |  | 10.2 |  | 9.4 | + | 6.1 | + |
| Manganese | $\mathrm{mg} / \mathrm{kg}$ | 1,800 | 748 | L+ | 680 | L | 610 |  | 443 |  | 649 | L | 133 | L | 370 | L+ | 361 | + |
| Mercury | $\mathrm{mg} / \mathrm{kg}$ | 0.78* | 0.83 | J | 0.19 | L | 0.27 |  | 0.14 |  | 0.052 | L | 0.11 | UL | 0.11 | UL | 0.044 | J |
| Nickel | $\mathrm{mg} / \mathrm{kg}$ | 160 | 83 |  | 89.6 |  | 63.5 |  | 38.3 |  | 86.6 |  | 67.1 |  | 84.4 |  | 84.2 |  |
| Thallium | $\mathrm{mg} / \mathrm{kg}$ | 2.20 | 21.7 | UJ+ | 2.2 | UJ | 2.2 | U | 2.2 | U | 2.3 | UJ | 2.2 | UJ | 4.6 | UJ+ | 5.2 | U+ |
| Vanadium | $\mathrm{mg} / \mathrm{kg}$ | 390.00 | 138 | + | 119 |  | 33.2 |  | 61 |  | 110 |  | 106 |  | 125 | + | 197 | + |
| Zinc | $\mathrm{mg} / \mathrm{kg}$ | 2,300 | 83.9 | + | 53.7 |  | 58.5 |  | 46.4 |  | 75.2 |  | 61.9 |  | 78.5 | + | 82.7 | + |
|  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| Comparison value based on the higher of the adjusted Sept 2008 Regional Screening Level (RSL) (if non-carcinogenic) or the 2007 Background value. |  |  |  |  | U - Analyte was analyzed for but not detected above the adjusted practical quantitation limit (PQL). |  |  |  |  |  |  |  |  |  |  |  |  |  |
| NA - Not available. |  |  |  |  | UJ - Analyte not detected, reported PQL may be inaccurate or imprecise. |  |  |  |  |  |  |  |  |  |  |  |  |  |
| BG - Background Value (2007 Study). |  |  |  |  | UL - Analyte not detected, reported PQL is biased low, actual PQL is expected to be higher. |  |  |  |  |  |  |  |  |  |  |  |  |  |
| SV - Spring Valley Remediation Goal. |  |  |  |  | J - Analyte detected, estimated concentration. |  |  |  |  |  |  |  |  |  |  |  |  |  |
| (NO CODE) - Confirmed identification. |  |  |  |  | L - Analyte detected, reported result is biased low, actual value is expected to be higher. |  |  |  |  |  |  |  |  |  |  |  |  |  |
| + - Result reported from diluted sample. |  |  |  |  | K - Analyte detected, reported result is biased high, actual value is expected to be lower. |  |  |  |  |  |  |  |  |  |  |  |  |  |
| Detections are bolded. |  |  |  |  | Detections exceeding the comparison level are shown shaded and bolded. |  |  |  |  |  |  |  |  |  |  |  |  |  |
| Sample has been removed |  |  |  |  | * RSL for methyl mercury since methyl mercury is on the AUES list. |  |  |  |  |  |  |  |  |  |  |  |  |  |
|  |  |  |  |  | Sample has been removed |  |  |  |  |  |  |  |  |  |  |  |  |  |
|  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
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Table A.7--PARSONS 2007-2008
ANALYTICAL RESULTS FOR 4835 TEST PIT SAMPLES
(12 Metals-Suite)

| SAMPLE ID: |  |  | $\begin{gathered} \text { SW-4835GB- } \\ (170,30) \text { SW-S(5)-LC5 } \end{gathered}$ |  | SW-4835GB <br> (-170,30)SW-S(5)LW4.5 |  | SW-4835GB (-170,30)SW-S(5)-4.5 |  | $\begin{gathered} \text { SW-4835GB } \\ (-150,-10) \text { SW-W }(10) L C 3 \end{gathered}$ |  | $\begin{gathered} \text { SW-4835GB } \\ (-150,-10) \text { SW-W(10)LC4 } \end{gathered}$ |  | SW-4835GB-$(-150,-10)-S W-W(10) L S-2.5$ |  | SW-4835GB-$(-150,-10)-S W-W(10)-2.5$ |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| DATE SAMPLED: |  |  | 04/01/08 |  | 04/02/08 |  | 04/02/08 |  | 04/02/08 |  | 04/02/08 |  | 04/03/08 |  | 04/03/08 |  |
| LAB SAMPLE ID: |  |  | 804006-001 |  | 804013-001 |  | 804013-003 |  | 804013-005 |  | 804013-006 |  | 804044-004 |  | 804044-002 |  |
| SAMPLE DEPTH(FT) |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
|  | Units | Regional Screening Level |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| Total Metals - ILM05.4 |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| Aluminum | $\mathrm{mg} / \mathrm{kg}$ | 19,100 | 32200 |  | 14000 |  | 40200 |  | 30300 | + | 27700 |  | 26700 |  | 40200 |  |
| Antimony | $\mathrm{mg} / \mathrm{kg}$ | 5.2(BG) | 1.1 | L | 1.9 | L | 2.2 | L | 0.81 | L+ | 1.3 | L | 0.41 | J | 0.79 | J |
| Arsenic | $\mathrm{mg} / \mathrm{kg}$ | 20(SV) | 4.1 | J | 3.1 | K | 3.1 | K | 9.7 | K | 1.2 | K | 17.1 | J | 0.69 | J |
| Barium | $\mathrm{mg} / \mathrm{kg}$ | 1,500 | 136 |  | 46.1 |  | 175 |  | 121 | + | 116 |  | 108 |  | 202 |  |
| Cadmium | $\mathrm{mg} / \mathrm{kg}$ | 7.0 | 0.2 | J | 0.5 | J | 0.26 | J | 0.12 | B+ | 0.14 | B | 0.25 | B | 0.33 | J |
| Copper | $\mathrm{mg} / \mathrm{kg}$ | 310 | 58.6 |  | 34.9 | K | 68.4 | K | 53.3 | K+ | 50.6 | K | 72.8 | J | 51.3 | J |
| Lead | $\mathrm{mg} / \mathrm{kg}$ | 400 | 10.5 | L | 5.7 | K | 9.9 | K | 6.3 | K+ | 7.7 | K | 10.7 |  | 8.4 |  |
| Manganese | $\mathrm{mg} / \mathrm{kg}$ | 1,800 | 893 | L | 1240 | K+ | 892 | K+ | 527 | K+ | 650 | K | 427 |  | 772 |  |
| Mercury | $\mathrm{mg} / \mathrm{kg}$ | 0.78* | 0.11 | UL | 0.1 | U | 0.11 | U | 0.1 | U | 0.1 | U | 0.079 | J | 0.035 | J |
| Nickel | $\mathrm{mg} / \mathrm{kg}$ | 160 | 82.4 |  | 104 |  | 73.3 |  | 43.7 |  | 12.3 | J | 41.8 |  | 42.4 |  |
| Thallium | $\mathrm{mg} / \mathrm{kg}$ | 2.20 | 2.5 | U | 2.1 | UL | 2.5 | UL | 4.4 | UL+ | 2.1 | UL | 2.1 | U | 2.2 | U |
| Vanadium | $\mathrm{mg} / \mathrm{kg}$ | 390.00 | 135 |  | 67.3 |  | 88 |  | 103 | + | 109 |  | 82.6 |  | 139 |  |
| Zinc | mg/kg | 2,300 | 59.2 |  | 44.3 |  | 68.7 |  | 118 | + | 90.2 |  | 85.2 |  | 83.1 |  |
|  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| Comparison value based on the higher of the adjusted Sept 2008 Regional Screening Level (RSL) (if non-carcinogenic) or the 2007 Background value. |  |  |  |  | U - Analyte was analyzed for but not detected above the adjusted practical quantitation limit (PQL). |  |  |  |  |  |  |  |  |  |  |  |
| NA - Not available. |  |  |  |  | UJ - Analyte not detected, reported PQL may be inaccurate or imprecise. |  |  |  |  |  |  |  |  |  |  |  |
| BG - Background Value (2007 Study). |  |  |  |  | UL - Analyte not detected, reported PQL is biased low, actual PQL is expected to be higher. |  |  |  |  |  |  |  |  |  |  |  |
| SV - Spring Valley Remediation Goal. |  |  |  |  | $J$ - Analyte detected, estimated concentration. |  |  |  |  |  |  |  |  |  |  |  |
| (NO CODE) - Confirmed identification. |  |  |  |  | L - Analyte detected, reported result is biased low, actual value is expected to be higher. |  |  |  |  |  |  |  |  |  |  |  |
| + - Result reported from diluted sample. |  |  |  |  | K - Analyte detected, reported result is biased high, actual value is expected to be lower. |  |  |  |  |  |  |  |  |  |  |  |
| Detections are bolded. |  |  |  |  | Detections exceeding the comparison level are shown shaded and bolded. |  |  |  |  |  |  |  |  |  |  |  |
| Sample has been removed |  |  |  |  | * RSL for methyl mercury since methyl mercury is on the AUES list. |  |  |  |  |  |  |  |  |  |  |  |
|  |  |  |  |  | Sample has been removed |  |  |  |  |  |  |  |  |  |  |  |
|  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
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Table A.7--PARSONS 2007-2008
ANALYTICAL RESULTS FOR 4835 TEST PIT SAMPLES
(12 Metals-Suite)

| SAMPLE ID: |  |  | $\begin{array}{r} S W-48 \\ (-190,50)-S \end{array}$ | $5)-4.5$ | $\begin{array}{r} S W-48: \\ (-190,50)-S \end{array}$ | $\mathrm{V}(5) \mathrm{LC}$ | $\begin{array}{r} \mathrm{SW}-48 \\ (-190,50)-\mathrm{S} \end{array}$ | (5)-4.5 | $\begin{array}{r} \text { SW-483 } \\ (-170,-10) \end{array}$ | GB- W-S-3 | $\begin{array}{r} \text { SW-48 } \\ (-190,50) \text { SU } \end{array}$ | $-w(5)$ | $\begin{array}{r} \text { SW-48: } \\ (-190,50) \text { SWN } \end{array}$ | $5)-4.5$ | $\begin{array}{r} \text { SW-48 } \\ (-190,50) \text { SWN } \end{array}$ | (5)LN |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| DATE SAMPLED: |  |  | 04/0 |  | 04/07 |  | 04/0 |  | 04/08 |  | 04/1 |  | 04/1 |  | 04/16 |  |
| LAB SAMPLE ID: |  |  | 80404 |  | 80404 |  | 80404 |  | 80404 |  | 80411 |  | 80411 |  | 804115 |  |
| SAMPLE DEPTH(FT) |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
|  | Units | Regional Screening Level |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| Total Metals - ILM05 |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| Aluminum | $\mathrm{mg} / \mathrm{kg}$ | 19,100 | 17700 |  | 24600 |  | 41800 |  | 46700 |  | 17400 |  | 12700 |  | 19200 |  |
| Antimony | mg/kg | 5.2(BG) | 0.45 |  | 0.67 |  | 1.8 |  | 0.92 |  | 0.32 | J | 5.6 | UL | 0.30 | J |
| Arsenic | $\mathrm{mg} / \mathrm{kg}$ | 20(SV) | 13.1 |  | 9.8 |  | 5 |  | 17.8 |  | 9.1 | J | 3.0 | J | 4.6 | J |
| Barium | mg/kg | 1,500 | 86.9 |  | 131 |  | 254 |  | 94.1 |  | 85.8 |  | 143 |  | 77.8 |  |
| Cadmium | mg/kg | 7.0 | 0.081 |  | 0.47 | U | 0.28 |  | 0.46 |  | 0.34 | J | 0.29 | J | 0.49 |  |
| Copper | $\mathrm{mg} / \mathrm{kg}$ | 310 | 41.3 |  | 73.4 |  | 343 |  | 222 |  | 42.4 |  | 21.7 |  | 44.0 |  |
| Lead | mg/kg | 400 | 15.7 |  | 13.7 |  | 7.2 |  | 6.5 |  | 23.0 |  | 8.9 |  | 8.3 |  |
| Manganese | $\mathrm{mg} / \mathrm{kg}$ | 1,800 | 960 |  | 1420 |  | 4110 |  | 645 |  | 982 | + | 2680 | + | 1120 | + |
| Mercury | $\mathrm{mg} / \mathrm{kg}$ | 0.78* | 0.095 | J | 0.05 | J | 0.11 | U | 0.042 | J | 0.089 | J | 0.011 | U | 0.054 | J |
| Nickel | $\mathrm{mg} / \mathrm{kg}$ | 160 | 55.6 |  | 71.2 |  | 138 |  | 54.5 |  | 56.6 |  | 72.4 |  | 64.1 |  |
| Thallium | mg/kg | 2.20 | 2.3 | U | 1.2 |  | 12 | UD | 4.3 | U+ | 2.2 | U | 0.88 | J | 2.2 | U |
| Vanadium | mg/kg | 390.00 | 67.9 |  | 95.7 |  | 265 |  | 345 |  | 77.5 | K | 74.3 | K | 93.6 | K |
| Zinc | $\mathrm{mg} / \mathrm{kg}$ | 2,300 | 60 |  | 66.2 |  | 94.4 |  | 84.5 |  | 61.0 |  | 46.9 |  | 51.7 |  |
|  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| Comparison value based on the higher of the adjusted Sept 2008 Regional Screening Level (RSL) (if non-carcinogenic) or the 2007 Background value. |  |  |  |  | U - Analyte was analyzed for but not detected above the adjusted practical quantitation limit (PQL). |  |  |  |  |  |  |  |  |  |  |  |
| NA - Not available. |  |  |  |  | UJ - Analyte not detected, reported PQL may be inaccurate or imprecise. |  |  |  |  |  |  |  |  |  |  |  |
| BG - Background Value (2007 Study). |  |  |  |  | UL - Analyte not detected, reported PQL is biased low, actual PQL is expected to be higher. |  |  |  |  |  |  |  |  |  |  |  |
| SV - Spring Valley Remediation Goal. |  |  |  |  | $J$ - Analyte detected, estimated concentration. |  |  |  |  |  |  |  |  |  |  |  |
| (NO CODE) - Confirmed identification. |  |  |  |  | L - Analyte detected, reported result is biased low, actual value is expected to be higher. |  |  |  |  |  |  |  |  |  |  |  |
| + - Result reported from diluted sample. |  |  |  |  | K - Analyte detected, reported result is biased high, actual value is expected to be lower. |  |  |  |  |  |  |  |  |  |  |  |
| Detections are bolded. |  |  |  |  | Detections exceeding the comparison level are shown shaded and bolded. |  |  |  |  |  |  |  |  |  |  |  |
| Sample has been removed |  |  |  |  | * RSL for methyl mercury since methyl mercury is on the AUES list. |  |  |  |  |  |  |  |  |  |  |  |
|  |  |  |  |  | Sample has been removed |  |  |  |  |  |  |  |  |  |  |  |
|  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
|  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |

Table A.7--PARSONS 2007-2008
ANALYTICAL RESULTS FOR 4835 TEST PIT SAMPLES
(12 Metals-Suite)


Table A.7--PARSONS 2007-2008
ANALYTICAL RESULTS FOR 4835 TEST PIT SAMPLES
(12 Metals-Suite)


Table A.7--PARSONS 2007-2008
ANALYTICAL RESULTS FOR 4835 TEST PIT SAMPLES
(12 Metals-Suite)

| SAMPLE ID: |  |  | SW-4835GB-(-170,30)SDPIPE(N) |  | $\begin{gathered} \text { SW-4835GB- } \\ (-90,30)-4 \end{gathered}$ |  | $\begin{gathered} \text { SW-4835GB- } \\ (-90,30) \text { SW-W(15)- } \\ 3.5 \end{gathered}$ |  | SW-4835GB-(-90,30)SW-W(15)-0.5 |  | SW-4835GB- <br> (-90,30)SW-W(15)LE-4.0 |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| DATE SAMPLED: |  |  | 05/06/08 |  | 11/12/08 |  | 11/19/08 |  | 11/25/08 |  | 11/25/08 |  |
| LAB SAMPLE ID: |  |  | 805036-003 |  | 811103-004 |  | 811174-002 |  | 811230-001 |  | 811230-002 |  |
| SAMPLE DEPTH(FT) |  |  |  |  | 4 |  | 3.5 |  | 0.5 |  | 4 |  |
|  | Units | Regional Screening Level |  |  |  |  |  |  |  |  |  |  |
| Total Metals - ILM05.4 |  |  |  |  |  |  |  |  |  |  |  |  |
| Aluminum | mg/kg | 19,100 | 23200 |  | 28200 |  | 29200 |  | 28800 |  | 26200 |  |
| Antimony | mg/kg | 5.2(BG) | 2.6 | J | 10.4 | U | 12.1 | U | 1 | J | 6.5 | U |
| Arsenic | mg/kg | 20(SV) | 31.6 | K | 19.4 |  | 10.2 |  | 14.4 |  | 9.6 |  |
| Barium | mg/kg | 1,500 | 57.4 |  | 105 |  | 194 |  | 120 |  | 147 |  |
| Cadmium | mg/kg | 7.0 | 0.39 | J | 0.87 | U | 1 | U | 0.53 | U | 0.54 | U |
| Copper | mg/kg | 310 | 83.5 | K | 60 |  | 31.0 |  | 88.6 |  | 33.2 |  |
| Lead | mg/kg | 400 | 9.4 |  | 7.8 |  | 8.8 |  | 11.6 |  | 10.2 |  |
| Manganese | mg/kg | 1,800 | 537 |  | 458 |  | 475 |  | 615 |  | 413 |  |
| Mercury | mg/kg | 0.78* | 0.20 | J | 0.06 | J | 0.11 | U | 0.06 | J | 0.03 | J |
| Nickel | mg/kg | 160 | 102 | K | 47.50 |  | 19.6 |  | 74.3 |  | 19.1 |  |
| Thallium | mg/kg | 2.20 | 2.8 | U | 4.3 | U | 5 | U | 0.67 | J | 2.7 | U |
| Vanadium | $\mathrm{mg} / \mathrm{kg}$ | 390.00 | 110 |  | 85.9 |  | 95.1 |  | 105 |  | 80.2 |  |
| Zinc | $\mathrm{mg} / \mathrm{kg}$ | 2,300 | 61.7 |  | 71.3 |  | 73.0 |  | 69.7 |  | 68.4 |  |
|  |  |  |  |  |  |  |  |  |  |  |  |  |
| Comparison value based on the higher of the adjusted Sept 2008 Regional Screening Level (RSL) (if non-carcinogenic) or the 2007 Background value. |  |  |  |  | U - Analyte was analyzed for but not detected above the adjusted practical quantitation limit (PQL). |  |  |  |  |  |  |  |
| NA - Not available. |  |  |  |  | UJ - Analyte not detected, reported PQL may be inaccurate or imprecise. |  |  |  |  |  |  |  |
| BG - Background Value (2007 Study). |  |  |  |  | UL - Analyte not detected, reported PQL is biased low, actual PQL is expected to be higher. |  |  |  |  |  |  |  |
| SV - Spring Valley Remediation Goal. |  |  |  |  | $J$ - Analyte detected, estimated concentration. |  |  |  |  |  |  |  |
| (NO CODE) - Confirmed identification. |  |  |  |  | L - Analyte detected, reported result is biased low, actual value is expected to be higher. |  |  |  |  |  |  |  |
| + - Result reported from diluted sample. |  |  |  |  | K - Analyte detected, reported result is biased high, actual value is expected to be lower. |  |  |  |  |  |  |  |
| Detections are bolded. |  |  |  |  | Detections exceeding the comparison level are shown shaded and bolded. |  |  |  |  |  |  |  |
| Sample has been removed |  |  |  |  | * RSL for methyl mercury since methyl mercury is on the AUES list. |  |  |  |  |  |  |  |
|  |  |  |  |  | Sample has been removed |  |  |  |  |  |  |  |
|  |  |  |  |  |  |  |  |  |  |  |  |  |
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Table A.8--PARSONS 2007
VALIDATED ANALYTICAL RESULTS FOR TEST PIT GRAB SOIL SAMPLES (SPRING VALLEY-SPECIFIC ANALYTES ONLY)


Table A.8--PARSONS 2007
VALIDATED ANALYTICAL RESULTS FOR TEST PIT GRAB SOIL SAMPLES (SPRING VALLEY-SPECIFIC ANALYTES ONLY)


Table A.8--PARSONS 2007
VALIDATED ANALYTICAL RESULTS FOR TEST PIT GRAB SOIL SAMPLES (SPRING VALLEY-SPECIFIC ANALYTES ONLY)

| SAMPLE ID: |  |  | SW-4835GB-01 (assoc w/TP-17) |  | SW-4835GB-04 (assoc w/ TP-40) |  | SW-4835GB-02 <br> (assoc wl TP-40) |  | SW-4835GB-TP56-001 (assoc w/ TP-56) |  | SW-4835GB-TP49-001 (assoc wl TP-49) |  | SW-4835GB-16 (assoc wl TP-49) |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| DATE SAMPLED: |  |  | 10/22/07 |  | 10/31/07 |  | 10/23/2007 |  | 11/28/2007 |  | 12/13/2007 |  | 12/13/2007 |  |
| SAMPLE DEPTH (FT) |  |  | $5{ }^{\prime}$ |  | 18" |  |  |  | $8{ }^{\prime}$ |  |  |  | $5{ }^{\prime}$ |  |
| LAB SAMPLE ID: |  |  | 710202-001 |  | 711019-001 |  | ECBC |  | ECBC |  | ECBC |  | ECBC |  |
| ECBC Parameters |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| Mustard | ug/kg | 550 | 10 | U | 10 | U | 10 | U | 10 | U | 10 | U | 10 | U |
| Lewisite | ug/kg |  | 100 | U | 100 | U | 100 | U | 100 | U | 100 | U | 100 | U |
| 1,4-Dithiane | ug/kg | 61,000 | 100 | U | 100 | U | 100 | U | 100 | U | 100 | U | 100 | $\cup$ |
| 1,4-Oxathiane | ug/kg | 61,000 | 100 | U | 100 | U | 100 | U | 100 | U | 100 | U | 100 | U |
| QA NOTES AND DATA QUALIFIERS: |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| Comparison value based on the higher of the Sept 2008 Regional Screen Level (RSL) (if non-carcinogenic) or the 2007 Background value. (NO CODE) - Confirmed identification. NA - Not available. NS - Not Sampled. <br> U - Analyte was analyzed for but not detected above the adjusted practical quantitation limit (PQL). <br> UJ - Analyte not detected, reported PQL may be inaccurate or imprecise. <br> J - Analyte detected, estimated concentration. <br> UL - Analyte not detected, reported PQL is biased low, actual PQL is expected to be higher. <br> L - Analyte detected, reported value is biased low, actual value is expected to be higher. <br> K - Analyte detected, reported value is biased high, actual value is expected to be lower. <br> + - Result reported from diluted sample. <br> Detections are bolded. <br> Detections exceeding the comparison level are shown shaded and bolded. <br> NS - Not Sampled BG - Background Value (2007 Study). SV - Spring Valley Remediation Goal. <br> * RSL for methyl mercury since methyl mercury is on the AUES list. |  |  |  |  |  |  |  |  |  |  |  |  |  |  |

Table A.9--PARSONS 2007

## VALIDATED ANALYTICAL RESULTS FOR TEST PIT GRAB SOIL SAMPLES (NON-SPRING VALLEY ANALYTES DETECTED)

| SAMPLE ID: |  |  | $\begin{aligned} & \text { SW-4835GB-01 } \\ & \text { (assoc wl TP-17) } \end{aligned}$ |  | SW-4835GB-04 (assoc wl TP-40) |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| DATE SAMPLED: |  |  | 10/22/07 |  | 10/31/07 |  |
| SAMPLE DEPTH (FT) |  |  | $5 '$ |  | 18" |  |
| LAB SAMPLE ID: |  |  | 710202-001 |  | 711019-001 |  |
|  | Units | Comparison Level |  |  |  |  |
| Non SV Specific Volatiles - OLM04.3 |  |  |  |  |  |  |
| Acetone | ug/kg | 6,100,000 | NS |  | 45 |  |
| Methylene Chloride | ug/kg | 11,000 | NS |  | 10 | B |
|  |  |  |  |  |  |  |
| Non SV Specific Volatile TICs |  |  |  |  |  |  |
| 1-Methyl-4-(1-methylethyl) benzene | ug/kg | NA |  |  | 3.99 | NJ |
|  |  |  |  |  |  |  |
| Non SV Specific Semivolatiles - OLM04.3 |  |  |  |  |  |  |
| Anthracene | ug/kg | 1,700,000 | NS |  | 52 | J |
| Benzo(a)anthracene | ug/kg | 150 | NS |  | 110 | J |
| Benzo(a)pyrene | ug/kg | 15 | NS |  | 83 | J |
| Benzo(b)fluoranthene | ug/kg | 150 | NS |  | 72 | J |
| Benzo(k)fluoranthene | ug/kg | 1,500 | NS |  | 92 | J |
| Bis(2-ethylhexyl)phthalate | ug/kg | 35,000 | NS |  | 67 | J |
| Chrysene | ug/kg | 15,000 | NS |  | 100 | J |
| Fluoranthene | ug/kg | 230,000 | NS |  | 230 | J |
| Phenanthrene | $\mathrm{ug} / \mathrm{kg}$ | NA | NS |  | 220 | J |
| Pyrene | ug/kg | 170,000 | NS |  | 240 | J |
|  |  |  |  |  |  |  |
| Non SV Specific Semivolatile TICs |  |  |  |  |  |  |
| Unknown (06.73) | ug/kg | NA |  |  | 100 | NJ |
| (+)-Cycloisosativene (15.29) | ug/kg | NA |  |  | 560 | NJ |
| Unknown (15.44) | ug/kg | NA |  |  | 540 | NJ |
| E-11, 13-Tetradecadien-1-ol (16.04) | ug/kg | NA |  |  | 140 | NJ |
| Unknown (16.61) | ug/kg | NA |  |  | 240 | NJ |
| Naphthalene, 1,2,3,4-tetrahydro-1,6-dimethyl-4-(1-methylethyl)- | ug/kg | NA |  |  | 240 | NJ |
| Unknown (18.01) | ug/kg | NA |  |  | 130 | NJ |
| Unknown (20.09) | ug/kg | NA |  |  | 150 | NJ |
| Unknown (20.36) | ug/kg | NA |  |  | 2600 | NJ |
| Unknown (29.96) | ug/kg | NA |  |  | 300 | NJ |
|  |  |  |  |  |  |  |
| Non SV Specific Metals - ILM05.3 |  |  |  |  |  |  |
| Iron | $\mathrm{mg} / \mathrm{kg}$ | 32,400 | 45,500 | + | 43,300 |  |
| Magnesium | $\mathrm{mg} / \mathrm{kg}$ | 6,950 | 8,840 | + | 5,730 |  |
|  |  |  |  |  |  |  |
| QA NOTES AND DATA QUALIFIERS: |  |  |  |  |  |  |
| Comparison value based on the higher of the adjusted Sept 2008 Regional Screening Level (RSL) (if non-carcinogenic) or the 2007 Background value. <br> (NO CODE) - Confirmed identification. NA - Not available. NS - Not Sampled. <br> J - Analyte detected, estimated concentration. <br> B - Blank contamination, the analyte was detected in the associated blank at a comparable concentration. <br> NJ - Tentatively idenfified compound (TIC). Presumptively present at approximate concentration. <br> + - Result reported from diluted sample. <br> Detections are bolded. <br> Detections exceeding the comparison level are shown shaded and bolded. |  |  |  |  |  |  |

Table A.10--PARSONS 2008
VALIDATED ANALYTICAL RESULTS FOR HIGH As GRAB SOIL SAMPLES (SPRING VALLEY COMPREHENSIVE LIST)

| SAMPLE ID: |  |  | $\begin{gathered} \text { 4835GB(-190,50) } \\ \text { SW-N(5)LW-4.5 } \end{gathered}$ |  | $\begin{aligned} & \text { 4835GB(-190,50) } \\ & \text { SW-N(5)LW-5 } \end{aligned}$ |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| DATE SAMPLED: |  |  |  |  |  |  |
| SAMPLE DEPTH (FT) |  |  | 4.5 |  | 5.0 |  |
|  | Units | Comparison Level |  |  |  |  |
| Volatile Oganic Compounds - OLM04.3_V |  |  |  |  |  |  |
| 1,1,1-Trichloroethane | ug/kg | 900,000 | 13 | UJ | 12 | U |
| 1,1,2,2-Tetrachloroethane | ug/kg | 590 | 13 | UJ | 12 | U |
| 1,1,2-Trichloro-1,2,2-trifluoroethane | ug/kg | 230,000,000 | 13 | UJ | 12 | U |
| 1,1,2-Trichloroethane | ug/kg | 1,100 | 13 | UJ | 12 | U |
| 1,1-Dichloroethane | ug/kg | 3,400 | 13 | UJ | 12 | U |
| 1,1-Dichloroethene | ug/kg | 25,000 | 13 | UJ | 12 | U |
| 1,2,4-Trichlorobenzene | ug/kg | 8,700 | 13 | UJ | 12 | U |
| 1,2-Dibromo-3-Chloropropane | ug/kg | 6 | 13 | UJ | 12 | U |
| 1,2-Dibromoethane | ug/kg | NA | 13 | UJ | 12 | U |
| 1,2-Dichlorobenzene | ug/kg | 200,000 | 13 | UJ | 12 | U |
| 1,2-Dichloroethane | ug/kg | 450 | 13 | UJ | 12 | U |
| 1,2-Dichloropropane | ug/kg | 9,300 | 13 | UJ | 12 | U |
| 1,3-Dichlorobenzene | ug/kg | 230,000 | 13 | UJ | 1.5 | J |
| 1,4-Dichlorobenzene | ug/kg | 2,600 | 13 | UJ | 1.6 | J |
| 2-Butanone | ug/kg | 2,800,000 | 13 | UJ | 12 | U |
| 2-Hexanone | ug/kg | NA | 13 | UJ | 12 | U |
| 4-Methyl-2-Pentanone | ug/kg | NA | 13 | UJ | 12 | U |
| Acetone | ug/kg | 6,100,000 | 30 | J | 12 | U |
| Acetonitrile | ug/kg | 87,000 | 130 | UJ | 120 | U |
| Acrolein | ug/kg | 16 | 63 | UJ | 61 | U |
| Benzene | ug/kg | 1,100 | 13 | UJ | 12 | U |
| Benzyl Bromide | ug/kg | 156,000 | 13 | UJ | 12 | U |
| Benzyl Chloride | ug/kg | 3,800 | 13 | UJ | 12 | U |
| Bromodichloromethane | ug/kg | 10,000 | 13 | UJ | 12 | U |
| Bromoform | ug/kg | 61,000 | 13 | UJ | 12 | U |
| Bromomethane | ug/kg | 790 | 13 | UJ | 12 | U |
| Carbon Disulfide | ug/kg | 67,000 | 13 | UJ | 12 | U |
| Carbon Tetrachloride | ug/kg | 250 | 13 | UJ | 12 | U |
| Chlorobenzene | ug/kg | 31,000 | 13 | UJ | 12 | U |
| Chloroethane | ug/kg | 1,500,000 | 13 | UJ | 12 | U |
| Chloroform | ug/kg | 300 | 13 | UJ | 12 | U |
| Chloromethane | ug/kg | NA | 13 | UJ | 12 | U |
| Chloropicrin | ug/kg | NA | 63 | UJ | 61 | U |
| cis-1,2-Dichloroethene | ug/kg | 78,000 | 13 | UJ | 12 | U |
| cis-1,3-Dichloropropene | ug/kg | 1,700 | 13 | UJ | 12 | U |
| Cyclohexane | ug/kg | NA | 13 | UJ | 12 | U |
| Dibromochloromethane | ug/kg | 5,800 | 13 | UJ | 12 | U |
| Dichlorodifluoromethane | ug/kg | 19,000 | 13 | UJ | 12 | U |
| Diphenyl Ether | ug/kg | NA | 13 | UJ | 12 | U |
| Ethylbenzene | ug/kg | 5,700 | 13 | UJ | 12 | U |
| Isopropylbenzene | ug/kg | 220,000 | 13 | UJ | 12 | U |
| Methyl Acetate | ug/kg | 7,800,000 | 13 | UJ | 12 | U |
| Methyl Tert-Butyl Ether | ug/kg | 39,000 | 13 | UJ | 12 | U |
| MethylCyclohexane | ug/kg | NA | 13 | UJ | 12 | U |
| Methylene Chloride | ug/kg | 11,000 | 6 | J | 14 |  |
| Styrene | ug/kg | 650,000 | 13 | UJ | 12 | U |
| Tetrachloroethene | ug/kg | 570 | 13 | UJ | 12 | U |
| Toluene | ug/kg | 500,000 | 13 | UJ | 12 | U |
| trans-1,2-dichloroethene | ug/kg | 1,100 | 13 | UJ | 12 | U |
| trans-1,3-dichloropropene | ug/kg | 1,700 | 13 | UJ | 12 | U |
| Trichloroethene | ug/kg | 2,800 | 13 | UJ | 12 | U |
| Trichlorofluoromethane | ug/kg | 80,000 | 13 | UJ | 12 | U |
| Vinyl Chloride | ug/kg | 60 | 13 | UJ | 12 | U |
| Xylenes (Total) | ug/kg | 450,000 | 13 | UJ | 2.7 | J |

Table A.10--PARSONS 2008
VALIDATED ANALYTICAL RESULTS FOR HIGH As GRAB SOIL SAMPLES (SPRING VALLEY COMPREHENSIVE LIST)

| SAMPLE ID: |  |  | $\begin{aligned} & \text { 4835GB(-190,50) } \\ & \text { SW-N(5)LW-4.5 } \end{aligned}$ |  | $\begin{aligned} & \text { 4835GB(-190,50) } \\ & \text { SW-N(5)LW-5 } \end{aligned}$ |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| DATE SAMPLED: |  |  | 04/0 |  | 04/ |  |
| SAMPLE DEPTH (FT) |  |  | 4.5 |  | 5.0 |  |
|  | Units | Comparison Level |  |  |  |  |
| Semivolatile Organic Compounds - OLM04.3_SV |  |  |  |  |  |  |
| 1,1'-Biphenyl | ug/kg | 390,000 | 420 | U | 400 | U |
| 1-chloro-2,4-dinitrobenzene | ug/kg | NA | 420 | U | 400 | U |
| 2,2'-Oxybis(1-chloropropane) | ug/kg | NA | 420 | U | 400 | U |
| 2,4,5-Trichlorophenol | ug/kg | 610,000 | 1100 | U | 1000 | U |
| 2,4,6-Trichlorophenol | ug/kg | 44,000 | 420 | U | 400 | U |
| 2,4-Dichlorophenol | ug/kg | 18,000 | 420 | U | 400 | U |
| 2,4-Dimethylphenol | ug/kg | 120,000 | 420 | U | 400 | U |
| 2,4-Dinitrophenol | ug/kg | 12,000 | 1100 | U | 1000 | U |
| 2-Bromo-4'-chloroacetophenone | ug/kg | NA | 420 | U | 400 | U |
| 2-Chloronaphthalene | ug/kg | 630,000 | 420 | U | 400 | U |
| 2-Chlorophenol | ug/kg | 39,000 | 420 | U | 400 | U |
| 2-Methylnaphthalene | ug/kg | 31,000 | 420 | U | 400 | U |
| 2-methylphenol | ug/kg | 310,000 | 420 | U | 400 | U |
| 2-Nitroaniline | ug/kg | 23,000 | 1100 | U | 1000 | U |
| 2-Nitrophenol | ug/kg | NA | 420 | U | 400 | U |
| 3,3'-Dichlorobenzidine | ug/kg | 1,100 | 420 | U | 400 | U |
| 3-Nitroaniline | ug/kg | 1,800 | 1100 | U | 1000 | U |
| 4,6-Dinitro-2-Methylphenol | ug/kg | 610 | 1100 | U | 1000 | U |
| 4-Bromophenyl-phenylether | ug/kg | NA | 420 | U | 400 | U |
| 4-chloro-3-methylphenol | ug/kg | NA | 420 | U | 400 | U |
| 4-Chloroacetophenone | ug/kg | NA | 420 | U | 400 | U |
| 4-Chloroaniline | ug/kg | 9,000 | 420 | U | 400 | U |
| 4-Chlorophenyl-PhenylEther | ug/kg | NA | 420 | U | 400 | U |
| 4-methylphenol | ug/kg | 31,000 | 420 | U | 400 | U |
| 4-Nitroaniline | ug/kg | 23,000 | 1100 | U | 1000 | U |
| 4-Nitrophenol | ug/kg | 63,000 | 1100 | U | 1000 | U |
| Acenaphthene | ug/kg | 340,000 | 420 | U | 400 | U |
| Acenaphthylene | ug/kg | 470,000 | 420 | U | 400 | U |
| Acetophenone | ug/kg | 780,000 | 420 | U | 400 | U |
| Anthracene | ug/kg | 1,700,000 | 420 | U | 400 | U |
| Atrazine | ug/kg | 2,100 | 420 | U | 400 | U |
| Benzal Chloride | ug/kg | NA | 420 | U | 400 | U |
| Benzaldehyde | ug/kg | 780,000 | 420 | U | 400 | U |
| Benzo(a)anthracene | ug/kg | 150 | 420 | U | 400 | U |
| Benzo(a)pyrene | ug/kg | 15 | 420 | U | 400 | U |
| Benzo(b)fluoranthene | ug/kg | 150 | 420 | U | 400 | U |
| Benzo(g,h,i)perylene | ug/kg | NA | 420 | U | 400 | U |
| Benzo(k)fluoranthene | ug/kg | 1,500 | 420 | U | 400 | U |
| Benzoic Acid | ug/kg | 24,000,000 | 420 | U | 400 | U |
| Bis(2-chloroethoxy)methane | ug/kg | NA | 420 | U | 400 | U |
| Bis(2-chloroethyl)ether | ug/kg | 190 | 420 | U | 400 | U |
| Bis(2-ethylhexyl)phthalate | ug/kg | 35,000 | 52 | J | 400 | U |
| Bromoacetophenone | ug/kg | NA | 420 | U | 400 | U |
| Butylbenzylphthalate | ug/kg | 260,000 | 420 | U | 400 | U |
| Caprolactam | ug/kg | 3,100,000 | 420 | U | 400 | U |
| Carbazole | ug/kg | 32,000 | 420 | U | 400 | U |
| Chrysene | ug/kg | 15,000 | 420 | U | 400 | U |
| Dibenz(a,h)Anthracene | ug/kg | 15 | 420 | U | 400 | U |
| Dibenzofuran | ug/kg | 7,800 | 420 | U | 400 | U |
| Diethylphthalate | ug/kg | 4,900,000 | 420 | U | 400 | U |

Table A.10--PARSONS 2008
VALIDATED ANALYTICAL RESULTS FOR HIGH As GRAB SOIL SAMPLES (SPRING VALLEY COMPREHENSIVE LIST)

| SAMPLE ID: |  |  | $\begin{aligned} & \text { 4835GB(-190,50) } \\ & \text { SW-N(5)LW-4.5 } \end{aligned}$ |  | $\begin{aligned} & \text { 4835GB(-190,50) } \\ & \text { SW-N(5)LW-5 } \end{aligned}$ |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| DATE SAMPLED: |  |  | 04/07/08 |  | 04/16/08 |  |
| SAMPLE DEPTH (FT) |  |  | 4.5 |  | 5.0 |  |
|  | Units | Comparison Level |  |  |  |  |
| Dimethylaniline | ug/kg | 16,000 | 420 | U | 400 | U |
| Dimethylphthalate | ug/kg | 78,000,000 | 420 | U | 400 | U |
| di-n-Butyl Phthalate | ug/kg | 610,000 | 46 | J | 79 | J |
| di-n-Octyl Phthalate | ug/kg | 310,000 | 420 | U | 400 | U |
| Ethylene Chlorohydrin | ug/kg | NA | 420 | U | 400 | U |
| Fluoranthene | ug/kg | 230,000 | 420 | U | 400 | U |
| Fluorene | ug/kg | 230,000 | 420 | U | 400 | U |
| Glycol-bromohydrin | ug/kg | NA | 420 | U | 400 | U |
| Hexachlorobenzene | ug/kg | 300 | 420 | U | 400 | U |
| Hexachlorobutadiene | ug/kg | 6,200 | 420 | U | 400 | U |
| Hexachlorocyclopentadiene | ug/kg | 37,000 | 420 | U | 400 | U |
| Hexachloroethane | ug/kg | 7,800 | 420 | U | 400 | U |
| Indeno(1,2,3-cd)pyrene | ug/kg | 150 | 420 | U | 400 | U |
| Isophorone | ug/kg | 510,000 | 420 | U | 400 | U |
| Naphthalene | ug/kg | 3,900 | 420 | U | 400 | U |
| N-nitroso-di-n-propylamine | ug/kg | NA | 420 | U | 400 | U |
| n -Nitrosodiphenylamine | ug/kg | 99,000 | 420 | U | 400 | U |
| Pentachlorophenol | ug/kg | 3,000 | 1100 | U | 1000 | U |
| Phenanthrene | ug/kg | NA | 420 | U | 400 | U |
| Phenol | ug/kg | 1,800,000 | 420 | U | 400 | U |
| Phenyl isocyanate | ug/kg | NA | 420 | U | 400 | U |
| Phenyl isothiocyanate | ug/kg | NA | 420 | U | 400 | U |
| Pyrene | ug/kg | 170,000 | 420 | U | 400 | U |
| Tolidine | ug/kg | NA | 420 | U | 400 | U |
|  |  |  |  |  |  |  |
| Non SV Specific Semivolatile TICs |  |  |  |  |  |  |
| Unknown (10:48) | ug/kg | NA |  |  | 2.6 | NJ |
| 2-Ethyl Hexanoic acid (11:49) | ug/kg | NA | 140 | NJ |  |  |
| Cyclotetradecane (19:03) | ug/kg | NA | 990 | NJ |  |  |
| Unknown (12:03) | ug/kg | NA | 110 | NJ |  |  |
|  |  |  |  |  |  |  |
| Explosives -SW8330A |  |  |  |  |  |  |
| 1,3,5-Trinitrobenzene | ug/kg | 220,000 | 40 | U | 40 | U |
| 1,3-Dinitrobenzene | ug/kg | 610 | 40 | U | 40 | U |
| 2,4,6-Trinitrotoluene (TNT) | ug/kg | 19,000 | 40 | U | 40 | U |
| 2,4-Dinitrotoluene | ug/kg | 12,000 | 40 | U | 40 | U |
| 2,6-Dinitrotoluene | ug/kg | 6,100 | 40 | U | 40 | U |
| 2-Amino-4,6-Dinitrotoluene | ug/kg | 710 | 40 | U | 40 | U |
| 2-NITROTOLUENE | ug/kg | 2,900 | 80 | U | 80 | U |
| 3-Nitrotoluene | ug/kg | 120,000 | 80 | U | 80 | U |
| 4-Amino-2,6-Dinitrotoluene | ug/kg | 710 | 40 | U | 40 | U |
| 4-Nitrotoluene | ug/kg | 30,000 | 80 | U | 80 | U |
| Hexahydro-1,3,5-trinitro-1,3,5-triazine (RDX) | ug/kg | 5,500 | 80 | U | 80 | U |
| Methyl-2,4,6-trinitrophenylnitramine (Tetryl) | ug/kg | 24,000 | 80 | U | 80 | U |
| Nitrobenzene | ug/kg | NA | 40 | U | 40 | U |
| Octahydro-1,3,5,7-tetranitro-1,3,5,7-tetrazocine (HMX) | ug/kg | 380,000 | 80 | U | 80 | U |
| Nitroglycerine | ug/kg | 610 | 4000 | U | 4000 | U |
|  |  |  |  |  |  |  |
| Agent Breakdown Products |  |  |  |  |  |  |
| 1,4-Dithiane | ug/kg | 61,000 | 13 | U | 12 | U |
| 1,4-Oxathiane | ug/kg | 61,000 | 25 | UJ | 24 | U |
| Thiodiglycol | ug/kg | 39,100 | 630 | U | 610 | U |

Table A.10--PARSONS 2008
VALIDATED ANALYTICAL RESULTS FOR HIGH As GRAB SOIL SAMPLES (SPRING VALLEY COMPREHENSIVE LIST)

| SAMPLE ID: |  |  | $\begin{aligned} & \text { 4835GB(-190,50) } \\ & \text { SW-N(5)LW-4.5 } \end{aligned}$ |  | $\begin{gathered} \text { 4835GB(-190,50) } \\ \text { SW-N(5)LW-5 } \end{gathered}$ |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| DATE SAMPLED: |  |  | 04/0 |  | 04/1 |  |
| SAMPLE DEPTH (FT) |  |  | 4.5 |  | 5.0 |  |
|  | Units | Comparison Level |  |  |  |  |
| Metals - ILM05.3 |  |  |  |  |  |  |
| Aluminum | $\mathrm{mg} / \mathrm{kg}$ | 19,100 | 15600 | + | 10500 |  |
| Antimony | $\mathrm{mg} / \mathrm{kg}$ | 5.2(BG) | 3.9 | J+ | 5.5 | U |
| Arsenic | $\mathrm{mg} / \mathrm{kg}$ | 20(SV) | 281 | + | 4.1 |  |
| Barium | $\mathrm{mg} / \mathrm{kg}$ | 1,500 | 96.2 | + | 54.5 |  |
| Beryllium | $\mathrm{mg} / \mathrm{kg}$ | 16 | 1.2 | J | 0.73 |  |
| Cadmium | $\mathrm{mg} / \mathrm{kg}$ | 7 | 2.4 | U+ | 0.22 | B |
| Chromium | $\mathrm{mg} / \mathrm{kg}$ | 12000 | 269 | J+ | 448 | J |
| Cobalt | $\mathrm{mg} / \mathrm{kg}$ | 17.8 | 232 | J+ | 23.6 | J |
| Copper | $\mathrm{mg} / \mathrm{kg}$ | 310 | 91.7 | J+ | 16.2 | J |
| Iron | $\mathrm{mg} / \mathrm{kg}$ | 32,400 | 50500 | J+ | 32900 | J |
| Lead | $\mathrm{mg} / \mathrm{kg}$ | 400 | 22.4 | + | 7.6 |  |
| Magnesium | $\mathrm{mg} / \mathrm{kg}$ | 6,950 | 6640 | + | 3300 |  |
| Manganese | $\mathrm{mg} / \mathrm{kg}$ | 1,800 | 3920 | J+ | 705 | J |
| Mercury | $\mathrm{mg} / \mathrm{kg}$ | 0.78* | 0.22 |  | 0.49 |  |
| Nickel | $\mathrm{mg} / \mathrm{kg}$ | 160 | 64 |  | 43.2 |  |
| Selenium | $\mathrm{mg} / \mathrm{kg}$ | 39 | 16.5 | U+ | 0.59 | J |
| Silver | $\mathrm{mg} / \mathrm{kg}$ | 39 | 0.95 | U | 0.91 | U |
| Strontium | $\mathrm{mg} / \mathrm{kg}$ | 4,700 | 14.7 | J+ | 17.5 | J+ |
| Tellurium | $\mathrm{mg} / \mathrm{kg}$ | 39.11 | 1.8 | J+ | 2.2 | J+ |
| Thallium | $\mathrm{mg} / \mathrm{kg}$ | 2.2 | 11.8 | U+ | 2.3 | U |
| Tin | $\mathrm{mg} / \mathrm{kg}$ | 4,700 | 23.6 | U+ | 1.4 | B |
| Titanium | $\mathrm{mg} / \mathrm{kg}$ | 31,000 | 550 | + | 325 |  |
| Vanadium | $\mathrm{mg} / \mathrm{kg}$ | 550 | 80.1 | J+ | 68.2 |  |
| Zinc | $\mathrm{mg} / \mathrm{kg}$ | 2,300 | 75.2 | + | 34.5 |  |
| Zirconium | $\mathrm{mg} / \mathrm{kg}$ | 48.3 | 12.2 | B+ | 16.9 | B+ |
|  |  |  |  |  |  |  |
| Other Parameters |  |  |  |  |  |  |
| Fluoride | $\mathrm{mg} / \mathrm{kg}$ | 470 |  |  | 11 |  |
| Total Cyanide | $\mathrm{mg} / \mathrm{kg}$ | 160 | 0.41 | J | 0.17 | U |
| Iodine (as lodide) | $\mathrm{mg} / \mathrm{kg}$ | NA |  |  | 0.02 | U |
| Iodine Pentafluoride (as lodate) | $\mathrm{mg} / \mathrm{kg}$ | NA |  |  | 110 | K |
| Perchlorate | ug/kg | 5,500 | 1.32 | J | 1.74 | J |
|  |  |  |  |  |  |  |
| NOTES: |  |  |  |  |  |  |
| Comparison value based on the higher of the adjusted Sept 2008 Regional Screening Level (RSL) (if non-carcinogenic) or the 2007 Background value. |  |  |  |  |  |  |
| * RSL for methyl mercury since methyl mercury is on the AUES list. |  |  |  |  |  |  |
| BG - Background value (2007 Study). SV - Spring Valley Remediation Goal. |  |  |  |  |  |  |

## APPENDIX B SAMPLE SIZE ESTIMATES

## APPENDIX B - SAMPLE SIZE ESTIMATES

For this site, one of the data quality objectives (DQOs) is that a sufficient number of soil samples are collected to detect a minimum of a $20 \%$ difference from the action level with $95 \%$ confidence and $80 \%$ power. This section presents the methodology used to estimate the required sample sizes and the results of the calculations.

## Equations

Data from the site may be compared to an action level using a one-sample t-test for normally distributed data and a Mann-Whitney $U$ test for data that is not normally distributed (USEPA 2002, 2006, and 2009). For data that is lognormally distributed, the data may be log-transformed and a one-sample t-test may be used on the log-transformed data. For data that contains nondetects (regardless of distribution), current USEPA (2009) guidance recommends the use of nonparametric tests. Therefore, the one-sample Mann-Whitney $U$ test was used here to compare data containing non-detects from the site to the action levels.
For data that is normally distributed without non-detects, USEPA (2000 and 2006) and PNNL (2007) provide an equation for estimating the sample size required for a one-sample t-test, as follows:

$$
n=\frac{S D^{2}\left(Z_{1-\alpha}+Z_{1-\beta}\right)^{2}}{\Delta^{2}}+\frac{Z_{1-\alpha}^{2}}{2}
$$

where:

| n | $=$ recommended minimum sample size |
| :--- | :--- |
| SD | $=\quad$arithmetic standard deviation; for lognormally distributed data, this is the <br> back-transformed SD of the log-transformed data |
| $\Delta$ | $=$ minimum detectable difference from the action level |
| Z | $=$ value from the Z-distribution |
| $\alpha$ | $=$ the false rejection (Type I) error rate (0.05) |
| $\beta$ | $=$ the false acceptance (Type II) error rate (0.20) |

For data that is not normally distributed or contains non-detects, USEPA (2006) and PNNL (2009) provide an equation for estimating the sample size required for a one-sample MannWhitney U test, as follows:

$$
n=1.16 \times\left(\frac{2 S D^{2}\left(Z_{1-\alpha}+Z_{1-\beta}\right)^{2}}{\Delta^{2}}+\frac{Z_{1-\alpha}^{2}}{4}\right)
$$

where:
$\mathrm{n} \quad=\quad$ recommended minimum sample size
$\mathrm{SD}=$ arithmetic standard deviation
$\Delta \quad=\quad$ minimum detectable difference from the action level
$\mathrm{Z} \quad=\quad$ value from the Z -distribution

```
\alpha = the false rejection (Type I) error rate (0.05)
\beta = the false acceptance (Type II) error rate (0.20)
```


## Existing Data

The equations presented above require the standard deviation of the data at the site. To determine the standard deviation, all data collected at the site were evaluated for use. The results of the data quality assessment (USACE, 2009) indicate that all of the data was usable. However, not all of the data is representative of current conditions at the site, as there have been several remedial actions. Therefore, those samples that were collected in areas that have been excavated were excluded. For metals, only samples collected by Parsons were used in this assessment. However, for non-metals, all samples that were collected at the site were used. A complete list of the samples included in this analysis is presented in Table B.1.

## Inputs for Metals

The data from the samples listed in Table B. 1 were used to calculate the arithmetic standard deviation for each metal. For lognormally distributed data without non-detects, the arithmetic standard deviation was calculated for the $\log _{10}$-transformed data and then back-transformed into normal space. For those metals with non-detects, the Kaplan-Meier standard deviation was calculated using ProUCL v4.00.04 (USEPA 2009). The data distributions were also determined using ProUCL. The output from ProUCL is shown in Table B.2. Table B. 3 shows the distributions of the detected data, the number of data points used in the calculations, the number of non-detects, and the standard deviations used in the calculations. However, the type of standard deviation used depends on the data distribution, as noted in Table B.3.

To achieve a power of $80 \%, \beta$ was set in the equations above to $20 \%$, as power $=1-\beta$. To achieve a confidence of $95 \%, \alpha$ was set in the equations above to $5 \%$, as confidence $=1-\alpha$.

The minimum detectable difference ( $\Delta$ ) used in the equation was set at $20 \%$ of the action level; i.e., 0.2 times the action level. Therefore, the sample size calculated here is the minimum necessary to show a significant difference between the site mean and the action level using a one-sample t-test or a one-sample Mann-Whitney U test, as appropriate. The action levels are shown in Table B.3. Note that the action level is the greater of the background concentration and the risk-based preliminary remedial goal (see Table B.4). All calculations were performed using Visual Sample Plan v5.4.2 (PNNL 2007).

## Results

As can be seen from Table B.3, the minimum number of samples necessary to detect a minimum of a $20 \%$ difference from the action level with $95 \%$ confidence and $80 \%$ power for all metals is less than the number of samples already collected. Thus, the sample size DQO was met.

## Analysis of Non-Metals

During the course of the many investigations conducted at this property, numerous analytes were evaluated, including the full suites of explosives, PCBs, pesticides, SVOCs, and VOCs, as well as chemical agents and their breakdown products.

Explosives and PCBs were not detected in any of the samples collected at the site. Therefore, sample sizes were not evaluated for explosives and PCBs.

For pesticides, only 4,4'-DDT, alpha-chlordane, gamma-chlordane, heptachlor-epoxide and 2,4,5-TP (silvex) were detected in one or two samples of the thirteen samples analyzed. However, sample sizes were not calculated since the detection limits were only available for one of the NDs. A review of the history of the chlorinated pesticides shows that the first chlorinated pesticide produced was DDT. Although discovered in 1874, it was not known that DDT could be used as an insecticide until 1939 (ATSDR 2002, WHO 1979). Further, the first DDT samples were only sent to the United States in 1942 (WHO 1979). Chlorinated pesticides, therefore, largely came into use in the 1940s. For example, heptachlor was not invented until 1946 (WHO 1984a), chlordane until the 1940s (WHO 1984b), and aldrin and dieldrin were first synthesized in 1948 (WHO 1989). Prior to that, metals (such as lead arsenate) were used as pesticides. Thus, any chlorinated pesticides present at the site are very unlikely to be due to military activities (which ceased in 1920s) and sampling at the site for chlorinated pesticides is not necessary.
Among the SVOCs, several polycyclic aromatic hydrocarbons (PAHs) were detected. Similar to metals, PAHs are ubiquitous in the environment. Therefore, the action levels for the PAHs are the greater of background or the risk-based concentrations (Table B.4). The sample sizes necessary to meet the DQOs for the PAHs were evaluated using the data obtained at the site that have not been removed following the same process outlined above. However, there were too many non-detects to calculate the Kaplan-Meier standard deviation in ProUCL. Therefore, the arithmetic standard deviation was estimated using $1 / 2$ the detection limit. The results of the sample size estimates (Table B.5) for the PAHs indicate that additional sampling may be warranted. It should be noted that these calculations are based on data that consists mostly of NDs, some of which have elevated detection limits. Thus, these sample size estimates contain a high degree of uncertainty. Nonetheless, if additional sampling is conducted for PAHs at the site, it is recommended that either EPA Method 8310 or 8270 C SIM be used and that the previously collected data be excluded from the analyses (due to the elevated detection limits).
Apex (1996) identified a layer of broken glassware and debris at approximately 2 ft bgs. To remove the contamination, Apex (1996) excavated a pit 12 feet in diameter and 6 ft deep. Removal was confirmed by five confirmation samples analyzed for metals, VOCs, pesticides, and herbicides. However, hexachlorobenzene was detected in the excavation at a concentration exceeding the action level and the confirmation samples were not analyzed for hexachlorobenzene (i.e., EPA Method 8270C). Thus, additional sampling for SVOCs (such as hexachlorobenzene) may be warranted at the site.

Outside of the excavation, Apex (1996) advanced 127 borings and dug 3 test pits. Parsons also dug 76 test pits at the site. No other potential source areas (e.g., burn pits, ash layers, glassware, debris, NAPLs, stained soils, etc.) were identified in the borings or test pits at the site. PID readings were taken by Apex (1996) from 91 soil probes in the backyard and 24 in the front yard. Where elevated PID readings were found, soil samples were collected and sent to a fixedlaboratory for analysis for VOCs. No elevated levels of VOCs were found in any of the soil samples collected where there were elevated PID readings (APEX 1996). From a qualitative site-characterization perspective, this indicates that there are no potential VOC sources at the site (aside from the area already excavated by Apex) and that there is little potential for VOC impacts at the site. Nonetheless, the sample sizes necessary to achieve the DQOs were calculated for two representative VOCs; i.e., 1,1,2,2-tetrachloroethane and 1,1,2-trichloroethane. These two VOCs were both detected in one of 20 samples. However, detection limits were available for only 17
of those samples; therefore, only the 17 data points for which detection limits were available were used. Since there were too few detects (i.e., 1 of 17) to calculate the Kaplan-Meier standard deviation in ProUCL, the arithmetic standard deviation was calculated by using $1 / 2$ the detection limit. The results of the sample size estimates (Table B.5) indicate that the minimum number of samples necessary to achieve the DQOs is less than the number of samples already collected.

Among the agent breakdown products, only thiodiglycol was detected. The results of the sample size estimates (Table B.5) indicate that the minimum number of samples necessary to achieve the DQOs for thiodiglycol is less than the number of samples already collected.
The results of the analyses presented above indicate that the sample sizes for metals, VOCs, and agent breakdown products are sufficient for the DQOs. Additionally, no further analyses at the site are recommended for 1) explosives, 2) PCBs, and 3) chlorinated pesticides. However, additional sampling at the site may be warranted for PAHs and select SVOCs, such as hexachlorobenzene. Based on previous investigations and test pit investigation results, all potential sources were identified and removed. The samples collected from the excavation performed by Apex (1996) or associated with AUES-related items found during test pit investigation represent the sources at the site. The detected SVOCs concentrations remaining at the site are below the Regional Screening Levels or site-specific background levels. Therefore, the site characterization for metals and non-metals (including VOCs, SVOCs, PCBs, pesticides, ABPs, and explosives) for 4835 Glenbrook Road is completed under current USEPA guidance.

## References

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World Health Organization (WHO). 1989. Environmental health criteria 91 - aldrin and dieldrin. Available online at: http://www.inchem.org/documents/ehc/ehc/ehc91.htm

Table B. 1
Unexcavated Samples Used to Estimate Sample Sizes 4835 Glenbrook Rd.

| Metals data (from Parsons only) | Non-Metals data |  |
| :---: | :---: | :---: |
| Sample ID | Sample ID | Collected By |
| SW-4835GB-(-170,10)SW-E(5) | 052692-1CM | EMS (1992) |
| SW-4835GB-(-170,10)SW-S | 9005 | Apex (1996) |
| SW-4835GB-(-170,10)SW-W | 9006 | Apex (1996) |
| SW-4835GB-(-190,10)-2 | 9007 | Apex (1996) |
| SW-4835GB-(-190,10)-N | 9008 | Apex (1996) |
| SW-4835GB-(-130,-30)-1.5 | 9009 | Apex (1996) |
| SW-4835GB-(-130,-30)SW-N | 9010 | Apex (1996) |
| SW-4835GB-(-130,-30)SW-W | 9011 | Apex (1996) |
| SW-4835GB-(-190,90)-2 | 9012 | Apex (1996) |
| SW-4835GB-(-190,90)SW-E(5) | 9013 | Apex (1996) |
| SW-4835GB-(-190,90)SW-S | 9014 | Apex (1996) |
| SW-4835GB-(-250,70)-2 | 9015 | Apex (1996) |
| SW-4835GB-(-250,70)SW-E | 9016 | Apex (1996) |
| SW-4835GB-(-250,70)SW-S | 9017 | Apex (1996) |
| SW-4835GB-(-150,50)-2 | 9018 | Apex (1996) |
| SW-4835GB-(-150,50)SW-E | 9019 | Apex (1996) |
| SW-4835GB-(-150,50)SW-N | G-01 | Apex (1996) |
| SW-4835GB-(-90,50)-2 | G-02 | Apex (1996) |
| SW-4835GB-(-90,50)SW-E | G-03 | Apex (1996) |
| SW-4835GB-(-90,30)SW-W(5) | OU3-SB02 | EPA (1999) |
| SW-4835GB-(-130,-30)SW-S(2.5) | OU3 MTL-4835-1 | Parsons |
| SW-4835GB-(190,90)SW-E(5)LC | OU3 MTL-4835-2 | Parsons |
| SW-4835GB-(-190,90)SW-E(5)LN | OU3 MTL-4835-3 | Parsons |
| SW-4835GB-(190,90)SW-E(5)LS | OU3 MTL-4835-4 | Parsons |
| SW-4835GB-(-190,90)SW-N(6) | OU3 MTL-4835-SB-(0-2) | Parsons |
| SW-4835GB-(-90,50)SW-N(5) | OU3 MTL-4835-SB-(2-4) | Parsons |
| SW-4835GB-(-170,10)-4 | OU3 MTL-4835-SB-(4-6) | Parsons |
| SW-4835GB-(-170,10)SW-E(5)LC-4 | SW-4835GB-01 (assoc w/TP-17) | Parsons |
| SW-4835GB-(-170,10)SW-E(5)LS | SW-4835GB-04 (assoc w/ TP-40) | Parsons |
| SW-4835GB-(-170,10)SW-S3.5 | SW-4835GB-02 (assoc w/ TP-40) | Parsons |
| SW-4835GB-(-170,10)-SW-W3.5 | SW-4835GB-TP56-001 (assoc w/ TP-56) | Parsons |
| SW-4835GB-(-190,10)SW-E(7) | SW-4835GB-TP49-001 (assoc w/ TP-49) | Parsons |
| SW-4835GB-(-190,90)SW-N(6)LC | SW-4835GB-16 (assoc w/ TP-49) | Parsons |
| SW-4835GB-(-190,90)SW-N(6)LE | 4835GB(-190,50) SW-N(5)LW-5 | Parsons |
| SW-4835GB-(-150,50)SW-S(8) |  |  |
| SW-4835GB-(-90,50)SW-N(5)LE |  |  |
| SW-4835GB(-90,50)-SW-N(5)LE2.5 |  |  |
| SW-4835GB(-90,50)-SW-N(5)LC-3 |  |  |
| SW-4835GB(-150,50)-SW-S(8)LE |  |  |
| SW-4835GB(-150,50)-SW-S(8)LC-3 |  |  |
| SW-4835GB(-190,10)SW-E(7)LN |  |  |
| SW-4835GB(-190,10)SW-E(7)LC |  |  |
| SW-4835GB(-150,50)SWS(8)2.5 |  |  |
| SW-4835GB(-150,50)SWS(8)LE2.5 |  |  |
| SW-4835GB(-90,50)SWN(5)2.5 |  |  |

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Table B. 1
Unexcavated Samples Used to Estimate Sample Sizes 4835 Glenbrook Rd.

```
Metals data (from Parsons only) Non-Metals data
SW-4835GB-(-170,50)
SW-4835GB-(-150,10)
SW-4835GB-(-150,30)-2
SW-4835GB-(-150,-10)SW-E
SW-4835GB-(-170,30)-4
SW-4835GB-(-150,-10)-2
SW-4835GB-(-190,50)-5
SW-4835GB-(-170,-10)-3
SW-4835GB-(-190,50)SW-N(5)
SW-4835GB-(-190,50)SW-S(5)
SW-4835GB-(-170,30)SW-E
SW-4835GB-(-170,30)SW-E-3.5
SW-4835GB-(-150,30)SW-E(5)LN
SW-4835GB-(-150,30)SW-W(5)
SW-4835GB-(-170,10)SW-N
4835GB-(-190,30)-5
4835GB-(-190,30)-SW-N(4.5)
4835GB-(-190,30)-SW-N
4835GB-(-170,30)SW-S(5)-3.5
4835GB-(-170,30)SW-S(5)LW
4835GB-(-150,30)SW-W(5)LC
SW-4835GB-(190,50)SW-S(5)LC
SW-4835GB-(170,30)SW-S(5)-LC5
SW-4835GB(-170,30)SW-S(5)LW4.5
SW-4835GB(-170,30)SW-S(5)-4.5
SW-4835GB(-150,10)SW-W(10)LC3
SW-4835GB(-150,-10)SW-W(10)LC4
4835GB(-150,-10)SW-W(10)LS-2.5
4835GB(-150,-10)SW-W(10)-2.5
4835GB(-190,50)-SW-N(5)-4.5
4835GB(-190,50)-SW-N(5)LC
4835GB(-190,50)SW-S(5)-4.5
4835GB(-170,-10)SW-S-3
SW-4835GB-(-190,50)SWN(5)LW(5)
SW-4835GB-(-190,50)SWN(5)LW(5)-4.5
SW-4835GB-(-190,50)SWN(5)LW(5)LN
SW-4835GB-(-190,50)SWN(5)LW(5)LN-4.5
SW-4835GB-(-190,50)SWN(5)LW(5)LE
SW-4835GB-(-190,50)SWN(5)LW(5)LE-4.5
SW-4835GB-(-170,10)SWN(5)-3.5
SW-4835(-170,10)SWN(5)LC5
SW-4835(-170,10)SWN(5)-4.5
SW-4835(-170,10)SWN(5)LE-4.5
SW-4835(-170,10)SWN(5)LW-4.5
(-170,30)SW-S(5)LE-4.5
(-150,30)SW-E(5)LN-2.5
```

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Table B. 1
Unexcavated Samples Used to Estimate Sample Sizes 4835 Glenbrook Rd.

```
Metals data (from Parsons only)
Non-Metals data
(-150,30)SW-E(5)LC-3.0
SW-4835GB(-90,30)-4
SW-4835GB-(90,30)-SW-W(15)-3.5
SW-4835GB-(90,30)-SW-W(15)-0.5
SW-4835GB-(90,30)-SW-W(15)LE-4.0
SW-4835GB-01
SW-4835GB-04
4835GB(-190,50)-SW-N(5)LW-5
OU3-MTL-4835(-100,0)
OU3-MTL-4835(-100,100)
OU3-MTL-4835(-100,120)
OU3-MTL-4835(-100,140)
OU3-MTL-4835(-100,20)
OU3-MTL-4835(-100,80)
OU3-MTL-4835(-120,100)
OU3-MTL-4835(-120,120)
OU3-MTL-4835(-120,140)
OU3-MTL-4835(-140,100)
OU3-MTL-4835(-140,120)
OU3-MTL-4835(-140,140)
OU3-MTL-4835(-160,100)
OU3-MTL-4835(-160,120)
OU3-MTL-4835(-160,140)
OU3-MTL-4835(-180,120)
OU3-MTL-4835(-180,140)
OU3-MTL-4835(-200,120)
OU3-MTL-4835(-200,140)
OU3-MTL-4835(-220,120)
OU3-MTL-4835(-220,140)
OU3-MTL-4835(-240,120)
OU3-MTL-4835(-240,140)
OU3-MTL-4835(-260,120)
OU3-MTL-4835(-260,140)
OU3-MTL-4835(-280,120)
OU3-MTL-4835(-320,0)
OU3-MTL-4835(-340,0)
OU3-MTL-4835(280,140)
OU3-MTL-4835-(-100,-20)
OU3-MTL-4835-(-100,-40)
OU3-MTL-4835-(-120,-20)
OU3-MTL-4835-(-120,-40)
OU3-MTL-4835-(-120,0)
OU3-MTL-4835-(-140,-40)
OU3-MTL-4835-(-140,0)
OU3-MTL-4835-(-160,80)
OU3-MTL-4835-(-180,100)
```

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Table B. 1
Unexcavated Samples Used to Estimate Sample Sizes 4835 Glenbrook Rd.

| Metals data (from Parsons only) | Non-Metals data |
| :--- | :--- |
| OU3-MTL-4835-(-180,60) |  |
| OU3-MTL-4835-(-180,80) |  |
| OU3-MTL-4835-(-200,60) |  |
| OU3-MTL-4835-(-200,80) |  |
| OU3-MTL-4835-(-220,100) |  |
| OU3-MTL-4835-(-220,40) |  |
| OU3-MTL-4835-(-220,60) |  |
| OU3-MTL-4835-(-220,80) |  |
| OU3-MTL-4835-(-240,100) |  |
| OU3-MTL-4835-(-240,60) |  |
| OU3-MTL-4835-(-240,80) |  |
| OU3-MTL-4835-(-260,100) |  |
| OU3-MTL-4835-(-280,100) |  |
| OU3-MTL-4835-(-300,0) |  |

Table B. 2

## ProUCL UCL Output for Metals Data from 4835 Glenbrook Rd

| User Selected Options |  |
| :--- | :---: |
| Full Precision | ON |
| Confidence Coefficient | $95 \%$ |
| Number of Bootstrap Operations | 2000 |


|  | Aluminum |  |
| :--- | :--- | ---: |
|  |  |  |
| General Statistics | 97 Number of Detected Data | 96 |
| Number of Valid Data | 82 Number of Non-Detect Data | 1 |
| Number of Distinct Detected Data | 2 Percent Non-Detects | $1.03 \%$ |
| Number of Missing Values | Log-transformed Statistics |  |
|  | 8960 Minimum Detected | 9.100526 |
| Raw Statistics | 55900 Maximum Detected | 10.93132 |
| Minimum Detected | 24176.67 Mean of Detected | 10.02508 |
| Maximum Detected | 9053.644 SD of Detected | 0.374439 |
| Mean of Detected | 18600 Minimum Non-Detect | 9.830917 |
| SD of Detected | 18600 Maximum Non-Detect | 9.830917 |
| Minimum Non-Detect |  |  |

UCL Statistics



| Arsenic |  |  |  |
| :---: | :---: | :---: | :---: |
| General Statistics |  |  |  |
| Number of Valid Observations | 151 | Number of Distinct Observations | 117 |
| Raw Statistics |  | Log-transformed Statistics |  |
| Minimum | 0.69 | Minimum of Log Data | -0.37106 |
| Maximum | 19.9 | Maximum of Log Data | 2.99072 |
| Mean | 9.274503 | Mean of log Data | 1.997531 |
| Median | 9.1 | SD of log Data | 0.761329 |
| SD | 5.340711 |  |  |
| Coefficient of Variation | 0.575849 |  |  |
| Skewness | 0.235048 |  |  |
| Relevant UCL Statistics |  |  |  |
| Normal Distribution Test |  | Lognormal Distribution Test |  |
| Lilliefors Test Statistic | 0.074177 | Lilliefors Test Statistic | 0.132277 |
| Lilliefors Critical Value | 0.072102 | Lilliefors Critical Value | 0.072102 |
| Data not Normal at 5\% Significance Level Data not Lognormal at 5\% Significance Level |  |  |  |
| Assuming Normal Distribution |  | Assuming Lognormal Distribution |  |
| 95\% Student's-t UCL | 9.993834 | 95\% H-UCL | 11.15573 |
| 95\% UCLs (Adjusted for Skewness) |  | 95\% Chebyshev (MVUE) UCL | 12.81379 |
| 95\% Adjusted-CLT UCL | 9.998274 | 97.5\% Chebyshev (MVUE) UCL | 14.10726 |
| 95\% Modified-t UCL | 9.995219 | 99\% Chebyshev (MVUE) UCL | 16.64802 |
| Gamma Distribution Test |  | Data Distribution |  |
| k star (bias corrected) | 2.287585 | Data do not follow a Discernable Distribution (0.05) |  |
| Theta Star | 4.054277 |  |  |
| MLE of Mean | 9.274503 |  |  |
| MLE of Standard Deviation | 6.131998 |  |  |
| nu star | 690.8507 |  |  |
| Approximate Chi Square Value (.05) | 630.8674 | Nonparametric Statistics |  |
| Adjusted Level of Significance | 0.048411 | 95\% CLT UCL | 9.989391 |
| Adjusted Chi Square Value | 630.3215 | 95\% Jackknife UCL | 9.993834 |
|  |  | 95\% Standard Bootstrap UCL | 9.978401 |
| Anderson-Darling Test Statistic | 1.970888 | 95\% Bootstrap-t UCL | 9.985983 |
| Anderson-Darling 5\% Critical Value | 0.763673 | 95\% Hall's Bootstrap UCL | 9.983493 |
| Kolmogorov-Smirnov Test Statistic | 0.090642 | 95\% Percentile Bootstrap UCL | 9.988874 |
| Kolmogorov-Smirnov 5\% Critical Value | 0.077136 | 95\% BCA Bootstrap UCL | 9.994106 |
| Data not Gamma Distributed at 5\% Significance Level |  | 95\% Chebyshev(Mean, Sd) UCL | 11.16897 |
|  |  | 97.5\% Chebyshev(Mean, Sd) UCL | 11.98871 |
| Assuming Gamma Distribution |  | 99\% Chebyshev(Mean, Sd) UCL | 13.59893 |
| 95\% Approximate Gamma UCL | 10.15633 |  |  |
| 95\% Adjusted Gamma UCL | 10.16513 |  |  |
| Potential UCL to Use |  | Use 95\% Chebyshev (Mean, Sd) UCL | 11.16897 |


| Barium |  |  |  |
| :---: | :---: | :---: | :---: |
| General Statistics |  |  |  |
| Number of Valid Data |  | Number of Detected Data | 98 |
| Number of Distinct Detected Data |  | Number of Non-Detect Data | 1 |
|  |  | Percent Non-Detects | 1.01\% |
| Raw Statistics |  | Log-transformed Statistics |  |
| Minimum Detected | 18.2 | Minimum Detected | 2.901422 |
| Maximum Detected | 254 | Maximum Detected | 5.537334 |
| Mean of Detected | 91.14796 | Mean of Detected | 4.394748 |
| SD of Detected | 44.17457 | SD of Detected | 0.50327 |
| Minimum Non-Detect |  | Minimum Non-Detect | 4.099332 |
| Maximum Non-Detect | 60.3 | Maximum Non-Detect | 4.099332 |
| UCL Statistics |  |  |  |
| Normal Distribution Test with Detected Values Only |  | Lognormal Distribution Test with Detected Values Only |  |
| Lilliefors Test Statistic | 0.087527 | Lilliefors Test Statistic | 0.07217 |
| 5\% Lilliefors Critical Value | 0.0895 | 5\% Lilliefors Critical Value | 0.0895 |
| Data appear Normal at 5\% Significance Level Data appear Lognormal at 5\% Significance Level |  |  |  |
| Assuming Normal Distribution |  | Assuming Lognormal Distribution |  |
| DL/2 Substitution Method DL/2 Substitution Method |  |  |  |
| Mean | 90.53182 | Mean | 4.384762 |
| SD | 44.37413 |  | 0.510458 |
| 95\% DL/2 (t) UCL | 97.93749 | 95\% H-Stat (DL/2) UCL | 100.5483 |
| Maximum Likelihood Estimate(MLE) Method |  | Log ROS Method |  |
| Mean | 84.93734 | Mean in Log Scale | 4.389324 |
| SD | 52.13543 | SD in Log Scale | 0.503595 |
| 95\% MLE (t) UCL | 93.63831 | Mean in Original Scale | 90.7057 |
| 95\% MLE (Tiku) UCL | 94.19525 | SD in Original Scale | 44.16837 |
|  |  | 95\% Percentile Bootstrap UCL | 98.43967 |
|  |  | 95\% BCA Bootstrap UCL | 97.70129 |
| Gamma Distribution Test with Detected Values Only |  | Data Distribution Test with Detected Values Only |  |
| k star (bias corrected) | 4.27849 | Data appear Normal at 5\% Significance Level |  |
| Theta Star | 21.30377 |  |  |
| nu star | 838.5841 |  |  |
| A-D Test Statistic | 0.275107 | Nonparametric Statistics |  |
| 5\% A-D Critical Value | 0.755466 | Kaplan-Meier (KM) Method |  |
| K-S Test Statistic | 0.755466 | Mean | 90.68649 |
| 5\% K-S Critical Value | 0.090653 |  | 43.97804 |
| Data appear Gamma Distributed at 5\% Significance Level | SE of Mean |  | 4.444143 |
|  |  | 95\% KM (t) UCL | 98.06621 |
| Assuming Gamma Distribution |  | 95\% KM (z) UCL | 97.99645 |
| Gamma ROS Statistics using Extrapolated Data |  | 95\% KM (jackknife) UCL | 98.066 |
| Minimum | 18.2 | 95\% KM (bootstrap t) UCL | 98.91016 |
| Maximum | 254 | 95\% KM (BCA) UCL | 98.56493 |
| Mean | 90.71173 | 95\% KM (Percentile Bootstrap) UCL | 97.96768 |
| Median | 84.6 | 95\% KM (Chebyshev) UCL | 110.0581 |
| SD | 44.16243 | 97.5\% KM (Chebyshev) UCL | 118.4402 |
| k star | 4.262372 | 99\% KM (Chebyshev) UCL | 134.9052 |
| Theta star | 21.28198 |  |  |
| Nu star | 843.9497 | Potential UCLs to Use |  |
| AppChi2 | 777.5285 | 95\% KM (t) UCL | 98.06621 |
| 95\% Gamma Approximate UCL | 98.46088 | 95\% KM (Percentile Bootstrap) UCL | 97.96768 |
| 95\% Adjusted Gamma UCL | 98.57906 |  |  |



| Copper |  |  |  |
| :---: | :---: | :---: | :---: |
| General Statistics |  |  |  |
| Number of Valid Observations |  | Number of Distinct Observations | 95 |
| Raw Statistics |  | Log-transformed Statistics |  |
| Minimum | 16.2 | Minimum of Log Data | 2.785011 |
| Maximum | 444 | Maximum of Log Data | 6.095825 |
| Mean | 78.72929 | Mean of log Data | 4.16272 |
| Median | 62.5 | SD of log Data | 0.596416 |
| SD | 65.80122 |  |  |
| Coefficient of Variation | 0.835791 |  |  |
| Skewness | 3.26125 |  |  |
| Relevant UCL Statistics |  |  |  |
| Normal Distribution Test |  | Lognormal Distribution Test |  |
| Lilliefors Test Statistic | 0.240049 L | Lilliefors Test Statistic | 0.09358 |
| Lilliefors Critical Value | 0.089046 | Lilliefors Critical Value | 0.089046 |
| Data not Normal at 5\% Significance Level Data not Lognormal at 5\% Significance Level |  |  |  |
| Assuming Normal Distribution |  | Assuming Lognormal Distribution |  |
| 95\% Student's-t UCL 89.71097 | 89.71097 | 95\% H-UCL | 86.14791 |
| 95\% UCLs (Adjusted for Skewness) |  | 95\% Chebyshev (MVUE) UCL | 98.1514 |
| 95\% Adjusted-CLT UCL | 91.92329 | 97.5\% Chebyshev (MVUE) UCL | 107.4809 |
| 95\% Modified-t UCL | 90.07224 | 99\% Chebyshev (MVUE) UCL | 125.8069 |
| Gamma Distribution Test |  | Data Distribution |  |
| k star (bias corrected) | 2.541646 | Data do not follow a Discernable Distribution (0.05) |  |
| Theta Star | 30.97572 |  |  |
| MLE of Mean | 78.72929 |  |  |
| MLE of Standard Deviation | 49.38316 |  |  |
| nu star | 503.2458 |  |  |
| Approximate Chi Square Value (.05) | 452.224 N | Nonparametric Statistics |  |
| Adjusted Level of Significance | 0.047576 | 95\% CLT UCL | 89.60716 |
| Adjusted Chi Square Value | 451.5162 | 95\% Jackknife UCL | 89.71097 |
|  |  | 95\% Standard Bootstrap UCL | 89.43421 |
| Anderson-Darling Test Statistic | 3.229802 | 95\% Bootstrap-t UCL | 92.89755 |
| Anderson-Darling 5\% Critical Value | 0.761093 | 95\% Hall's Bootstrap UCL | 92.90683 |
| Kolmogorov-Smirnov Test Statistic | 0.142362 | 95\% Percentile Bootstrap UCL | 90.31818 |
| Kolmogorov-Smirnov 5\% Critical Value | 0.090771 | 95\% BCA Bootstrap UCL | 91.33737 |
| Data not Gamma Distributed at 5\% Significance Level |  | 95\% Chebyshev(Mean, Sd) UCL | 107.5559 |
|  |  | 97.5\% Chebyshev(Mean, Sd) UCL | 120.0292 |
| Assuming Gamma Distribution |  | 99\% Chebyshev(Mean, Sd) UCL | 144.5305 |
| 95\% Approximate Gamma UCL | 87.61188 |  |  |
| 95\% Adjusted Gamma UCL | 87.74921 |  |  |
| Potential UCL to Use |  | Use 95\% Chebyshev (Mean, Sd) UCL | 107.5559 |


|  | Fluoranthene |
| :--- | :--- |
|  |  |
| General Statistics |  |
| Number of Valid Observations | 5 Number of Distinct Observations |
| Number of Missing Values | 1 |
| Raw Statistics | Log-transformed Statistics |
| Minimum | 5 Minimum of Log Data |
| Maximum | 400 Maximum of Log Data |
| Mean | 158 Mean of log Data |
| Median | 100 SD of log Data |
| SD | 159.0047 |
| Coefficient of Variation | 1.006359 |
| Skewness | 1.002926 |

Warning: A sample size of ' $n$ ' = 5 may not adequate enough to compute meaningful and reliable test statistics and estimates!
It is suggested to collect at least 8 to 10 observations using these statistical methods!
If possible compute and collect Data Quality Objectives (DQO) based sample size and analytical results.

Warning: There are only 5 Values in this data
Note: It should be noted that even though bootstrap methods may be performed on this data set, the resulting calculations may not be reliable enough to draw conclusions

The literature suggests to use bootstrap methods on data sets having more than 10-15 observations.

| Relevant UCL Statistics |  |  |  |
| :---: | :---: | :---: | :---: |
| Normal Distribution Test |  | Lognormal Distribution Test |  |
| Shapiro Wilk Test Statistic | 0.918315 | Shapiro Wilk Test Statistic | 0.917191 |
| Shapiro Wilk Critical Value | 0.762 | Shapiro Wilk Critical Value | 0.762 |
| Data appear Normal at 5\% Significance Level | Data appear Lognormal at 5\% Significance Level |  |  |
| Assuming Normal Distribution Assuming Lognormal Distribution |  |  |  |
| 95\% Student's-t UCL | 309.5937 | 95\% H-UCL | 297062.2 |
| 95\% UCLs (Adjusted for Skewness) |  | 95\% Chebyshev (MVUE) UCL | 797.7536 |
| 95\% Adjusted-CLT UCL | 309.0432 | 97.5\% Chebyshev (MVUE) UCL | 1054.193 |
| 95\% Modified-t UCL | 314.9093 | 99\% Chebyshev (MVUE) UCL | 1557.917 |
| Gamma Distribution Test |  | Data Distribution |  |
| k star (bias corrected) | 0.457026 | Data appear Normal at 5\% Significance Level |  |
| Theta Star | 345.7136 |  |  |
| MLE of Mean | 158 |  |  |
| MLE of Standard Deviation | 233.7151 |  |  |
| nu star | 4.570257 |  |  |
| Approximate Chi Square Value (.05) | 0.959079 | Nonparametric Statistics |  |
| Adjusted Level of Significance | 0.0086 | 95\% CLT UCL | 274.964 |
| Adjusted Chi Square Value | 0.424094 | 95\% Jackknife UCL | 309.5937 |
|  |  | 95\% Standard Bootstrap UCL | 263.4805 |
| Anderson-Darling Test Statistic | 0.193526 | 95\% Bootstrap-t UCL | 540.7539 |
| Anderson-Darling 5\% Critical Value | 0.696613 | 95\% Hall's Bootstrap UCL | 1157.799 |
| Kolmogorov-Smirnov Test Statistic | 0.165777 | 95\% Percentile Bootstrap UCL | 271 |
| Kolmogorov-Smirnov 5\% Critical Value | 0.366135 | 95\% BCA Bootstrap UCL | 272 |
| Data appear Gamma Distributed at 5\% Significance Level |  | 95\% Chebyshev(Mean, Sd) UCL | 467.9573 |
|  |  | 97.5\% Chebyshev(Mean, Sd) UCL | 602.076 |
| Assuming Gamma Distribution |  | 99\% Chebyshev(Mean, Sd) UCL | 865.5263 |
| 95\% Approximate Gamma UCL | 752.9106 |  |  |
| 95\% Adjusted Gamma UCL | 1702.69 |  |  |
| Potential UCL to Use |  | Use 95\% Student's-t UCL | 309.5937 |





| Nickel |  |  |  |
| :---: | :---: | :---: | :---: |
| General Statistics |  |  |  |
| Number of Valid Observations |  | Number of Distinct Observations | 95 |
| Raw Statistics |  | Log-transformed Statistics |  |
| Minimum | 12.3 | Minimum of Log Data | 2.509599 |
| Maximum | 345 | Maximum of Log Data | 5.843544 |
| Mean | 66.04748 | Mean of log Data | 4.061088 |
| Median | 58.5 | SD of log Data | 0.506569 |
| SD | 40.06534 |  |  |
| Coefficient of Variation | 0.606614 |  |  |
| Skewness | 3.766068 |  |  |
| Relevant UCL Statistics |  |  |  |
| Normal Distribution Test |  | Lognormal Distribution Test |  |
| Lilliefors Test Statistic | 0.151736 | Lilliefors Test Statistic | 0.090287 |
| Lilliefors Critical Value | 0.089046 | Lilliefors Critical Value | 0.089046 |
| Data not Normal at 5\% Significance Level |  | Data not Lognormal at 5\% Significance Level |  |
| Assuming Normal Distribution |  | Assuming Lognormal Distribution |  |
| 95\% Student's-t UCL | 72.73405 | 95\% H-UCL | 72.55989 |
| 95\% UCLs (Adjusted for Skewness) |  | 95\% Chebyshev (MVUE) UCL | 81.33741 |
| 95\% Adjusted-CLT UCL | 74.29939 | 97.5\% Chebyshev (MVUE) UCL | 88.02294 |
| 95\% Modified-t UCL | 72.98807 | 99\% Chebyshev (MVUE) UCL | 101.1554 |
| Gamma Distribution Test |  | Data Distribution |  |
| k star (bias corrected) | 3.91124 | Data Follow Appr. Gamma Distributi | Level |
| Theta Star | 16.88658 |  |  |
| MLE of Mean | 66.04748 |  |  |
| MLE of Standard Deviation | 33.39635 |  |  |
| nu star | 774.4254 |  |  |
| Approximate Chi Square Value (.05) | 710.8483 | Nonparametric Statistics |  |
| Adjusted Level of Significance | 0.047576 | 95\% CLT UCL | 72.67084 |
| Adjusted Chi Square Value | 709.9577 | 95\% Jackknife UCL | 72.73405 |
|  |  | 95\% Standard Bootstrap UCL | 72.38582 |
| Anderson-Darling Test Statistic | 1.117951 | 95\% Bootstrap-t UCL | 75.25746 |
| Anderson-Darling 5\% Critical Value | 0.756231 | 95\% Hall's Bootstrap UCL | 79.68338 |
| Kolmogorov-Smirnov Test Statistic | 0.089042 | 95\% Percentile Bootstrap UCL | 72.86869 |
| Kolmogorov-Smirnov 5\% Critical Value | 0.090239 | 95\% BCA Bootstrap UCL | 75.25657 |
| Data follow Appr. Gamma Distribution at 5\% Significance Level |  | 95\% Chebyshev(Mean, Sd) UCL | 83.59953 |
|  |  | 97.5\% Chebyshev(Mean, Sd) UCL | 91.19432 |
| Assuming Gamma Distribution |  | 99\% Chebyshev(Mean, Sd) UCL | 106.1128 |
| 95\% Approximate Gamma UCL | 71.95466 |  |  |
| 95\% Adjusted Gamma UCL | 72.04492 |  |  |
| Potential UCL to Use |  | Use 95\% Approximate Gamma UCL | 71.95466 |




| Vanadium |  |  |  |
| :---: | :---: | :---: | :---: |
| General Statistics |  |  |  |
| Number of Valid Observations |  | Number of Distinct Observations | 84 |
| Raw Statistics |  | Log-transformed Statistics |  |
| Minimum | 33.2 N | Minimum of Log Data | 3.50255 |
| Maximum | 345 | Maximum of Log Data | 5.843544 |
| Mean | 100.8253 | Mean of log Data | 4.535009 |
| Median | 93.7 | SD of log Data | 0.383957 |
| SD | 46.12171 |  |  |
| Coefficient of Variation | 0.457442 |  |  |
| Skewness | 2.518952 |  |  |
| Relevant UCL Statistics |  |  |  |
| Normal Distribution Test |  | Lognormal Distribution Test |  |
| Lilliefors Test Statistic | 0.158535 L | Lilliefors Test Statistic | 0.071974 |
| Lilliefors Critical Value | 0.089046 L | Lilliefors Critical Value | 0.089046 |
| Data not Normal at 5\% Significance Level |  | Data appear Lognormal at 5\% Significance Level |  |
| Assuming Normal Distribution |  | Assuming Lognormal Distribution |  |
| 95\% Student's-t UCL | 108.5226 | 95\% H-UCL | 107.559 |
| 95\% UCLs (Adjusted for Skewness) |  | 95\% Chebyshev (MVUE) UCL | 117.7015 |
| 95\% Adjusted-CLT UCL | 109.7037 | 97.5\% Chebyshev (MVUE) UCL | 125.2417 |
| 95\% Modified-t UCL | 108.7182 | 99\% Chebyshev (MVUE) UCL | 140.0529 |
| Gamma Distribution Test |  | Data Distribution |  |
| k star (bias corrected) | 6.349875 D | Data appear Lognormal at 5\% Significance Level |  |
| Theta Star | 15.8783 |  |  |
| MLE of Mean | 100.8253 |  |  |
| MLE of Standard Deviation | 40.01167 |  |  |
| nu star | 1257.275 |  |  |
| Approximate Chi Square Value (.05) | 1175.946 | Nonparametric Statistics |  |
| Adjusted Level of Significance | 0.047576 | 95\% CLT UCL | 108.4498 |
| Adjusted Chi Square Value | 1174.797 | 95\% Jackknife UCL | 108.5226 |
|  |  | 95\% Standard Bootstrap UCL | 108.7321 |
| Anderson-Darling Test Statistic | 1.664037 | 95\% Bootstrap-t UCL | 110.1939 |
| Anderson-Darling 5\% Critical Value | 0.753691 | 95\% Hall's Bootstrap UCL | 111.2009 |
| Kolmogorov-Smirnov Test Statistic | 0.098348 | 95\% Percentile Bootstrap UCL | 108.6525 |
| Kolmogorov-Smirnov 5\% Critical Value | 0.090015 | 95\% BCA Bootstrap UCL | 109.3333 |
| Data not Gamma Distributed at 5\% Significance Level |  | 95\% Chebyshev(Mean, Sd) UCL | 121.0305 |
|  |  | 97.5\% Chebyshev(Mean, Sd) UCL | 129.7734 |
| Assuming Gamma Distribution |  | 99\% Chebyshev(Mean, Sd) UCL | 146.947 |
| 95\% Approximate Gamma UCL | 107.7984 |  |  |
| 95\% Adjusted Gamma UCL | 107.9038 |  |  |
| Potential UCL to Use |  | Use 95\% Student's-t UCL | 108.5226 |
|  |  | or 95\% Modified-t UCL | 108.7182 |
|  |  | or 95\% H-UCL | 107.559 |



Table B. 3
Estimated Sample Sizes for Metals and
Supporting Inputs to the Calculations 4835 Glenbrook Rd.

| Metal | Existing Data |  |  |  |  | Action Level (mg/kg) | $\begin{gathered} \Delta \\ (\mathrm{mg} / \mathrm{kg}) \\ \hline \end{gathered}$ | Test to use | Recommended Sample Size |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | Sample Size | Number of Non-detects | Distribution ${ }^{1}$ | Central Tendency ${ }^{2,3,4}$ (mg/kg) | $\begin{gathered} \mathrm{SD}^{2,3,4} \\ \text { (unitless) } \end{gathered}$ |  |  |  |  |
| Aluminum | 97 | 1 | Lognormal | 24,075.1 | 9,018 | 77,000 | 15,400 | Mann-Whitney U | 5 |
| Antimony | 99 | 31 | Lognormal | 0.9 | 0.70 | 31 | 6.2 | Mann-Whitney U | 2 |
| Arsenic | 151 | 0 | Nonparametric | 9.1 | 5.34 | 20 | 4.0 | Mann-Whitney U | 15 |
| Barium | 99 | 1 | Normal | 90.7 | 43.98 | 15,000 | 3,000 | Mann-Whitney U | 2 |
| Cadmium | 99 | 48 | Lognormal | 0.2 | 0.18 | 70 | 14 | Mann-Whitney U | 2 |
| Copper | 99 | 0 | Nonparametric | 62.5 | 65.80 | 3,100 | 620 | Mann-Whitney U | 2 |
| Lead | 99 | 1 | Nonparametric | 13.8 | 10.40 | 400 | 80 | Mann-Whitney U | 2 |
| Manganese | 99 | 1 | Lognormal | 669.4 | 512.05 | 1,800 | 360 | Mann-Whitney U | 17 |
| Mercury | 99 | 31 | Lognormal | 0.1 | 0.12 | 7.8 | 1.56 | Mann-Whitney U | 2 |
| Nickel | 99 | 0 | Nonparametric | 58.5 | 40.07 | 1,600 | 320 | Mann-Whitney U | 2 |
| Thallium | 98 | 64 | Nonparametric | 1.2 | 0.91 | 5.1 | 1.02 | Mann-Whitney U | 8 |
| Vanadium | 99 | 0 | Lognormal | 93.2 | 1.47 | 390 | 78 | t-test of logged data | 2 |
| Zinc | 99 | 0 | Lognormal | 67.1 | 1.38 | 23,000 | 4,600 | t-test of logged data | 2 |

Notes:
1 - of the detected data
2 - for lognormal data without NDs, the central tendency is the mean of the log10-transformed data back-transformed into normal space and the SD is the SD of the log10-transformed data back-transformed into normal space.
3 - for nonparametric data without NDs, the central tendency is the median and the SD is the arithmetic SD
4 - for data with NDs, the central tendency is the Kaplan-Meier mean and the SD is the Kaplan-Meier SD
Definitions:
$\Delta$ - minimum detectable difference from action level
ND - non-detect
SD - standard deviation

Table B. 4
Action Levels 4835 Glenbrook Rd.

| Metal | Background UTL (mg/kg) | Background Central Tendency ( $\mathrm{mg} / \mathrm{kg}$ ) | Residential PRG ${ }^{1}$ (mg/kg) | $\begin{gathered} \text { Action Level² } \\ (\mathrm{mg} / \mathrm{kg}) \end{gathered}$ |
| :---: | :---: | :---: | :---: | :---: |
| Aluminum ${ }^{\text {a }}$ | 19,100 | 11,500 | 77,000 | 77,000 |
| Antimony ${ }^{\text {b,c }}$ | 5.2 | 0.52 | 31 | 31 |
| Arsenic ${ }^{\text {d }}$ | 12.6 | 5 | 20 | 20 |
| Barium ${ }^{\text {a }}$ | 172 | 69.65 | 15,000 | 15,000 |
| Benzo(a)anthracene ${ }^{\text {e }}$ | 0.36 | 182.1 | 0.15 | 182.10 |
| Benzo(a)pyrene ${ }^{\text {e,f }}$ | 0.40 | 0.19 | 0.015 | 0.40 |
| Benzo(b)fluoranthene ${ }^{\text {e }}$ | 0.37 | 157.7 | 0.15 | 157.70 |
| Benzo(k)fluoranthene ${ }^{\text {e }}$ | 0.37 | 162.3 | 1.5 | 162.3 |
| Cadmium ${ }^{\text {e }}$ | 2.36 | 0.72 | 70 | 70 |
| Chrysene ${ }^{\text {e }}$ | 0.40 | 178.7 | 15 | 178.7 |
| Copper ${ }^{\text {a }}$ | 49.65 | 26.1 | 3,100 | 3,100 |
| Fluoranthene ${ }^{\text {e }}$ | 0.70 | 254.2 | 2,300 | 2,300 |
| $L_{\text {Lead }}{ }^{\text {a }}$ | 194 | 49.4 | 400 | 400 |
| Manganese ${ }^{\text {a }}$ | 968 | 357 | 1,800 | 1,800 |
| Mercury ${ }^{\text {3,a }}$ | 0.25 | 0.086 | 7.8 | 7.8 |
| Nickel ${ }^{\text {a }}$ | 33.5 | 19.5 | 1,600 | 1,600 |
| Pyrene ${ }^{\text {e,f }}$ | 0.63 | 0.23 | 1,700 | 1,700 |
| Thallium ${ }^{\text {b,g }}$ | 2.2 | 0.96 | 5.1 | 5.1 |
| 1,1,2,2-Tetrachloroethane | NA | NA | 0.6 | 0.6 |
| 1,1,2-Trichloroethane | NA | NA | 1.1 | 1.1 |
| Thiodiglycol ${ }^{4}$ | NA | NA | 39 | 39 |
| Vanadium ${ }^{\text {a }}$ | 75.5 | 45.9 | 390 | 390 |
| Zinc ${ }^{\text {a }}$ | 158 | 81.3 | 23,000 | 23,000 |

Notes:
1 - the residential PRGs listed here are the lesser of the cancer-based and non-cancer based September 2008 USEPA Regional Screening Levels (RSL), except for arsenic, which is the Spring Valley remediation goal agreed by USACE, USEPA and DDOE.
2 - the greater of background and the residential PRG.
3 - To be health-protective, the residential RSL for methyl mercury was used.
4 - From Remedial Investigation Report for the Operation Safe Removal Formerly Used Defense Site, Washington, D.C. (Parsons, 1995).
a - Background UTL is the nonparametric upper 90th percentile with $95 \%$ confidence (Parsons, 2008; see Table B.6); central tendency is the median (see Table B.6)
b-Background UTL is the maximum detection limit of the non-detects (Parsons, 2008)
c - Background central tendency is the KM mean (see Table B.6)
d - Background UTL is $11.1 \mathrm{mg} / \mathrm{kg}$. However, the UTL calculated in a previous report using a smaller dataset of $12.6 \mathrm{mg} / \mathrm{kg}$ was retained (Parsons, 2008); central tendency is the median (see Table B.6)
e - Background UTL is the 95\% KM UTL with 90\% coverage (Parsons, 2008; see Table B.6); central tendency is the KM mean (see Table B.6)
f The UTL presented here was calculated (see Table B.6) using the data presented in the Background Sampling Report (Parsons, 2008); however, the UTL calculated here differs from the Background Sampling Report due to a discrepancy in the number of non-detects
g - Background central tendency is the single detected value

## Definitions:

> KM - Kaplan-Meier
> NA - Not applicable
> PRG - Preliminary Remediation Goal
> RSL - Regional Screening Level
> UTL - Upper tolerance limit

Table B. 5
Estimated Sample Sizes for Organics and Supporting Inputs to the Calculations 4835 Glenbrook Rd.

| VOC | Existing Data |  |  |  |  |  | Action Level (ug/kg) | $\begin{gathered} \Delta \\ (\mathrm{ug} / \mathrm{kg}) \\ \hline \end{gathered}$ | Test to use | RecommendedSample Size |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | Sample | Number of |  | DLs for NDs (ug/kg) |  | SD |  |  |  |  |
|  | Size | Non-detects | Distribution | Minimum | Maximum | (unitless) |  |  |  |  |
| Benzo(a)anthracene ${ }^{1}$ | 6 | 5 | NA | 100 | 400 | 61.24 | 357.5 | 72 | Mann-Whitney U | 7 |
| Benzo(a)pyrene ${ }^{1}$ | 6 | 5 | NA | 100 | 400 | 66.01 | 375.0 | 75 | Mann-Whitney U | 8 |
| Benzo(b)fluoranthene ${ }^{1}$ | 6 | 5 | NA | 100 | 400 | 68.37 | 365.7 | 73 | Mann-Whitney U | 8 |
| Benzo(k)fluoranthene ${ }^{1}$ | 6 | 5 | NA | 100 | 400 | 64.25 | 1,500 | 300 | Mann-Whitney U | 2 |
| Chrysene ${ }^{1}$ | 6 | 5 | NA | 100 | 400 | 62.82 | 15,000 | 3,000 | Mann-Whitney U | 2 |
| Fluoranthene ${ }^{1}$ | 5 | 2 | Normal | 100 | 400 | 100.16 | 230,000 | 46,000 | Mann-Whitney U | 2 |
| Pyrene ${ }^{1}$ | 4 | 2 | NA | 100 | 400 | 100.07 | 170,000 | 34,000 | Mann-Whitney U | 2 |
| 1,1,2,2-Tetrachloroethane ${ }^{1}$ | 17 | 16 | NA | 1 | 12 | 91.54 | 590 | 118 | Mann-Whitney U | 6 |
| 1,1,2-Trichloroethane ${ }^{1}$ | 17 | 16 | NA | 1 | 12 | 76.99 | 1,100 | 220 | Mann-Whitney U | 3 |
| Thiodiglycol ${ }^{2}$ | 8 | 4 | Normal | 575 | 610 | 131.36 | 39,100 | 7,820 | Mann-Whitney U | 2 | Notes:

1 - arithmetic SD calculated using 1/2 DL.
2 - arithmetic SD calculated using Kaplan-Meier method in ProUCL
Definitions:
$\Delta$ - minimum detectable difference from action level
DL - detection limit
ND - non-detect
SD - standard deviation

Table B. 6

## ProUCL UTL Output for Metals Data

 4835 Glenbrook Rd| Full Precision | OFF |
| :--- | :---: |
| Confidence Coefficient | $95 \%$ |
| Coverage | $90 \%$ |
| Different or Future K Values | 1 |
| Number of Bootstrap Operations | 2000 |


| Aluminum |  |  |
| :---: | :---: | :---: |
| General Statistics |  |  |
| Total Number of Observations | 45 Number of Distinct Observations | 40 |
| Raw Statistics | Log-Transformed Statistics |  |
| Minimum | 5300 Minimum | 8.575 |
| Maximum | 21600 Maximum | 9.98 |
| Second Largest | 21300 Second Largest | 9.966 |
| First Quartile | 9055 First Quartile | 9.11 |
| Median | 11500 Median | 9.35 |
| Third Quartile | 14050 Third Quartile | 9.55 |
| Mean | 11694 Mean | 9.313 |
| SD | 3863 SD | 0.338 |
| Coefficient of Variation | 0.33 |  |
| Skewness | 0.59 |  |
| Background Statistics |  |  |
| Normal Distribution Test | Lognormal Distribution Test |  |
| Shapiro Wilk Test Statistic | 0.956 Shapiro Wilk Test Statistic | 0.973 |
| Shapiro Wilk Critical Value | 0.945 Shapiro Wilk Critical Value | 0.945 |
| Data appear Normal at 5\% Significance Level | Data appear Lognormal at 5\% Significance Level |  |
| Assuming Normal Distribution | Assuming Lognormal Distribution |  |
| 95\% UTL with 90\% Coverage | 18116 95\% UTL with 90\% Coverage | 19427 |
| 95\% UPL (t) | 18256 95\% UPL (t) | 19667 |
| 90\% Percentile (z) | 16645 90\% Percentile (z) | 17080 |
| 95\% Percentile (z) | 18048 95\% Percentile (z) | 19311 |
| 99\% Percentile (z) | 20680 99\% Percentile (z) | 24313 |
| Gamma Distribution Test | Data Distribution Test |  |
| k star | 8.766 Data appear Normal at 5\% Significance Level |  |
| Theta Star | 1334 |  |
| MLE of Mean | 11694 |  |
| MLE of Standard Deviation | 3950 |  |
| nu star | 789 |  |
| A-D Test Statistic | 0.248 Nonparametric Statistics |  |
| 5\% A-D Critical Value | 0.749 90\% Percentile | 16520 |
| K-S Test Statistic | 0.0752 95\% Percentile | 20640 |
| $5 \% \mathrm{~K}-\mathrm{S}$ Critical Value | 0.132 99\% Percentile | 21600 |
| Data appear Gamma Distributed at 5\% Significance Level |  |  |
| Assuming Gamma Distribution | 95\% UTL with 90\% Coverage | 19100 |
| 90\% Percentile | 16955 95\% Percentile Bootstrap UTL with 90\% Coverage | 19100 |
| 95\% Percentile | 18857 95\% BCA Bootstrap UTL with 90\% Coverage | 19100 |
| 99\% Percentile | 22781 95\% UPL | 20640 |
|  | 95\% Chebyshev UPL | 28718 |
| 95\% WH Approx. Gamma UPL | 18980 Upper Threshold Limit Based upon IQR | 21543 |
| 95\% HW Approx. Gamma UPL | 19124 |  |
| 95\% WH Approx. Gamma UTL with 90\% Coverage | 18787 |  |
| 95\% HW Approx. Gamma UTL with 90\% Coverage | 18921 |  |






| Copper |  |  |  |
| :---: | :---: | :---: | :---: |
| General Statistics |  |  |  |
| Total Number of Observations |  | Number of Distinct Observations | 43 |
| Raw Statistics |  | Log-Transformed Statistics |  |
| Minimum |  | Minimum | 2.23 |
| Maximum | 76.7 | Maximum | 4.34 |
| Second Largest | 57.8 | Second Largest | 4.057 |
| First Quartile | 19.5 | First Quartile | 2.97 |
| Median | 26.1 | Median | 3.262 |
| Third Quartile | 35.45 | Third Quartile | 3.568 |
| Mean | 28.49 | Mean | 3.253 |
| SD | 13.11 |  | 0.448 |
| Coefficient of Variation | 0.46 |  |  |
| Skewness | 1.356 |  |  |
| Background Statistics |  |  |  |
| Normal Distribution Test |  | Lognormal Distribution Test |  |
| Shapiro Wilk Test Statistic | 0.916 | Shapiro Wilk Test Statistic | 0.988 |
| Shapiro Wilk Critical Value | 0.945 | Shapiro Wilk Critical Value | 0.945 |
| Data not Normal at 5\% Significance Level |  | Data appear Lognormal at 5\% Significance Level |  |
| Assuming Normal Distribution |  | Assuming Lognormal Distribution |  |
| 95\% UTL with 90\% Coverage$95 \%$ UPL (t) | 50.28 | 95\% UTL with 90\% Coverage | 54.47 |
|  | 50.76 | 95\% UPL (t) | 55.37 |
| 90\% Percentile (z) | 45.29 | 90\% Percentile (z) | 45.92 |
| 95\% Percentile (z) | 50.05 | 95\% Percentile (z) | 54.04 |
| 99\% Percentile (z) | 58.98 | 99\% Percentile (z) | 73.35 |
| Gamma Distribution Test |  | Data Distribution Test |  |
| k star | 4.991 Data appear Gamma Distributed at 5\% Significance Level |  |  |
| Theta Star | 5.708 ( |  |  |
| MLE of Mean | 28.49 |  |  |
| MLE of Standard Deviation | 12.75 |  |  |
| nu star | 449.2 |  |  |
| A-D Test Statistic | 0.175 | Nonparametric Statistics |  |
| 5\% A-D Critical Value | 0.753 | 90\% Percentile | 44.24 |
| K-S Test Statistic | 0.0747 | 95\% Percentile | 55.36 |
| 5\% K-S Critical Value | 0.132 | 99\% Percentile | 76.7 |
| Data appear Gamma Distributed at 5\% Significance Level |  |  |  |
| Assuming Gamma Distribution |  | 95\% UTL with 90\% Coverage | 49.65 |
| 90\% Percentile | 45.56 | 95\% Percentile Bootstrap UTL with 90\% Coverage | 49.65 |
| 95\% Percentile | 52.17 | 95\% BCA Bootstrap UTL with 90\% Coverage | 49.65 |
| 99\% Percentile | 66.16 | 95\% UPL | 55.36 |
|  |  | 95\% Chebyshev UPL | 86.26 |
| 95\% WH Approx. Gamma UPL | 52.58 | Upper Threshold Limit Based upon IQR | 59.38 |
| 95\% HW Approx. Gamma UPL | 53.11 ( |  |  |
| 95\% WH Approx. Gamma UTL with 90\% Coverage | 51.9 |  |  |
| 95\% HW Approx. Gamma UTL with 90\% Coverage | 52.38 |  |  |



|  | Manganese |  |
| :---: | :---: | :---: |
| General Statistics |  |  |
| Total Number of Observations | 45 Number of Distinct Observations | 44 |
| Raw Statistics | Log-Transformed Statistics |  |
| Minimum | 143 Minimum | 4.963 |
| Maximum | 1000 Maximum | 6.908 |
| Second Largest | 981 Second Largest | 6.889 |
| First Quartile | 230.5 First Quartile | 5.44 |
| Median | 357 Median | 5.878 |
| Third Quartile | 647.3 Third Quartile | 6.472 |
| Mean | 442.7 Mean | 5.933 |
| SD | 249.8 SD | 0.581 |
| Coefficient of Variation | 0.564 |  |
| Skewness | 0.732 |  |
| Background Statistics |  |  |
| Normal Distribution Test | Lognormal Distribution Test |  |
| Shapiro Wilk Test Statistic | 0.898 Shapiro Wilk Test Statistic | 0.943 |
| Shapiro Wilk Critical Value | 0.945 Shapiro Wilk Critical Value | 0.945 |
| Data not Normal at 5\% Significance Level | Data not Lognormal at 5\% Significance Level |  |
| Assuming Normal Distribution | Assuming Lognormal Distribution |  |
| 95\% UTL with 90\% Coverage | 858 95\% UTL with 90\% Coverage | 991 |
| 95\% UPL (t) | 867.1 95\% UPL (t) | 1012 |
| 90\% Percentile (z) | 762.9 90\% Percentile (z) | 794.3 |
| 95\% Percentile (z) | 853.6 95\% Percentile (z) | 980.9 |
| 99\% Percentile (z) | 1024 99\% Percentile (z) | 1457 |
| Gamma Distribution Test | Data Distribution Test |  |
| k star | 3.083 Data appear Gamma Distributed at 5\% Significance Level |  |
| Theta Star | 143.6 |  |
| MLE of Mean | 442.7 |  |
| MLE of Standard Deviation | 252.1 |  |
| nu star | 277.5 |  |
| A-D Test Statistic | 0.603 Nonparametric Statistics |  |
| 5\% A-D Critical Value | 0.755 90\% Percentile | 811.4 |
| K-S Test Statistic | 0.11 95\% Percentile | 977.1 |
| 5\% K-S Critical Value | 0.133 99\% Percentile | 1000 |
| Data appear Gamma Distributed at 5\% Significance Level |  |  |
| Assuming Gamma Distribution | 95\% UTL with 90\% Coverage | 968 |
| 90\% Percentile | 780.8 95\% Percentile Bootstrap UTL with 90\% Coverage | 968 |
| 95\% Percentile | 921.8 95\% BCA Bootstrap UTL with 90\% Coverage | 968 |
| 99\% Percentile | 1227 95\% UPL | 977.1 |
|  | 95\% Chebyshev UPL | 1544 |
| 95\% WH Approx. Gamma UPL | 932.6 Upper Threshold Limit Based upon IQR | 1272 |
| 95\% HW Approx. Gamma UPL | 947.9 |  |
| 95\% WH Approx. Gamma UTL with 90\% Coverage | 917.9 |  |
| 95\% HW Approx. Gamma UTL with 90\% Coverage | 932 |  |


| Mercury (no outliers) |  |  |  |
| :---: | :---: | :---: | :---: |
| General Statistics |  |  |  |
| Total Number of Observations |  | Number of Distinct Observations | 33 |
| Raw Statistics |  | Log-Transformed Statistics |  |
| Minimum | 0.019 | Minimum | -3.963 |
| Maximum | 0.46 | Maximum | -0.777 |
| Second Largest | 0.29 | Second Largest | -1.238 |
| First Quartile | 0.064 | First Quartile | -2.749 |
| Median | 0.086 | Median | -2.453 |
| Third Quartile | 0.14 | Third Quartile | -1.966 |
| Mean | 0.113 | Mean | -2.383 |
| SD | 0.0827 |  | 0.633 |
| Coefficient of Variation | 0.731 |  |  |
| Skewness | 2.209 |  |  |
| Background Statistics |  |  |  |
| Normal Distribution Test |  | Lognormal Distribution Test |  |
| Shapiro Wilk Test Statistic | 0.802 | Shapiro Wilk Test Statistic | 0.993 |
| Shapiro Wilk Critical Value | 0.943 | Shapiro Wilk Critical Value | 0.943 |
| Data not Normal at 5\% Significance Level |  | Data appear Lognormal at 5\% Significance Level |  |
| Assuming Normal Distribution |  | Assuming Lognormal Distribution |  |
| 95\% UTL with 90\% Coverage$95 \%$ UPL (t) | 0.251 | 95\% UTL with 90\% Coverage | 0.266 |
|  | 0.254 | 95\% UPL (t) | 0.271 |
| 90\% Percentile (z) | 0.219 | 90\% Percentile (z) | 0.208 |
| 95\% Percentile (z) | 0.249 | 95\% Percentile (z) | 0.261 |
| 99\% Percentile (z) | 0.305 | 99\% Percentile (z) | 0.403 |
| Gamma Distribution Test |  | Data Distribution Test |  |
| k star | 2.45 Data appear Gamma Distributed at 5\% Significance Level |  |  |
| Theta Star | 0.0462 20 |  |  |
| MLE of Mean | 0.113 |  |  |
| MLE of Standard Deviation | 0.0722 |  |  |
| nu star | 210.7 |  |  |
| A-D Test Statistic | 0.65 | Nonparametric Statistics |  |
| 5\% A-D Critical Value | 0.757 | 90\% Percentile | 0.232 |
| K-S Test Statistic | 0.122 | 95\% Percentile | 0.282 |
| 5\% K-S Critical Value | 0.136 | 99\% Percentile | 0.46 |
| Data appear Gamma Distributed at 5\% Significance Level |  |  |  |
| Assuming Gamma Distribution |  | 95\% UTL with 90\% Coverage | 0.25 |
| 90\% Percentile | 0.21 | 95\% Percentile Bootstrap UTL with 90\% Coverage | 0.269 |
| 95\% Percentile | 0.252 | 95\% BCA Bootstrap UTL with 90\% Coverage | 0.25 |
| 99\% Percentile | 0.344 | 95\% UPL | 0.282 |
|  |  | 95\% Chebyshev UPL | 0.478 |
| 95\% WH Approx. Gamma UPL | 0.254 | Upper Threshold Limit Based upon IQR | 0.254 |
| 95\% HW Approx. Gamma UPL | 0.256 |  |  |
| 95\% WH Approx. Gamma UTL with 90\% Coverage | 0.25 |  |  |
| 95\% HW Approx. Gamma UTL with 90\% Coverage | 0.252 |  |  |



|  | Thallium |  |
| :--- | ---: | ---: |
| General Statistics |  |  |
| Number of Valid Data | 45 Number of Detected Data | 1 |
| Number of Distinct Detected Data | 1 Number of Non-Detect Data | 44 |
| Warning: Only one distinct data value was detected! ProUCL (or any other software) should not be used on such a data set! |  |  |
| It is suggested to use alternative site specific values determined by the Project Team to estimate environmental parameters (e.g., EPC, E |  |  |
| The data set for variable Thallium was not processed! |  |  |


| Vanadium |  |  |
| :---: | :---: | :---: |
| General Statistics |  |  |
| Total Number of Observations | 45 Number of Distinct Observations | 44 |
| Raw Statistics | Log-Transformed Statistics |  |
| Minimum | 26.5 Minimum | 3.277 |
| Maximum | 85 Maximum | 4.443 |
| Second Largest | 79.3 Second Largest | 4.373 |
| First Quartile | 37.55 First Quartile | 3.626 |
| Median | 45.9 Median | 3.826 |
| Third Quartile | 63.5 Third Quartile | 4.151 |
| Mean | 48.91 Mean | 3.842 |
| SD | 15.36 SD | 0.314 |
| Coefficient of Variation | 0.314 |  |
| Skewness | 0.535 |  |
| Background Statistics |  |  |
| Normal Distribution Test | Lognormal Distribution Test |  |
| Shapiro Wilk Test Statistic | 0.938 Shapiro Wilk Test Statistic | 0.959 |
| Shapiro Wilk Critical Value | 0.945 Shapiro Wilk Critical Value | 0.945 |
| Data not Normal at 5\% Significance Level | Data appear Lognormal at 5\% Significance Level |  |
| Assuming Normal Distribution | Assuming Lognormal Distribution |  |
| 95\% UTL with 90\% Coverage$95 \%$ UPL (t) | 74.45 95\% UTL with 90\% Coverage | 78.54 |
|  | 75.01 95\% UPL (t) | 79.44 |
| 90\% Percentile (z) | 68.6 90\% Percentile (z) | 69.7 |
| 95\% Percentile (z) | 74.18 95\% Percentile (z) | 78.11 |
| 99\% Percentile (z) | 84.65 99\% Percentile (z) | 96.73 |
| Gamma Distribution Test | Data Distribution Test |  |
| k star | 9.92 Data appear Gamma Distributed at 5\% Significance Level |  |
| Theta Star | 4.931 |  |
| MLE of Mean | 48.91 |  |
| MLE of Standard Deviation | 15.53 |  |
| nu star | 892.8 |  |
| A-D Test Statistic | 0.486 Nonparametric Statistics |  |
| 5\% A-D Critical Value | 0.748 90\% Percentile | 71.02 |
| K-S Test Statistic | 0.0977 95\% Percentile | 78.16 |
| $5 \% \mathrm{~K}$-S Critical Value | 0.132 99\% Percentile | 85 |
| Data appear Gamma Distributed at 5\% Significance Level |  |  |
| Assuming Gamma Distribution | 95\% UTL with 90\% Coverage | 75.5 |
| 90\% Percentile | 69.57 95\% Percentile Bootstrap UTL with 90\% Coverage | 75.5 |
| 95\% Percentile | 76.94 95\% BCA Bootstrap UTL with 90\% Coverage | 75.5 |
| 99\% Percentile | 92.07 95\% UPL | 78.16 |
|  | 95\% Chebyshev UPL | 116.6 |
| 95\% WH Approx. Gamma UPL | 77.41 Upper Threshold Limit Based upon IQR | 102.4 |
| 95\% HW Approx. Gamma UPL | 77.85 |  |
| 95\% WH Approx. Gamma UTL with 90\% Coverage | 76.66 |  |
| 95\% HW Approx. Gamma UTL with 90\% Coverage | 77.07 |  |



| Benzo(a)anthracene |  |  |  |
| :---: | :---: | :---: | :---: |
| General Statistics |  |  |  |
| Number of Valid Data |  | Number of Detected Data | 13 |
| Number of Distinct Detected Data |  | Number of Non-Detect Data | 8 |
|  |  | Percent Non-Detects | 38.10\% |
| Raw Statistics |  | Log-transformed Statistics |  |
| Minimum Detected |  | Minimum Detected | 3.761 |
| Maximum Detected |  | Maximum Detected | 6.163 |
| Mean of Detected | 191.2 | Mean of Detected | 5.1 |
| SD of Detected |  | SD of Detected | 0.609 |
| Minimum Non-Detect |  | Minimum Non-Detect | 5.886 |
| Maximum Non-Detect |  | Maximum Non-Detect | 6.064 |
| Data with Multiple Detection Limits |  | Single Detection Limit Scenario |  |
| Note: Data have multiple DLs - Use of KM Method is recommended |  | Number treated as Non-Detect with Single DL | 20 |
| For all methods (except KM, DL/2, and ROS Methods), |  | Number treated as Detected with Single DL | 1 |
| Observations < Largest ND are treated as NDs |  | Single DL Non-Detect Percentage | 95.24\% |
| Background Statistics |  |  |  |
| Normal Distribution Test with Detected Values Only |  | Lognormal Distribution Test with Detected Values Only |  |
| Shapiro Wilk Test Statistic | 0.88 | Shapiro Wilk Test Statistic | 0.94 |
| 5\% Shapiro Wilk Critical Value | 0.866 | 5\% Shapiro Wilk Critical Value | 0.866 |
| Data appear Normal at 5\% Significance Level |  | Data appear Lognormal at 5\% Significance Level |  |
| Assuming Normal Distribution |  | Assuming Lognormal Distribution |  |
| DL/2 Substitution Method |  | DL/2 Substitution Method |  |
| Mean |  | Mean (Log Scale) | 5.156 |
| SD | 84.05 | SD (Log Scale) | 0.479 |
| 95\% UTL 90\% Coverage | 351.1 | 95\% UTL 90\% Coverage | 432.3 |
| 95\% UPL (t) | 339.3 | 95\% UPL (t) | 404.3 |
| 90\% Percentile (z) | 298.7 | 90\% Percentile (z) | 320.7 |
| 95\% Percentile (z) | 329.2 | 95\% Percentile (z) | 381.7 |
| 99\% Percentile (z) | 386.5 | 99\% Percentile (z) | 529.1 |
| Maximum Likelihood Estimate(MLE) Method N/A | N/A | Log ROS Method |  |
|  |  | Mean in Original Scale | 177.2 |
|  |  | SD in Original Scale | 88.57 |
|  |  | Mean in Log Scale | 5.067 |
|  |  | SD in Log Scale | 0.496 |
|  |  | 95\% UTL 90\% Coverage | 408.2 |
|  |  | 95\% UPL (t) | 380.9 |
|  |  | 90\% Percentile (z) | 299.6 |
|  |  | 95\% Percentile (z) | 358.8 |
|  |  | 99\% Percentile (z) | 503 |
| Gamma Distribution Test with Detected Values Only k star (bias corrected) | Data Distribution Test with Detected Values Only |  |  |
|  | 2.677 | Data appear Normal at 5\% Significance Level |  |
| Theta Star | 71.4 |  |  |
| nu star | 69.61 |  |  |
| A-D Test Statistic | 0.348 | Nonparametric Statistics |  |
| 5\% A-D Critical Value | 0.738 | Kaplan-Meier (KM) Method |  |
| K-S Test Statistic | 0.166 | Mean | 182.1 |
| $5 \% \mathrm{~K}$-S Critical Value | 0.238 |  | 92.07 |
| Data appear Gamma Distributed at 5\% Significance Level | SE of Mean |  | 24.11 |
|  |  | 95\% KM UTL with 90\% Coverage | 357.5 |
| Assuming Gamma Distribution |  | 95\% KM Chebyshev UPL | 592.9 |
| Gamma ROS Statistics with Extrapolated Data |  | 95\% KM UPL (t) | 344.7 |
| Mean | 194.1 | 90\% Percentile (z) | 300.1 |
| Median | 200.8 | 95\% Percentile (z) | 333.6 |
| SD | 84.99 | 99\% Percentile (z) | 396.3 |
| k star | 4.551 |  |  |
| Theta star | 42.66 | Gamma ROS Limits with Extrapolated Data |  |
| Nu star | 191.1 | 95\% Wilson Hilferty (WH) Approx. Gamma UPL | 370.8 |
| 95\% Percentile of Chisquare (2k) | 17.06 | 95\% Hawkins Wixley (HW) Approx. Gamma UPL | 378.8 |
|  |  | 95\% WH Approx. Gamma UTL with 90\% Coverage | 389.6 |
| 90\% Percentile | 316 | 95\% HW Approx. Gamma UTL with 90\% Coverage | 399.4 |
| 95\% Percentile | 363.9 |  |  |
| 99\% Percentile | 465.5 |  |  |



| Benzo(b)fluoranthene |  |  |  |
| :---: | :---: | :---: | :---: |
| General Statistics |  |  |  |
| Number of Valid Data |  | Number of Detected Data | 16 |
| Number of Distinct Detected Data |  | Number of Non-Detect Data | 5 |
|  |  | Percent Non-Detects | 23.81\% |
| Raw Statistics |  | Log-transformed Statistics |  |
| Minimum Detected |  | Minimum Detected | 3.689 |
| Maximum Detected |  | Maximum Detected | 6.215 |
| Mean of Detected | 163.1 | Mean of Detected | 4.85 |
| SD of Detected | 120.2 | SD of Detected | 0.736 |
| Minimum Non-Detect |  | Minimum Non-Detect | 5.886 |
| Maximum Non-Detect |  | Maximum Non-Detect | 6.064 |
| Data with Multiple Detection Limits |  | Single Detection Limit Scenario |  |
| Note: Data have multiple DLs - Use of KM Method is recommended |  | Number treated as Non-Detect with Single DL | 20 |
| For all methods (except KM, DL/2, and ROS Methods), |  | Number treated as Detected with Single DL | 1 |
| Observations < Largest ND are treated as NDs |  | Single DL Non-Detect Percentage | 95.24\% |
| Background Statistics |  |  |  |
| Normal Distribution Test with Detected Values Only |  | Lognormal Distribution Test with Detected Values Only |  |
| Shapiro Wilk Test Statistic | 0.86 | Shapiro Wilk Test Statistic | 0.965 |
| 5\% Shapiro Wilk Critical Value | 0.887 | 5\% Shapiro Wilk Critical Value | 0.887 |
| Data not Normal at 5\% Significance Level |  | Data appear Lognormal at 5\% Significance Level |  |
| Assuming Normal Distribution | Assuming Lognormal Distribution |  |  |
| DL/2 Substitution Method | DL/2 Substitution Method |  |  |
| Mean | 170.7 | Mean (Log Scale) | 4.95 |
| SD | 105.2 SD (Log Scale) |  | 0.664 |
| 95\% UTL 90\% Coverage | 371.1 | 95\% UTL 90\% Coverage | 500.3 |
| 95\% UPL (t) | 356.4 | 95\% UPL (t) | 456 |
| 90\% Percentile (z) | 305.5 | 90\% Percentile (z) | 330.8 |
| 95\% Percentile (z) | 343.7 | 95\% Percentile (z) | 421 |
| 99\% Percentile (z) | 415.4 | 99\% Percentile (z) | 661.9 |
| Maximum Likelihood Estimate(MLE) Method N/A | N/A Log ROS Method |  |  |
|  |  | Mean in Original Scale | 152.8 |
|  |  | SD in Original Scale | 106.4 |
|  |  | Mean in Log Scale | 4.831 |
|  |  | SD in Log Scale | 0.646 |
|  |  | 95\% UTL 90\% Coverage | 429.1 |
|  |  | 95\% UPL (t) | 392.1 |
|  |  | 90\% Percentile (z) | 286.8 |
|  |  | 95\% Percentile (z) | 362.7 |
|  |  | 99\% Percentile (z) | 563.4 |
| Gamma Distribution Test with Detected Values Only |  | Data Distribution Test with Detected Values Only |  |
| k star (bias corrected) | 1.833 | Data appear Gamma Distributed at 5\% Significance Lev |  |
| Theta Star | 88.98 |  |  |
| nu star | 58.64 |  |  |
| A-D Test Statistic | 0.247 | Nonparametric Statistics |  |
| 5\% A-D Critical Value | 0.749 | Kaplan-Meier (KM) Method |  |
| K-S Test Statistic | 0.105 | Mean | 157.7 |
| $5 \% \mathrm{~K}$-S Critical Value | 0.218 |  | 109.2 |
| Data appear Gamma Distributed at 5\% Significance Level | el SE of Mean |  | 26.61 |
|  |  | 95\% KM UTL with 90\% Coverage | 365.7 |
| Assuming Gamma Distribution | 95\% KM Chebyshev UPL |  | 644.8 |
| Gamma ROS Statistics with Extrapolated Data | 95\% KM UPL (t) |  | 350.4 |
| Mean | 165.1 90\% Percentile (z) |  | 297.6 |
| Median | 160 95\% Percentile (z) |  | 337.3 |
| SD | 104.8 99\% Percentile (z) |  | 411.7 |
| k star | 2.442 |  |  |
| Theta star | 67.6 Gamma ROS Limits with Extrapolated Data |  |  |
| Nu star | 102.6 | 95\% Wilson Hilferty (WH) Approx. Gamma UPL | 378.9 |
| 95\% Percentile of Chisquare (2k) | 10.89 | 95\% Hawkins Wixley (HW) Approx. Gamma UPL | 388.3 |
|  |  | 95\% WH Approx. Gamma UTL with 90\% Coverage | 403.9 |
| 90\% Percentile | 306.6 | 95\% HW Approx. Gamma UTL with 90\% Coverage | 416 |
| 95\% Percentile | 368.2 |  |  |
| 99\% Percentile | 503.1 |  |  |


| Benzo(k)fluoranthene |  |  |  |
| :---: | :---: | :---: | :---: |
| General Statistics |  |  |  |
| Number of Valid Data |  | Number of Detected Data | 16 |
| Number of Distinct Detected Data |  | Number of Non-Detect Data | 5 |
|  |  | Percent Non-Detects | 23.81\% |
| Raw Statistics |  | Log-transformed Statistics |  |
| Minimum Detected |  | Minimum Detected | 3.714 |
| Maximum Detected |  | Maximum Detected | 6.064 |
| Mean of Detected | 166.5 | Mean of Detected | 4.873 |
| SD of Detected |  | SD of Detected | 0.764 |
| Minimum Non-Detect |  | Minimum Non-Detect | 5.886 |
| Maximum Non-Detect |  | Maximum Non-Detect | 6.064 |
| Data with Multiple Detection Limits |  | Single Detection Limit Scenario |  |
| Note: Data have multiple DLs - Use of KM Method is recommended |  | Number treated as Non-Detect with Single DL | 20 |
| For all methods (except KM, DL/2, and ROS Methods), |  | Number treated as Detected with Single DL | 1 |
| Observations < Largest ND are treated as NDs |  | Single DL Non-Detect Percentage | 95.24\% |
| Background Statistics |  |  |  |
| Normal Distribution Test with Detected Values Only |  | Lognormal Distribution Test with Detected Values Only |  |
| Shapiro Wilk Test Statistic | 0.923 | Shapiro Wilk Test Statistic | 0.918 |
| 5\% Shapiro Wilk Critical Value | 0.887 | 5\% Shapiro Wilk Critical Value | 0.887 |
| Data appear Normal at 5\% Significance Level |  | Data appear Lognormal at 5\% Significance Level |  |
| Assuming Normal Distribution |  | Assuming Lognormal Distribution |  |
| DL/2 Substitution Method |  | DL/2 Substitution Method |  |
| Mean | 172.1 | Mean (Log Scale) | 4.962 |
| SD | 96.03 | SD (Log Scale) | 0.682 |
| 95\% UTL 90\% Coverage |  | 95\% UTL 90\% Coverage | 523.8 |
| 95\% UPL (t) | 341.6 | 95\% UPL (t) | 476.2 |
| 90\% Percentile (z) | 295.2 | 90\% Percentile (z) | 342.4 |
| 95\% Percentile (z) | 330.1 | 95\% Percentile (z) | 438.6 |
| 99\% Percentile (z) | 395.5 | 99\% Percentile (z) | 698.2 |
| Maximum Likelihood Estimate(MLE) Method N/A | N/A | Log ROS Method |  |
|  |  | Mean in Original Scale | 156.9 |
|  |  | SD in Original Scale | 98.9 |
|  |  | Mean in Log Scale | 4.853 |
|  |  | SD in Log Scale | 0.681 |
|  |  | 95\% UTL 90\% Coverage | 469.3 |
|  |  | 95\% UPL (t) | 426.7 |
|  |  | 90\% Percentile (z) | 306.8 |
|  |  | 95\% Percentile (z) | 393 |
|  |  | 99\% Percentile (z) | 625.3 |
| Gamma Distribution Test with Detected Values Only |  | Data Distribution Test with Detected Values Only |  |
| k star (bias corrected) | 1.846 | Data appear Normal at 5\% Significance Level |  |
| Theta Star | 90.18 |  |  |
| nu star | 59.08 |  |  |
| A-D Test Statistic | 0.399 | Nonparametric Statistics |  |
| 5\% A-D Critical Value | 0.749 | Kaplan-Meier (KM) Method |  |
| K-S Test Statistic | 0.135 | Mean | 162.3 |
| $5 \% \mathrm{~K}$-S Critical Value | 0.218 |  | 102 |
| Data appear Gamma Distributed at 5\% Significance Level |  | SE of Mean | 25.37 |
|  |  | 95\% KM UTL with 90\% Coverage | 356.6 |
| Assuming Gamma Distribution |  | 95\% KM Chebyshev UPL | 617.2 |
| Gamma ROS Statistics with Extrapolated Data |  | 95\% KM UPL (t) | 342.3 |
| Mean | 168.6 | 90\% Percentile (z) | 293 |
| Median |  | 95\% Percentile (z) | 330 |
| SD | 96.97 | 99\% Percentile (z) | 399.5 |
| k star | 2.423 |  |  |
| Theta star | 69.6 | Gamma ROS Limits with Extrapolated Data |  |
| Nu star | 101.7 | 95\% Wilson Hilferty (WH) Approx. Gamma UPL | 389.1 |
| 95\% Percentile of Chisquare (2k) | 10.83 | 95\% Hawkins Wixley (HW) Approx. Gamma UPL | 402.3 |
|  |  | 95\% WH Approx. Gamma UTL with 90\% Coverage | 414.8 |
| 90\% Percentile | 313.7 | 95\% HW Approx. Gamma UTL with 90\% Coverage | 431.4 |
| 95\% Percentile | 376.9 |  |  |
| 99\% Percentile | 515.4 |  |  |



| Fluoranthene |  |  |
| :---: | :---: | :---: |
| General Statistics |  |  |
| Number of Valid Data | 21 Number of Detected Data | 20 |
| Number of Distinct Detected Data | 19 Number of Non-Detect Data | 1 |
|  | Percent Non-Detects | 4.76\% |
| Raw Statistics | Log-transformed Statistics |  |
| Minimum Detected | 39 Minimum Detected | 3.664 |
| Maximum Detected | 1005 Maximum Detected | 6.913 |
| Mean of Detected | 259.3 Mean of Detected | 5.115 |
| SD of Detected | 243.3 SD of Detected | 1.008 |
| Minimum Non-Detect | 390 Minimum Non-Detect | 5.966 |
| Maximum Non-Detect | 390 Maximum Non-Detect | 5.966 |
| Background Statistics |  |  |
| Normal Distribution Test with Detected Values Only | Lognormal Distribution Test with Detected Values Only |  |
| Shapiro Wilk Test Statistic | 0.819 Shapiro Wilk Test Statistic | 0.902 |
| 5\% Shapiro Wilk Critical Value | 0.905 5\% Shapiro Wilk Critical Value | 0.905 |
| Data not Normal at 5\% Significance Level | Data not Lognormal at 5\% Significance Level |  |
| Assuming Normal Distribution | Assuming Lognormal Distribution |  |
| DL/2 Substitution Method | DL/2 Substitution Method |  |
| Mean | 256.2 Mean (Log Scale) | 5.122 |
| SD | 237.5 SD (Log Scale) | 0.983 |
| 95\% UTL 90\% Coverage | 708.7 95\% UTL 90\% Coverage | 1090 |
| 95\% UPL (t) | 675.6 95\% UPL (t) | 950.4 |
| 90\% Percentile (z) | 560.7 90\% Percentile (z) | 590.8 |
| 95\% Percentile (z) | 646.9 95\% Percentile (z) | 844.3 |
| 99\% Percentile (z) | 808.8 99\% Percentile (z) | 1649 |
| Maximum Likelihood Estimate(MLE) Method | Log ROS Method |  |
| Mean | 118 Mean in Original Scale | 252.6 |
| SD | 366.2 SD in Original Scale | 239.1 |
| 95\% UTL with 90\% Coverage | 815.6 95\% UTL with 90\% Coverage | 1069 |
|  | 95\% BCA UTL with 90\% Coverage | 945.5 |
|  | 95\% Bootstrap (\%) UTL with 90\% Coverage | 959.5 |
| 95\% UPL (t) | 764.4 95\% UPL (t) | 931.6 |
| 90\% Percentile (z) | 587.3 90\% Percentile (z) | 578.5 |
| 95\% Percentile (z) | 720.3 95\% Percentile (z) | 827.4 |
| 99\% Percentile (z) | 969.9 99\% Percentile (z) | 1619 |
| Gamma Distribution Test with Detected Values Only |  |  |
| k star (bias corrected) | 1.112 Data do not follow a Discernable Distribution (0.05) |  |
| Theta Star | 233.2 ( |  |
| nu star | 44.47 |  |
| A-D Test Statistic | 0.836 Nonparametric Statistics |  |
| 5\% A-D Critical Value | 0.762 Kaplan-Meier (KM) Method |  |
| K-S Test Statistic | 0.232 Mean | 254.2 |
| 5\% K-S Critical Value | 0.198 SD | 233.9 |
| Data not Gamma Distributed at 5\% Significance Level | SE of Mean | 52.71 |
|  | 95\% KM UTL with 90\% Coverage | 699.9 |
| Assuming Gamma Distribution | 95\% KM Chebyshev UPL | 1298 |
| Gamma ROS Statistics with Extrapolated Data | 95\% KM UPL (t) | 667.2 |
| Mean | 256 90\% Percentile (z) | 554 |
| Median | 210 95\% Percentile (z) | 639 |
| SD | 237.6 99\% Percentile (z) | 798.5 |
| k star | 1.164 |  |
| Theta star | 220 Gamma ROS Limits with Extrapolated Data |  |
| Nu star | 48.88 95\% Wilson Hilferty (WH) Approx. Gamma UPL | 758.5 |
| 95\% Percentile of Chisquare (2k) | 6.611 95\% Hawkins Wixley (HW) Approx. Gamma UPL | 788.6 |
|  | 95\% WH Approx. Gamma UTL with 90\% Coverage | 826 |
| 90\% Percentile | 567.9 95\% HW Approx. Gamma UTL with 90\% Coverage | 866.6 |
| 95\% Percentile | 727.3 |  |
| 99\% Percentile | 1093 |  |



## APPENDIX C <br> STATISTICAL ANALYSIS

Table C. 1
Summary Statistics
4835 Glenbrook Road
Spring Valley, Washington, D.C.

| Depth | Chemical | N | \#D | \%D | Units | Mind | MaxD | Distribution | UCL Calculated Using ${ }^{1}$ | Central Tendency ${ }^{2}$ | UCL | RME |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 0-2 | Aluminum | 45 | 45 | 100\% | $\mathrm{mg} / \mathrm{kg}$ | 8960 | 35300 | None | 95\% Student's-t UCL | 21,000 | 23,116.14 | 23,116.14 |
| 0-10 | Aluminum | 99 | 98 | 99\% | $\mathrm{mg} / \mathrm{kg}$ | 8960 | 55900 | Kaplan-Meir | 95\% KM (BCA) UCL | 24,020.22 | 25,532.93 | 25,532.93 |
| 0-2 | Cobalt | 1 | 1 | 100\% | $\mathrm{mg} / \mathrm{kg}$ | 42 | 42 | NA | NA; 3 | 42 | -- | 42 |
| 0-10 | Cobalt | 3 | 3 | 100\% | $\mathrm{mg} / \mathrm{kg}$ | 18.4 | 42 | NA | NA; 3 | 28 | -- | 42 |
| 0-2 | Copper | 45 | 45 | 100\% | $\mathrm{mg} / \mathrm{kg}$ | 24.5 | 314 | Gamma | 95\% Approximate Gamma UCL | 70.35 | 79.24 | 79.24 |
| 0-10 | Copper | 99 | 99 | 100\% | $\mathrm{mg} / \mathrm{kg}$ | 16.2 | 444 | None | 95\% Chebyshev UCL | 62.50 | 107.56 | 107.56 |
| 0-2 | Manganese | 45 | 44 | 98\% | $\mathrm{mg} / \mathrm{kg}$ | 144 | 1200 | Kaplan-Meir | 95\% KM (t) UCL | 542.70 | 603.73 | 603.73 |
| 0-10 | Manganese | 99 | 98 | 99\% | $\mathrm{mg} / \mathrm{kg}$ | 133 | 4110 | Kaplan-Meir | 95\% KM (BCA) UCL | 669.40 | 772.96 | 772.96 |
| 0-2 | Mercury | 45 | 31 | 69\% | $\mathrm{mg} / \mathrm{kg}$ | 0.018 | 0.83 | Kaplan-Meir | 95\% KM (BCA) UCL | 0.11 | 0.15 | 0.15 |
| 0-10 | Mercury | 99 | 68 | 69\% | $\mathrm{mg} / \mathrm{kg}$ | 0.013 | 0.83 | Kaplan-Meir | 95\% KM (Percentile Bootstrap) UCL | 0.10 | 0.12 | 0.12 |
| 0-2 | Nickel | 45 | 45 | 100\% | $\mathrm{mg} / \mathrm{kg}$ | 16.3 | 345 | Gamma | 95\% Approximate Gamma UCL | 63.72 | 73.73 | 73.73 |
| 0-10 | Nickel | 99 | 99 | 100\% | $\mathrm{mg} / \mathrm{kg}$ | 12.3 | 345 | Gamma | 95\% Approximate Gamma UCL | 66.05 | 71.95 | 71.95 |
| 0-2 | Tellurium | 1 | 1 | 100\% | $\mathrm{mg} / \mathrm{kg}$ | 2.5 | 2.5 | NA | NA; 3 | 2.5 | -- | 2.5 |
| 0-10 | Tellurium | 3 | 3 | 100\% | $\mathrm{mg} / \mathrm{kg}$ | 2.2 | 6.6 | NA | NA; 3 | 3.77 | -- | 6.6 |
| 0-2 | Thallium | 45 | 18 | 40\% | $\mathrm{mg} / \mathrm{kg}$ | 0.55 | 2.6 | Kaplan-Meir | 95\% KM (t) UCL | 0.97 | 1.09 | 1.09 |
| 0-10 | Thallium | 99 | 34 | 34\% | $\mathrm{mg} / \mathrm{kg}$ | 0.55 | 8.7 | Kaplan-Meir | 95\% KM (t) UCL | 1.17 | 1.35 | 1.35 |
| 0-2 | Vanadium | 45 | 45 | 100\% | $\mathrm{mg} / \mathrm{kg}$ | 38.4 | 142 | None | 95\% Student's-t UCL | 83.80 | 94.26 | 94.26 |
| 0-10 | Vanadium | 99 | 99 | 100\% | $\mathrm{mg} / \mathrm{kg}$ | 33.2 | 345 | None | 95\% Student's-t UCL | 93.70 | 108.52 | 108.52 |

## Notes:

1 UCLs were calculated by ProUCL using the indicated technique
2 Value presented as the Central Tendency is determined by the distribution as follows:
Kaplan-Meier: the Kaplan-Meier mean
None: data is not parametrically distributed. The median is presented.
Lognormal: the backtransformed mean of the lognormal data
Gamma: k star * theta star
3 UCLs and Central Tendencies not calculated for datasets with less than ten samples [ $\mathrm{n}<10$ ] and/or less than 20 percent detections.
Definitions:
N Total number of samples analyzed
NA Not applicable
ND Number of non-detects
\%D Percentage of detects
MinD Minimum detected value
MaxD Maximum detected value
UCL Upper confidence limit
RME Reasonable maximum exposure

Table C. 2

## ProUCL Output for 0-2 feet bgs <br> 4835 Glenbrook Road Spring Valley, Washington, D.C.

| User Selected Options |  |
| :--- | :---: |
| Full Precision | ON |
| Confidence Coefficient | $95 \%$ |
| Number of Bootstrap Operations | 2000 |


| Aluminum |  |  |  |
| :---: | :---: | :---: | :---: |
| General Statistics |  |  |  |
| Number of Valid Observations | 45 | Number of Distinct Observations | 45 |
| Raw Statistics |  | Log-transformed Statistics |  |
| Minimum | 8960 | Minimum of Log Data | 9.1005255 |
| Maximum | 35300 | Maximum of Log Data | 10.471638 |
| Mean | 21408 | Mean of $\log$ Data | 9.9191812 |
| Median | 21000 S | SD of log Data | 0.3341177 |
| SD | 6819.642 |  |  |
| Coefficient of Variation | 0.318556 |  |  |
| Skewness | 0.299619 |  |  |
| Relevant UCL Statistics |  |  |  |
| Normal Distribution Test |  | Lognormal Distribution Test |  |
| Shapiro Wilk Test Statistic | 0.962085 | Shapiro Wilk Test Statistic | 0.9684656 |
| Shapiro Wilk Critical Value | 0.945 | Shapiro Wilk Critical Value | 0.945 |
| Data appear Normal at 5\% Significance Level Data appear Lognormal at 5\% Significance Level |  |  |  |
| Assuming Normal Distribution |  | Assuming Lognormal Distribution |  |
| 95\% Student's-t UCL | 23116.14 | 95\% H-UCL | 23523.569 |
| 95\% UCLs (Adjusted for Skewness) |  | 95\% Chebyshev (MVUE) UCL | 26217.785 |
| 95\% Adjusted-CLT UCL | 23128.7 | 97.5\% Chebyshev (MVUE) UCL | 28278.826 |
| 95\% Modified-t UCL | 23123.71 | 99\% Chebyshev (MVUE) UCL | 32327.341 |
| Gamma Distribution Test |  | Data Distribution |  |
| k star (bias corrected) | 9.083866 D | Data appear Normal at 5\% Significance Level |  |
| Theta Star | 2356.706 |  |  |
| MLE of Mean | 21408 |  |  |
| MLE of Standard Deviationnu star | 7102.982 |  |  |
|  | 817.548 |  |  |
| Approximate Chi Square Value (.05) | 752.1924 N | Nonparametric Statistics |  |
| Adjusted Level of Significance | 0.044667 | 95\% CLT UCL | 23080.178 |
| Adjusted Chi Square Value | 750.1263 | 95\% Jackknife UCL | 23116.142 |
|  |  | 95\% Standard Bootstrap UCL | 23083.989 |
| Anderson-Darling Test Statistic | 0.226324 | 95\% Bootstrap-t UCL | 23176.498 |
| Anderson-Darling 5\% Critical Value | 0.748663 | 95\% Hall's Bootstrap UCL | 23051.244 |
| Kolmogorov-Smirnov Test Statistic | 0.081499 | 95\% Percentile Bootstrap UCL | 23011.556 |
| Kolmogorov-Smirnov 5\% Critical Value <br> Data appear Gamma Distributed at 5\% Significance Level | 0.131718 | 95\% BCA Bootstrap UCL | 23204.889 |
|  |  | 95\% Chebyshev(Mean, Sd) UCL | 25839.31 |
|  |  | 97.5\% Chebyshev(Mean, Sd) UCL | 27756.741 |
| Assuming Gamma Distribution 95\% Approximate Gamma UCL 95\% Adjusted Gamma UCL |  | 99\% Chebyshev(Mean, Sd) UCL | 31523.164 |
|  | 23268.07 |  |  |
|  | 23332.16 |  |  |
|  |  | Potential UCL to Use Use 95\% Student's-t UCL |  |
|  |  | Use 95\% Student's-t UCL | 23116.142 |


|  | Cobalt |
| :--- | :--- |
| General Statistics <br> Number of Valid Observations <br>  <br> Warning: This data set only has 1 observations! <br> Data set is too small to compute reliable and meaningful statistics and estimates! <br> The data set for variable Cobalt was not processed! <br> It is suggested to collect at least 8 to 10 observations before using these statistical methods! <br> If possible, compute and collect Data Quality Objectives (DQO) based sample size and analytical results. |  |



| Manganese |  |  |  |
| :---: | :---: | :---: | :---: |
| General Statistics |  |  |  |
| Number of Valid Data |  | Number of Detected Data | 44 |
| Number of Distinct Detected Data |  | Number of Non-Detect Data | 1 |
|  |  | Percent Non-Detects | 2.22\% |
| Raw Statistics |  | Log-transformed Statistics |  |
| Minimum Detected | 144 | Minimum Detected | 4.9698133 |
| Maximum Detected | 1200 | Maximum Detected | 7.0900768 |
| Mean of Detected | 542.7046 | Mean of Detected | 6.1937259 |
| SD of Detected | 240.9209 | SD of Detected | 0.4753468 |
| Minimum Non-Detect | 1290 | Minimum Non-Detect | 7.1623975 |
| Maximum Non-Detect | 1290 | Maximum Non-Detect | 7.1623975 |
| UCL Statistics |  |  |  |
| Normal Distribution Test with Detected Values Only |  | Lognormal Distribution Test with Detected Values Only |  |
| Shapiro Wilk Test Statistic | 0.954982 | Shapiro Wilk Test Statistic | 0.9729495 |
| 5\% Shapiro Wilk Critical Value | 0.944 | 5\% Shapiro Wilk Critical Value | 0.944 |
| Data appear Normal at 5\% Significance Level |  | Data appear Lognormal at 5\% Significance Level |  |
| Assuming Normal Distribution |  | Assuming Lognormal Distribution |  |
| DL/2 Substitution Method |  | DL/2 Substitution Method |  |
| Mean | 544.9778 | Mean | 6.1998487 |
| SD | 238.6551 |  | 0.4717057 |
| 95\% DL/2 (t) UCL | 604.7547 | 95\% H-Stat (DL/2) UCL | 635.18829 |
| Maximum Likelihood Estimate(MLE) Method MLE method failed to converge properly | N/A | Log ROS Method |  |
|  |  | Mean in Log Scale | 6.1937259 |
|  |  | SD in Log Scale | 0.4699141 |
|  |  | Mean in Original Scale | 541.52594 |
|  |  | SD in Original Scale | 238.29859 |
|  |  | 95\% Percentile Bootstrap UCL | 601.86667 |
|  |  | 95\% BCA Bootstrap UCL | 605.40743 |
| Gamma Distribution Test with Detected Values Only |  | Data Distribution Test with Detected Values Only |  |
| k star (bias corrected) | 4.695366 | Data appear Normal at 5\% Significance Level |  |
| Theta Star | 115.583 |  |  |
| nu star | 413.1922 |  |  |
| A-D Test Statistic | 0.23837 | Nonparametric Statistics |  |
| 5\% A-D Critical Value | 0.752847 | Kaplan-Meier (KM) Method |  |
| K-S Test Statistic | 0.752847 | Mean | 542.70455 |
| $5 \%$ K-S Critical Value | 0.13372 | SE of Mean | 238.1674 |
| Data appear Gamma Distributed at 5\% Significance Level |  |  | 36.320187 |
|  |  | 95\% KM (t) UCL | 603.73081 |
| Assuming Gamma Distribution |  | 95\% KM (z) UCL | 602.44594 |
| Gamma ROS Statistics using Extrapolated Data |  | 95\% KM (jackknife) UCL | 603.74658 |
| Minimum |  | 95\% KM (bootstrap t) UCL | 611.27581 |
| Maximum | 1200 | 95\% KM (BCA) UCL | 602.81818 |
| Mean | 543.7704 | 95\% KM (Percentile Bootstrap) UCL | 602.06667 |
| Median | 530 | 95\% KM (Chebyshev) UCL | 701.02057 |
| SD | 238.2747 | 97.5\% KM (Chebyshev) UCL | 769.52404 |
| k star | 4.802202 | 99\% KM (Chebyshev) UCL | 904.08585 |
| Theta star | 113.2335 |  |  |
| Nu star | 432.1982 P | Potential UCLs to Use |  |
| AppChi2 | 385.0019 | 95\% KM (t) UCL | 603.73081 |
| 95\% Gamma Approximate UCL | 610.4296 | 95\% KM (Percentile Bootstrap) UCL | 602.06667 |
| 95\% Adjusted Gamma UCL | 612.7685 |  |  |
| Note: DL/2 is not a recommended method. |  |  |  |


| Mercury |  |  |  |
| :---: | :---: | :---: | :---: |
| General Statistics |  |  |  |
| Number of Valid Data |  | Number of Detected Data | 31 |
| Number of Distinct Detected Data |  | Number of Non-Detect Data | 14 |
|  |  | Percent Non-Detects | 31.11\% |
| Raw Statistics |  | Log-transformed Statistics |  |
| Minimum Detected | 0.018 | Minimum Detected | -4.017384 |
| Maximum Detected | 0.83 | Maximum Detected | -0.18633 |
| Mean of Detected | 0.131226 | Mean of Detected | -2.35406 |
| SD of Detected | 0.14347 | SD of Detected | 0.7988115 |
| Minimum Non-Detect | 0.076 | Minimum Non-Detect | -2.577022 |
| Maximum Non-Detect |  | Maximum Non-Detect | -2.207275 |
| Note: Data have multiple DLs - Use of KM Method is reco | ommended | Number treated as Non-Detect | 30 |
| For all methods (except KM, DL/2, and ROS Methods), |  | Number treated as Detected | 15 |
| Observations < Largest ND are treated as NDs |  | Single DL Non-Detect Percentage | 66.67\% |
| UCL Statistics |  |  |  |
| Normal Distribution Test with Detected Values Only |  | Lognormal Distribution Test with Detected Values Only |  |
| Shapiro Wilk Test Statistic | 0.570456 | Shapiro Wilk Test Statistic | 0.9488993 |
| 5\% Shapiro Wilk Critical Value | 0.929 | 5\% Shapiro Wilk Critical Value | 0.929 |
| Data not Normal at 5\% Significance Level |  | Data appear Lognormal at 5\% Significance Level |  |
| Assuming Normal Distribution |  | Assuming Lognormal Distribution |  |
| DL/2 Substitution Method |  | DL/2 Substitution Method |  |
| Mean | 0.105656 | Mean | -2.561747 |
| SD | 0.124596 |  | 0.7328014 |
| 95\% DL/2 (t) UCL | 0.136864 | 95\% H-Stat (DL/2) UCL | 0.1300145 |
| Maximum Likelihood Estimate(MLE) Method | N/A | Log ROS Method |  |
| MLE yields a negative mean |  | Mean in Log Scale | -2.565648 |
|  |  | SD in Log Scale | 0.7507258 |
|  |  | Mean in Original Scale | 0.1060296 |
|  |  | SD in Original Scale | 0.1246754 |
|  |  | 95\% Percentile Bootstrap UCL | 0.1399289 |
|  |  | 95\% BCA Bootstrap UCL | 0.1522572 |
| Gamma Distribution Test with Detected Values Only |  | Data Distribution Test with Detected Values Only |  |
| k star (bias corrected) | 1.551966 | Data appear Lognormal at 5\% Significance Level |  |
| Theta Star | 0.084555 |  |  |
| nu star | 96.22187 |  |  |
| A-D Test Statistic | 0.830296 | Nonparametric Statistics |  |
| 5\% A-D Critical Value | 0.761787 | Kaplan-Meier (KM) Method |  |
| K-S Test Statistic | 0.761787 | Mean | 0.1071976 |
| 5\% K-S Critical Value | 0.1603 |  | 0.1234288 |
| Data not Gamma Distributed at 5\% Significance Level | SE of Mean |  | 0.0189308 |
|  |  | 95\% KM (t) UCL | 0.1390056 |
| Assuming Gamma Distribution |  | 95\% KM (z) UCL | 0.1383359 |
| Gamma ROS Statistics using Extrapolated Data |  | 95\% KM (jackknife) UCL | 0.1388959 |
| Minimum | 0.018 | 95\% KM (bootstrap t) UCL | 0.1645956 |
| Maximum | 0.83 | 95\% KM (BCA) UCL | 0.1459022 |
| Mean | 0.127384 | 95\% KM (Percentile Bootstrap) UCL | 0.1405671 |
| Median | 0.112889 | 95\% KM (Chebyshev) UCL | 0.1897149 |
| SD | 0.121665 | 97.5\% KM (Chebyshev) UCL | 0.2254202 |
| k star | 2.009646 | 99\% KM (Chebyshev) UCL | 0.2955563 |
| Theta star | 0.063386 |  |  |
| Nu star | 180.8681 | Potential UCLs to Use |  |
| AppChi2 | 150.7616 | 95\% KM (BCA) UCL | 0.1459022 |
| 95\% Gamma Approximate UCL | 0.152822 |  |  |
| 95\% Adjusted Gamma UCL | 0.153748 |  |  |
| Note: DL/2 is not a recommended method. |  |  |  |



| Thallium |  |  |  |
| :---: | :---: | :---: | :---: |
| General Statistics |  |  |  |
| Number of Valid Data |  | Number of Detected Data | 18 |
| Number of Distinct Detected Data |  | Number of Non-Detect Data | 27 |
|  |  | Percent Non-Detects | 60.00\% |
| Raw Statistics |  | Log-transformed Statistics |  |
| Minimum Detected |  | Minimum Detected | -0.597837 |
| Maximum Detected |  | Maximum Detected | 0.9555114 |
| Mean of Detected | 1.033333 | Mean of Detected | -0.035525 |
| SD of Detected | 0.455748 | SD of Detected | 0.3613424 |
| Minimum Non-Detect |  | Minimum Non-Detect | -0.510826 |
| Maximum Non-Detect |  | Maximum Non-Detect | 3.0955776 |
| Note: Data have multiple DLs - Use of KM Method is recom | mmended | Number treated as Non-Detect | 45 |
| For all methods (except KM, DL/2, and ROS Methods), |  | Number treated as Detected | 0 |
| Observations < Largest ND are treated as NDs |  | Single DL Non-Detect Percentage | 100.00\% |
| UCL Statistics |  |  |  |
| Normal Distribution Test with Detected Values Only |  | Lognormal Distribution Test with Detected Values Only |  |
| Shapiro Wilk Test Statistic | 0.748122 | Shapiro Wilk Test Statistic | 0.9217477 |
| 5\% Shapiro Wilk Critical Value | 0.897 | 5\% Shapiro Wilk Critical Value | 0.897 |
| Data not Normal at 5\% Significance Level |  | Data appear Lognormal at 5\% Significance Level |  |
| Assuming Normal Distribution |  | Assuming Lognormal Distribution |  |
| DL/2 Substitution Method |  | DL/2 Substitution Method |  |
| Mean | 1.712222 | Mean | 0.233133 |
| SD | 2.153943 | SD | 0.6527997 |
| 95\% DL/2 (t) UCL | 2.251729 | 95\% H-Stat (DL/2) UCL | 2.3302176 |
| Maximum Likelihood Estimate(MLE) Method N/A | N/A | Log ROS Method |  |
| MLE method failed to converge properly |  | Mean in Log Scale | -0.08406 |
|  |  | SD in Log Scale | 0.2693212 |
|  |  | Mean in Original Scale | 0.9560966 |
|  |  | SD in Original Scale | 0.3158941 |
|  |  | 95\% Percentile Bootstrap UCL | 1.0358092 |
|  |  | 95\% BCA Bootstrap UCL | 1.0699403 |
| Gamma Distribution Test with Detected Values Only |  | Data Distribution Test with Detected Values Only |  |
| k star (bias corrected) | 6.271909 | Data appear Gamma Distributed at 5\% Significance Level |  |
| Theta Star | 0.164756 |  |  |
| nu star | 225.7887 |  |  |
| A-D Test Statistic | 0.659275 | Nonparametric Statistics |  |
| 5\% A-D Critical Value | 0.741099 | Kaplan-Meier (KM) Method |  |
| K-S Test Statistic | 0.741099 | Mean | 0.9651351 |
| 5\% K-S Critical Value | 0.203772 |  | 0.3636573 |
| Data appear Gamma Distributed at 5\% Significance Level | l SE of Mean |  | 0.0737865 |
|  |  | 95\% KM (t) UCL | 1.0891135 |
| Assuming Gamma Distribution |  | 95\% KM (z) UCL | 1.0865032 |
| Gamma ROS Statistics using Extrapolated Data |  | 95\% KM (jackknife) UCL | 1.0897951 |
| Minimum | 0.459038 | 95\% KM (bootstrap t) UCL | 1.1245251 |
| Maximum | 2.6 | 95\% KM (BCA) UCL | 1.0974802 |
| Mean | 1.053792 | 95\% KM (Percentile Bootstrap) UCL | 1.0882838 |
| Median |  | 95\% KM (Chebyshev) UCL | 1.2867631 |
| SD | 0.323377 | 97.5\% KM (Chebyshev) UCL | 1.4259318 |
| k star | 11.66712 | 99\% KM (Chebyshev) UCL | 1.6993017 |
| Theta star | 0.090322 |  |  |
| Nu star | 1050.04 | Potential UCLs to Use |  |
| AppChi2 | 975.8163 | 95\% KM (t) UCL | 1.0891135 |
| 95\% Gamma Approximate UCL | 1.133947 |  |  |
| 95\% Adjusted Gamma UCL | 1.136693 |  |  |
| Note: DL/2 is not a recommended method. |  |  |  |


| Vanadium |  |  |  |
| :---: | :---: | :---: | :---: |
| General Statistics |  |  |  |
| Number of Valid Observations |  | Number of Distinct Observations | 42 |
| Raw Statistics |  | Log-transformed Statistics |  |
| Minimum | 38.4 | Minimum of Log Data | 3.6480575 |
| Maximum | 142 | Maximum of Log Data | 4.9558271 |
| Mean | 87.68667 | Mean of log Data | 4.4260527 |
| Median | 83.8 | SD of log Data | 0.321628 |
| SD | 26.246 |  |  |
| Coefficient of Variation | 0.299316 |  |  |
| Skewness | 0.128221 |  |  |
| Relevant UCL Statistics |  |  |  |
| Normal Distribution Test |  | Lognormal Distribution Test |  |
| Shapiro Wilk Test Statistic | 0.973779 | Shapiro Wilk Test Statistic | 0.9598491 |
| Shapiro Wilk Critical Value | 0.945 | Shapiro Wilk Critical Value | 0.945 |
| Data appear Normal at 5\% Significance Level |  | Data appear Lognormal at 5\% Significance Level |  |
| Assuming Normal Distribution |  | Assuming Lognormal Distribution |  |
| 95\% Student's-t UCL | 94.26061 | 95\% H-UCL | 96.038399 |
| 95\% UCLs (Adjusted for Skewness) |  | 95\% Chebyshev (MVUE) UCL | 106.69515 |
| 95\% Adjusted-CLT UCL | 94.2021 | 97.5\% Chebyshev (MVUE) UCL | 114.81374 |
| 95\% Modified-t UCL | 94.27307 | 99\% Chebyshev (MVUE) UCL | 130.76114 |
| Gamma Distribution Test |  | Data Distribution |  |
| k star (bias corrected) | 9.947688 | Data appear Normal at 5\% Significance Level |  |
| Theta Star | 8.814779 |  |  |
| MLE of Mean | 87.68667 |  |  |
| MLE of Standard Deviation | 27.80177 |  |  |
| nu star | 895.2919 |  |  |
| Approximate Chi Square Value (.05) | 826.8449 | Nonparametric Statistics |  |
| Adjusted Level of Significance | 0.044667 | 95\% CLT UCL | 94.122194 |
| Adjusted Chi Square Value | 824.6771 | 95\% Jackknife UCL | 94.260605 |
|  |  | 95\% Standard Bootstrap UCL | 93.944818 |
| Anderson-Darling Test Statistic | 0.261388 | 95\% Bootstrap-t UCL | 94.257512 |
| Anderson-Darling 5\% Critical Value | 0.74838 | 95\% Hall's Bootstrap UCL | 94.398006 |
| Kolmogorov-Smirnov Test Statistic | 0.091076 | 95\% Percentile Bootstrap UCL | 94.151111 |
| Kolmogorov-Smirnov 5\% Critical Value | 0.13168 | 95\% BCA Bootstrap UCL | 93.986667 |
| Data appear Gamma Distributed at 5\% Significance Level |  | 95\% Chebyshev(Mean, Sd) UCL | 104.74096 |
|  |  | 97.5\% Chebyshev(Mean, Sd) UCL | 112.12036 |
| Assuming Gamma Distribution |  | 99\% Chebyshev(Mean, Sd) UCL | 126.61578 |
| 95\% Approximate Gamma UCL | 94.94545 |  |  |
| 95\% Adjusted Gamma UCL | 95.19503 |  |  |
|  | Potential UCL to Use |  |  |
|  |  | Use 95\% Student's-t UCL | 94.260605 |

Table C. 3

## ProUCL Output for 0-10 feet bgs <br> 4835 Glenbrook Road Spring Valley, Washington, D.C.



|  | Cobalt |
| :--- | :--- |
| General Statistics <br> Number of Valid Observations <br>  <br> Warning: This data set only has 3 observations! <br> Data set is too small to compute reliable and meaningful statistics and estimates! <br> The data set for variable Cobalt was not processed! <br> It is suggested to collect at least 8 to 10 observations before using these statistical methods! <br> If possible, compute and collect Data Quality Objectives (DQO) based sample size and analytical results. |  |


| Copper |  |  |  |
| :---: | :---: | :---: | :---: |
| General Statistics |  |  |  |
| Number of Valid Observations |  | Number of Distinct Observations | 95 |
| Raw Statistics |  | Log-transformed Statistics |  |
| Minimum | 16.2 | Minimum of Log Data | 2.785011 |
| Maximum | 444 | Maximum of Log Data | 6.095825 |
| Mean | 78.72929 | Mean of log Data | 4.16272 |
| Median | 62.5 | SD of log Data | 0.596416 |
| SD | 65.80122 |  |  |
| Coefficient of Variation | 0.835791 |  |  |
| Skewness | 3.26125 |  |  |
| Relevant UCL Statistics |  |  |  |
| Normal Distribution Test |  | Lognormal Distribution Test |  |
| Lilliefors Test Statistic | 0.240049 | Lilliefors Test Statistic | 0.09358 |
| Lilliefors Critical Value | 0.089046 | Lilliefors Critical Value | 0.089046 |
| Data not Normal at 5\% Significance Level Data not Lognormal at 5\% Significance Level |  |  |  |
| Assuming Normal Distribution |  | Assuming Lognormal Distribution |  |
| 95\% Student's-t UCL | 89.71097 | 95\% H-UCL | 86.14791 |
| 95\% UCLs (Adjusted for Skewness) |  | 95\% Chebyshev (MVUE) UCL | 98.1514 |
| 95\% Adjusted-CLT UCL | 91.92329 | 97.5\% Chebyshev (MVUE) UCL | 107.4809 |
| 95\% Modified-t UCL | 90.07224 | 99\% Chebyshev (MVUE) UCL | 125.8069 |
| Gamma Distribution Test |  | Data Distribution |  |
| k star (bias corrected) 2.541646 |  | Data do not follow a Discernable Distribution (0.05) |  |
| Theta Star | 30.97572 |  |  |
| MLE of Mean | 78.72929 |  |  |
| MLE of Standard Deviation | 49.38316 |  |  |
| nu star | 503.2458 |  |  |
| Approximate Chi Square Value (.05) | 452.224 | Nonparametric Statistics |  |
| Adjusted Level of Significance | 0.047576 | 95\% CLT UCL | 89.60716 |
| Adjusted Chi Square Value | 451.5162 | 95\% Jackknife UCL | 89.71097 |
|  |  | 95\% Standard Bootstrap UCL | 89.74498 |
| Anderson-Darling Test Statistic | 3.229802 | 95\% Bootstrap-t UCL | 93.41967 |
| Anderson-Darling 5\% Critical Value | 0.761093 | 95\% Hall's Bootstrap UCL | 93.75702 |
| Kolmogorov-Smirnov Test Statistic | 0.142362 | 95\% Percentile Bootstrap UCL | 89.73939 |
| Kolmogorov-Smirnov 5\% Critical Value | 0.090771 | 95\% BCA Bootstrap UCL | 92.15758 |
| Data not Gamma Distributed at 5\% Significance Level | 95\% Chebyshev(Mean, Sd) UCL |  | 107.5559 |
|  | 97.5\% Chebyshev(Mean, Sd) UCL |  | 120.0292 |
| Assuming Gamma Distribution |  | 99\% Chebyshev(Mean, Sd) UCL | 144.5305 |
| 95\% Approximate Gamma UCL | 87.61188 |  |  |
| 95\% Adjusted Gamma UCL | 87.74921 |  |  |
|  | Potential UCL to Use |  |  |
|  |  | Use 95\% Chebyshev (Mean, Sd) UCL | 107.5559 |


| Manganese |  |  |  |
| :---: | :---: | :---: | :---: |
| General Statistics |  |  |  |
| Number of Valid Data |  | Number of Detected Data | 98 |
| Number of Distinct Detected Data |  | Number of Non-Detect Data | 1 |
|  |  | Percent Non-Detects | 1.01\% |
| Raw Statistics |  | Log-transformed Statistics |  |
| Minimum Detected | 133 | Minimum Detected | 4.890349 |
| Maximum Detected | 4110 | Maximum Detected | 8.321178 |
| Mean of Detected | 670.4388 | Mean of Detected | 6.318089 |
| SD of Detected | 516.5584 | SD of Detected | 0.605222 |
| Minimum Non-Detect | 1290 | Minimum Non-Detect | 7.162398 |
| Maximum Non-Detect | 1290 | Maximum Non-Detect | 7.162398 |
| UCL Statistics |  |  |  |
| Normal Distribution Test with Detected Values Only |  | Lognormal Distribution Test with Detected Values Only |  |
| Lilliefors Test Statistic | 0.194194 | Lilliefors Test Statistic | 0.075841 |
| 5\% Lilliefors Critical Value | 0.0895 | 5\% Lilliefors Critical Value | 0.0895 |
| Data not Normal at 5\% Significance Level |  | Data appear Lognormal at 5\% Significance Level |  |
| Assuming Normal Distribution |  | Assuming Lognormal Distribution |  |
| DL/2 Substitution Method |  | DL/2 Substitution Method |  |
| Mean | 670.1818 | Mean | 6.319616 |
| SD | 513.9225 |  | 0.602318 |
| 95\% DL/2 (t) UCL | 755.9512 | 95\% H-Stat (DL/2) UCL | 755.5591 |
| Maximum Likelihood Estimate(MLE) Method | N/A | Log ROS Method |  |
| MLE yields a negative mean |  | Mean in Log Scale | 6.317513 |
|  |  | SD in Log Scale | 0.602153 |
|  |  | Mean in Original Scale | 668.9574 |
|  |  | SD in Original Scale | 514.1274 |
|  |  | 95\% Percentile Bootstrap UCL | 759.4647 |
|  |  | 95\% BCA Bootstrap UCL | 773.6139 |
| Gamma Distribution Test with Detected Values Only |  | Data Distribution Test with Detected Values Only |  |
| k star (bias corrected) | 2.710627 | Data appear Lognormal at 5\% Significance Level |  |
| Theta Star | 247.3371 |  |  |
| nu star | 531.283 |  |  |
| A-D Test Statistic | 1.148656 | Nonparametric Statistics |  |
| 5\% A-D Critical Value | 0.760018 | Kaplan-Meier (KM) Method |  |
| K-S Test Statistic | 0.760018 | Mean | 669.4002 |
| 5\% K-S Critical Value | 0.091152 |  | 512.0504 |
| Data not Gamma Distributed at 5\% Significance Level |  | SE of Mean | 51.79216 |
|  |  | 95\% KM (t) UCL | 755.4037 |
| Assuming Gamma Distribution |  | 95\% KM (z) UCL | 754.5907 |
| Gamma ROS Statistics using Extrapolated Data |  | 95\% KM (jackknife) UCL | 755.4016 |
| Minimum | 133 | 95\% KM (bootstrap t) UCL | 789.4754 |
| Maximum | 4110 | 95\% KM (BCA) UCL | 772.9605 |
| Mean | 670.7939 | 95\% KM (Percentile Bootstrap) UCL | 760.261 |
| Median | 596 | 95\% KM (Chebyshev) UCL | 895.157 |
| SD | 513.9283 | 97.5\% KM (Chebyshev) UCL | 992.8421 |
| k star | 2.737411 | 99\% KM (Chebyshev) UCL | 1184.726 |
| Theta star | 245.0468 |  |  |
| Nu star | 542.0074 | Potential UCLs to Use |  |
| AppChi2 | 489.0122 | 95\% KM (BCA) UCL | 772.9605 |
| 95\% Gamma Approximate UCL | 743.4891 |  |  |
| 95\% Adjusted Gamma UCL | 744.6105 |  |  |
| Note: DL/2 is not a recommended method. |  |  |  |




| Thallium |  |  |  |
| :---: | :---: | :---: | :---: |
| General Statistics |  |  |  |
| Number of Valid Data |  | Number of Detected Data | 34 |
| Number of Distinct Detected Data |  | Number of Non-Detect Data | 65 |
|  |  | Percent Non-Detects | 65.66\% |
| Raw Statistics |  | Log-transformed Statistics |  |
| Minimum Detected | 0.55 | Minimum Detected | -0.59784 |
| Maximum Detected |  | Maximum Detected | 2.163323 |
| Mean of Detected | 1.428529 | Mean of Detected | 0.160436 |
| SD of Detected | 1.407099 | SD of Detected | 0.53699 |
| Minimum Non-Detect |  | Minimum Non-Detect | -0.51083 |
| Maximum Non-Detect |  | Maximum Non-Detect | 3.152736 |
| Note: Data have multiple DLs - Use of KM Method is recommended |  | Number treated as Non-Detect | 99 |
| For all methods (except KM, DL/2, and ROS Methods), |  | Number treated as Detected | 0 |
| Observations < Largest ND are treated as NDs |  | Single DL Non-Detect Percentage | 100.00\% |
| UCL Statistics |  |  |  |
| Normal Distribution Test with Detected Values Only |  | Lognormal Distribution Test with Detected Values Only |  |
| Shapiro Wilk Test Statistic | 0.490149 | Shapiro Wilk Test Statistic | 0.854784 |
| 5\% Shapiro Wilk Critical Value | 0.933 | 5\% Shapiro Wilk Critical Value | 0.933 |
| Data not Normal at 5\% Significance Level |  | Data not Lognormal at 5\% Significance Level |  |
| Assuming Normal Distribution |  | Assuming Lognormal Distribution |  |
| DL/2 Substitution Method |  | DL/2 Substitution Method |  |
| Mean | 2.006768 | Mean | 0.373055 |
| SD | 2.401948 | SD | 0.677475 |
| 95\% DL/2 (t) UCL | 2.407633 | 95\% H-Stat (DL/2) UCL | 2.3586 |
| Maximum Likelihood Estimate(MLE) Method | N/A | Log ROS Method |  |
| MLE method failed to converge properly |  | Mean in Log Scale | 0.060957 |
|  |  | SD in Log Scale | 0.393505 |
|  |  | Mean in Original Scale | 1.178821 |
|  |  | SD in Original Scale | 0.86778 |
|  |  | 95\% Percentile Bootstrap UCL | 1.337166 |
|  |  | 95\% BCA Bootstrap UCL | 1.411555 |
| Gamma Distribution Test with Detected Values Only |  |  |  |
| k star (bias corrected) | 2.484454 | Data do not follow a Discernable Distribution (0.05) |  |
| Theta Star | 0.574987 |  |  |
| nu star | 168.9428 |  |  |
| A-D Test Statistic | 2.651085 | Nonparametric Statistics |  |
| 5\% A-D Critical Value | 0.755213 | Kaplan-Meier (KM) Method |  |
| K-S Test Statistic | 0.755213 | Mean | 1.171218 |
| 5\% K-S Critical Value | 0.152248 |  | 0.909707 |
| Data not Gamma Distributed at 5\% Significance Level |  | SE of Mean | 0.105357 |
|  |  | 95\% KM (t) UCL | 1.346168 |
| Assuming Gamma Distribution |  | 95\% KM (z) UCL | 1.344514 |
| Gamma ROS Statistics using Extrapolated Data |  | 95\% KM (jackknife) UCL | 1.345439 |
| Minimum | 0.474393 | 95\% KM (bootstrap t) UCL | 1.441529 |
| Maximum | 8.7 | 95\% KM (BCA) UCL | 1.376461 |
| Mean | 1.454444 | 95\% KM (Percentile Bootstrap) UCL | 1.351188 |
| Median | 1.425742 | 95\% KM (Chebyshev) UCL | 1.630456 |
| SD | 0.864347 | 97.5\% KM (Chebyshev) UCL | 1.829169 |
| k star | 5.495104 | 99\% KM (Chebyshev) UCL | 2.219501 |
| Theta star | 0.26468 |  |  |
| Nu star | 1088.031 | Potential UCLs to Use |  |
| AppChi2 | 1012.455 | 95\% KM (t) UCL | 1.346168 |
| 95\% Gamma Approximate UCL | 1.563013 | 95\% KM (\% Bootstrap) UCL | 1.351188 |
| 95\% Adjusted Gamma UCL | 1.564659 |  |  |
| Note: DL/2 is not a recommended method. |  |  |  |


| Vanadium |  |  |  |
| :---: | :---: | :---: | :---: |
| General Statistics |  |  |  |
| Number of Valid Observations |  | Number of Distinct Observations | 84 |
| Raw Statistics |  | Log-transformed Statistics |  |
| Minimum | 33.2 | Minimum of Log Data | 3.50255 |
| Maximum | 345 | Maximum of Log Data | 5.843544 |
| Mean | 100.8253 | Mean of log Data | 4.535009 |
| Median | 93.7 | SD of log Data | 0.383957 |
| SD | 46.12171 |  |  |
| Coefficient of Variation | 0.457442 |  |  |
| Skewness | 2.518952 |  |  |
| Relevant UCL Statistics |  |  |  |
| Normal Distribution Test |  | Lognormal Distribution Test |  |
| Lilliefors Test Statistic | 0.158535 | Lilliefors Test Statistic | 0.071974 |
| Lilliefors Critical Value | 0.089046 | Lilliefors Critical Value | 0.089046 |
| Data not Normal at 5\% Significance Level |  | Data appear Lognormal at 5\% Significance Level |  |
| Assuming Normal Distribution |  | Assuming Lognormal Distribution |  |
| 95\% Student's-t UCL | 108.5226 | 95\% H-UCL | 107.559 |
| 95\% UCLs (Adjusted for Skewness) |  | 95\% Chebyshev (MVUE) UCL | 117.7015 |
| 95\% Adjusted-CLT UCL | 109.7037 | 97.5\% Chebyshev (MVUE) UCL | 125.2417 |
| 95\% Modified-t UCL | 108.7182 | 99\% Chebyshev (MVUE) UCL | 140.0529 |
| Gamma Distribution Test |  | Data Distribution |  |
| k star (bias corrected) | 6.349875 | Data appear Lognormal at 5\% Significance Level |  |
| Theta Star | 15.8783 |  |  |
| MLE of Mean | 100.8253 |  |  |
| MLE of Standard Deviation | 40.01167 |  |  |
| nu star | 1257.275 |  |  |
| Approximate Chi Square Value (.05) | 1175.946 | Nonparametric Statistics |  |
| Adjusted Level of Significance | 0.047576 | 95\% CLT UCL | 108.4498 |
| Adjusted Chi Square Value | 1174.797 | 95\% Jackknife UCL | 108.5226 |
|  |  | 95\% Standard Bootstrap UCL | 108.5582 |
| Anderson-Darling Test Statistic | 1.664037 | 95\% Bootstrap-t UCL | 111.1056 |
| Anderson-Darling 5\% Critical Value | 0.753691 | 95\% Hall's Bootstrap UCL | 111.334 |
| Kolmogorov-Smirnov Test Statistic | 0.098348 | 95\% Percentile Bootstrap UCL | 108.7616 |
| Kolmogorov-Smirnov 5\% Critical Value | 0.090015 | 95\% BCA Bootstrap UCL | 110.3384 |
| Data not Gamma Distributed at 5\% Significance Level |  | 95\% Chebyshev(Mean, Sd) UCL | 121.0305 |
|  |  | 97.5\% Chebyshev(Mean, Sd) UCL | 129.7734 |
| Assuming Gamma Distribution |  | 99\% Chebyshev(Mean, Sd) UCL | 146.947 |
| 95\% Approximate Gamma UCL | 107.7984 |  |  |
| 95\% Adjusted Gamma UCL | 107.9038 |  |  |
|  | Potential UCL to Use |  |  |
|  |  | Use 95\% Student's-t UCL | 108.5226 |
|  |  | or 95\% Modified-t UCL | 108.7182 |
|  |  | or 95\% H-UCL | 107.559 |

## APPENDIX D <br> DERIVATION OF THE PARTICULATE EMISSIONS FACTORS (PEFS)

Appendix D. 1
Particulate Emissions Factor for Residents and Outdoor Workers 4835 Glenbrook Road
Spring Valley, Washington, D.C.


Appendix D. 2
Dust Dispersion Factor
4835 Glenbrook Road Spring Valley, Washington, D.C.

$$
Q / C=A \times \exp \left[\frac{\left(\ln A_{\text {site }}-B\right)^{2}}{C}\right]
$$

| Parameter | Value | Units | Reference |  |
| :---: | :---: | :---: | :---: | :---: |
| Q/C | Dispersion factor | 87.37 | $\mathrm{~g} / \mathrm{m}^{2}$-s per $\mathrm{kg} / \mathrm{m}^{3}$ | site-specific |
| $\mathrm{A}_{\text {site }}$ | Area of the site | 0.5 | acres | Conservative assumption |
| A | Constant | 14.0111 | - | USEPA (2002) for Philadelphia, PA |

## APPENDIX E

## RAGS PART D TABLES



OCCURRENCE, DISTRIBUTION, AND SELECTION OF CHEMICALS OF POTENTIAL CONCERN 4835 Glenbrook Road, Spring Valley Formerly Used Defense Site (SVFUDS) Washington, DC

| Scenario Timefi Medium: <br> Exposure Medi |  | Current and future Soils <br> Mixed soils ( $0-10 \mathrm{ft} \mathrm{bgs}$ ) |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Exposure <br> Point | $\begin{gathered} \text { CAS } \\ \text { Number } \\ \hline \end{gathered}$ |  | Chemical | Minimum Concentration | Maximum Concentration | Units | Location <br> of Maximum <br> Concentration | Detection <br> Frequency | Range of <br> Detection <br> Limits <br> (1) | Concentration <br> Used for <br> Screening <br> (2) | Background Value (3) | $\begin{gathered} \text { Screening } \\ \text { Toxicity Value } \\ \text { (4) } \\ \hline \hline \end{gathered}$ | Potential ARAR/TBC Value | Potential ARARTBC Source | $\begin{gathered} \text { COPC } \\ \text { Flag } \\ \text { (YIN) } \\ \hline \hline \end{gathered}$ | Rationale for Selection or Deletion |
| Soil | 129-00-0 | Pyrene |  | 0.048 | 0.24 | mg/kg | SW-4835GB-04 (assoc w/ TP-40) | $2 / 6$ | ND - 0.4 | 170 | 0.63 | 170 N | NA | NA | No | BSL |
| Soil | 7782-49-2 | Selenium |  | 0.59 | 0.83 | mg/kg | SW-4835GB-04 (assoc w/ TP-40) | $2 / 3$ | 5.7 | 39 | 1.20 | 39 N | NA | NA | No | BSL |
| Soil | 7440-22-4 | Silver |  | 0.12 | 0.12 | mg/kg | SW-4835GB-01 (assoc w/TP-17) | $1 / 3$ | 0.91-0.91 | 39 | 0.87 | 39 N | NA | NA | No | BSL |
| Soil | 7440-24-6 | Strontium |  | 14.5 | 26.1 | $\mathrm{mg} / \mathrm{kg}$ | SW-4835GB-01 (assoc w/TP-17) | 3/3 | NA | 4,700 | 53.0 | 4,700 N | NA | NA | No | BSL |
| Soil | 13494-80-9 | Tellurium |  | 2.2 | 6.6 | mg/kg | SW-4835GB-01 (assoc w/TP-17) | 3/3 | NA | 5 | 5.0 | NA | NA | NA | No | NSL |
| Soil | 7440-28-0 | Thallium |  | 0.55 | 8.7 | mg/kg | SW-4835GB(-90,50)SWN(5)2.5 | $34 / 98$ | 0.6-23.4 | 2.2 | 2.2 | 0.5 N | NA | NA | Yes | ASL |
| Soil | 7440-31-5 | Tin |  | 14.6 | 14.6 | mg/kg | SW-4835GB-01 (assoc w/TP-17) | $1 / 3$ | 1.4-4.6 | 4,700 | 8.4 | $4,700 \mathrm{~N}$ | NA | NA | No | BSL |
| Soil | 7440-32-6 | Titanium |  | 325 | 867 | mg/kg | SW-4835GB-01 (assoc w/TP-17) | 3/3 | NA | 2,690 | 2,690 | NA | NA | NA | No | BSL |
| Soil | 7440-62-2 | Vanadium |  | 33.2 | 345 | $\mathrm{mg} / \mathrm{kg}$ | 4835GB(-170,-10)SW-S-3 | 99 / 99 | NA | 75.5 | 75.5 | 39.0 N | NA | NA | Yes | ASL |
| Soil | 7440-66-6 | Zinc |  | 31.7 | 180 | $\mathrm{mg} / \mathrm{kg}$ | SW-4835GB-01 | 99 / 99 | NA | 2,300 | 158 | 2,300 N | NA | NA | No | BSL |
| Soil | 7440-67-7 | Zirconium |  | 12.2 | 13.6 | $\mathrm{mg} / \mathrm{kg}$ | SW-4835GB-01 (assoc w/TP-17) | 2/3 | 16.9 | 48 | 48.3 | NA | NA | NA | No | BSL |

Footnotes:
(1) For the NDs
(2) Greater of the background and screening toxicity values (3) USACE (2008)
(4) USEPA (2009d) Regional Screening Levels, residentia
efinition
ASL - Above screening level
BSL - Below screening level
C - carcinogenic
N - Noncarcinogenic
ND - Not detetected
NSL - no screening leve

EXPOSURE POINT CONCENTRATION SUMMARY
REASONABLE MAXIMUM EXPOSURE (RME)

4835 Glenbrook Road, Spping Valley Formerly Used Defense Site (SVFUDS) Washington, DC

| Scenario Timeframe | Current and future |
| :--- | :--- |
| Medium: | Soils |
| Exposure Medium: | Soils |


| Exposure Point | Chemical of Potential Concern | Units | Arithmetic Mean <br> (1) | 95\% UCL <br> (Distribution) <br> (2) | Maximum <br> Detected Concentration | Exposure Point Concentration |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  |  |  |  |  |  | Value | Units | Statistic | Rationale (3) |
| 0-2 | Aluminum | $\mathrm{mg} / \mathrm{kg}$ | 21408 | 23116.14 n | 35300 | 23116.14 | $\mathrm{mg} / \mathrm{kg}$ | Student's-t UCL | >UCL and max |
| 0-10 | Aluminum | mg/kg | 24117.959 | 25532.93 k | 55900 | 25532.93 | $\mathrm{mg} / \mathrm{kg}$ | Kaplan-Meier (BCA) UCL | >UCL and max |
| 0-2 | Cobalt | mg/kg | 42 | -- | 42 | 42.00 | $\mathrm{mg} / \mathrm{kg}$ | Maximum detected | Maximum detected |
| 0-10 | Cobalt | mg/kg | 28 | -- | 42 | 42.00 | $\mathrm{mg} / \mathrm{kg}$ | Maximum detected | Maximum detected |
| 0-2 | Copper | mg/kg | 70.35111 | 79.24 g | 314 | 79.24 | $\mathrm{mg} / \mathrm{kg}$ | Approximate Gamma UCL | >UCL and max |
| 0-10 | Copper | $\mathrm{mg} / \mathrm{kg}$ | 78.729293 | 107.56 n | 444 | 107.56 | $\mathrm{mg} / \mathrm{kg}$ | Chebyshev UCL | >UCL and max |
| 0-2 | Manganese | $\mathrm{mg} / \mathrm{kg}$ | 542.70455 | 603.73 k | 1200 | 603.73 | $\mathrm{mg} / \mathrm{kg}$ | Kaplan-Meier (t) UCL | >UCL and max |
| 0-10 | Manganese | $\mathrm{mg} / \mathrm{kg}$ | 670.43878 | 772.96 k | 4110 | 772.96 | $\mathrm{mg} / \mathrm{kg}$ | Kaplan-Meier (BCA) UCL | >UCL and max |
| 0-2 | Mercury | mg/kg | 0.1312258 | 0.15 k | 0.83 | 0.15 | $\mathrm{mg} / \mathrm{kg}$ | Kaplan-Meier (BCA) UCL | >UCL and max |
| 0-10 | Mercury | mg/kg | 0.1226176 | 0.12 k | 0.83 | 0.12 | $\mathrm{mg} / \mathrm{kg}$ | Kaplan-Meier (Percentile Bootstrap) UCL | >UCL and max |
| 0-2 | Nickel | $\mathrm{mg} / \mathrm{kg}$ | 63.722222 | 73.73 g | 345 | 73.73 | $\mathrm{mg} / \mathrm{kg}$ | Approximate Gamma UCL | >UCL and max |
| 0-10 | Nickel | $\mathrm{mg} / \mathrm{kg}$ | 66.047475 | 71.95 g | 345 | 71.95 | $\mathrm{mg} / \mathrm{kg}$ | Approximate Gamma UCL | >UCL and max |
| 0-2 | Tellurium | mg/kg | 2.5 | -- | 2.5 | 2.50 | $\mathrm{mg} / \mathrm{kg}$ | Maximum detected | Maximum detected |
| 0-10 | Tellurium | $\mathrm{mg} / \mathrm{kg}$ | 3.767 | -- | 6.6 | 6.60 | $\mathrm{mg} / \mathrm{kg}$ | Maximum detected | Maximum detected |
| 0-2 | Thallium | $\mathrm{mg} / \mathrm{kg}$ | 1.0333333 | 1.09 k | 2.6 | 1.09 | $\mathrm{mg} / \mathrm{kg}$ | Kaplan-Meier (t) UCL | >UCL and max |
| 0-10 | Thallium | mg/kg | 1.4285294 | 1.35 k | 8.7 | 1.35 | $\mathrm{mg} / \mathrm{kg}$ | Kaplan-Meier (t) UCL | >UCL and max |
| 0-2 | Vanadium | mg/kg | 87.686667 | 94.26 n | 142 | 94.26 | $\mathrm{mg} / \mathrm{kg}$ | Student's-t UCL | >UCL and max |
| 0-10 | Vanadium | mg/kg | 100.82525 | 108.52 n | 345 | 108.52 | $\mathrm{mg} / \mathrm{kg}$ | Student's-t UCL | >UCL and max |

Footnotes:
(1) Arithmetic mean for detected concentrations only
(2) k-Kaplan-Meier distribution
n - data is not parametrically distributed
g- Gamma distribution
(3) The minimum UCL and maximum detected concentration was selected as the exposure point concentration

CENTRAL TENDENCY (CT)
4835 Glenbrook Road, Spping Valley Formerly Used Defense Site (SVFUDS) Washington, DC

| Scenario Timeframe: | Current and future |
| :--- | :--- |
| Medium: | Soils |
| Exposure Medium: | Soils |


| Exposure Point | Chemical of Potential Concern | Units | Arithmetic Mean <br> (1) | 95\% UCL <br> (Distribution) <br> (2) | Maximum <br> Detected Concentration | Exposure Point Concentration |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  |  |  |  |  |  | Value | Units | Statistic | Rationale <br> (3) |
| 0-2 | Aluminum | mg/kg | 21408 | \#\#\#\#\#\#\# n | 35300 | 21,000.00 | $\mathrm{mg} / \mathrm{kg}$ | Median | Data is not parametrically distributed |
| 0-10 | Aluminum | $\mathrm{mg} / \mathrm{kg}$ | 24117.959 | \#\#\#\#\#\#\# k | 55900 | 24,020.22 | $\mathrm{mg} / \mathrm{kg}$ | Kaplan-Meier mean | Kaplan-Meier distribution |
| 0-2 | Cobalt | $\mathrm{mg} / \mathrm{kg}$ | 42 | -- | 42 | 42 | $\mathrm{mg} / \mathrm{kg}$ | Arithmetic mean | Sample size too small; normality assumed |
| 0-10 | Cobalt | mg/kg | 28 | -- | 42 | 28 | $\mathrm{mg} / \mathrm{kg}$ | Arithmetic mean | Sample size too small; normality assumed |
| 0-2 | Copper | $\mathrm{mg} / \mathrm{kg}$ | 70.35111 | 79.24 g | 314 | 70.35 | $\mathrm{mg} / \mathrm{kg}$ | k star * theta star | Gamma distribution |
| 0-10 | Copper | mg/kg | 78.729293 | 107.56 n | 444 | 62.50 | $\mathrm{mg} / \mathrm{kg}$ | Median | Data is not parametrically distributed |
| 0-2 | Manganese | mg/kg | 542.70455 | 603.73 k | 1200 | 542.70 | $\mathrm{mg} / \mathrm{kg}$ | Kaplan-Meier mean | Kaplan-Meier distribution |
| 0-10 | Manganese | $\mathrm{mg} / \mathrm{kg}$ | 670.43878 | 772.96 k | 4110 | 669.40 | $\mathrm{mg} / \mathrm{kg}$ | Kaplan-Meier mean | Kaplan-Meier distribution |
| 0-2 | Mercury | mg/kg | 0.1312258 | 0.15 k | 0.83 | 0.11 | $\mathrm{mg} / \mathrm{kg}$ | Kaplan-Meier mean | Kaplan-Meier distribution |
| 0-10 | Mercury | mg/kg | 0.1226176 | 0.12 k | 0.83 | 0.10 | $\mathrm{mg} / \mathrm{kg}$ | Kaplan-Meier mean | Kaplan-Meier distribution |
| 0-2 | Nickel | $\mathrm{mg} / \mathrm{kg}$ | 63.722222 | 73.73 g | 345 | 63.72 | $\mathrm{mg} / \mathrm{kg}$ | k star * theta star | Gamma distribution |
| 0-10 | Nickel | mg/kg | 66.047475 | 71.95 g | 345 | 66.05 | $\mathrm{mg} / \mathrm{kg}$ | k star * theta star | Gamma distribution |
| 0-2 | Tellurium | mg/kg | 2.5 | -- | 2.5 | 2.5 | $\mathrm{mg} / \mathrm{kg}$ | Arithmetic mean | Sample size too small; normality assumed |
| 0-10 | Tellurium | $\mathrm{mg} / \mathrm{kg}$ | 3.767 | -- | 6.6 | 3.767 | $\mathrm{mg} / \mathrm{kg}$ | Arithmetic mean | Sample size too small; normality assumed |
| 0-2 | Thallium | $\mathrm{mg} / \mathrm{kg}$ | 1.0333333 | 1.09 k | 2.6 | 0.97 | $\mathrm{mg} / \mathrm{kg}$ | Kaplan-Meier mean | Kaplan-Meier distribution |
| 0-10 | Thallium | mg/kg | 1.4285294 | 1.35 k | 8.7 | 1.17 | $\mathrm{mg} / \mathrm{kg}$ | Kaplan-Meier mean | Kaplan-Meier distribution |
| 0-2 | Vanadium | mg/kg | 87.686667 | 94.26 n | 142 | 83.80 | $\mathrm{mg} / \mathrm{kg}$ | Median | Data is not parametrically distributed |
| 0-10 | Vanadium | $\mathrm{mg} / \mathrm{kg}$ | 100.82525 | 108.52 n | 345 | 93.70 | $\mathrm{mg} / \mathrm{kg}$ | Median | Data is not parametrically distributed |

Footnotes:
(1) Arithmetic mean for detected concentrations only.
(2) k- Kaplan-Meier distribution
n - data is not parametrically distributed
g - Gamma distribution
(3) Value presented as the Central Tendency is determined by the distribution

## APPENDIX F RISK CHARACTERIZATION TABLES

TABLE F. 1
RME HYPOTHETICAL ADULT RESIDENT
CARCINOGENIC AND NONCARCINOGENIC RISK ESTIMATES -- INGESTION OF SURFACE SOIL 4835 GLENBROOK ROAD - SPRING VALLEY

WASHINGTON, DC

| Exposure Assumptions |  |  |  | Risk and Hazard Equations |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Receptor |  | RME Hypothetical Resident |  | Carcinogenic: |  |  |  |  |  |
| COPC Concentration in Soil/Sediment ( $\mathrm{C}_{\text {soil/sed }}$ ) <br> chemical-specific $\quad \mathrm{mg} / \mathrm{Kg}$ |  |  |  | Risk $=\underline{\left(\mathrm{C}_{\text {soil/sed }}\right)\left(\mathrm{IR}_{\text {soil/sed }}\right)(\mathrm{EF})(\mathrm{ED})(\mathrm{FI})(\mathrm{CF})\left(\mathrm{SF}_{\mathrm{o}}\right)}$ |  |  |  |  |  |
| Soil Ingestion Rate ( $\mathrm{IR}_{\text {soil/sed }}$ ) |  |  | mg/day | (BW)(AT ${ }_{\text {c }}$ )(365days/year) |  |  |  |  |  |
| Exposure Frequency (EF) |  |  | days/yr | Non-Carcinogenic: |  |  |  |  |  |
| Exposure Durati |  | 30 | yrs |  |  |  |  |  |  |
| Fraction Contaminated Soil/Sediment Ingested <br> (FI) <br> Conversion Factor (CF) |  |  | unitless | HQ = | $\left(\mathrm{C}_{\text {soil/sed }}\right)\left(\mathrm{IR}_{\text {soil/sed }}\right)(\mathrm{EF})(\mathrm{ED})(\mathrm{FI})(\mathrm{CF})$ |  |  |  |  |
| Conversion Fact Averaging Time | $\text { nogens }\left(\mathrm{AT}_{\mathrm{c}}\right)$ | 0.00 | $\mathrm{Kg} / \mathrm{mg}$ <br> yrs |  | $\left(\mathrm{RfD}_{0}\right)(\mathrm{BW})\left(\mathrm{AT}_{\mathrm{nc}}\right)(365 d a y s /$ year $)$ |  |  |  |  |
| Averaging Time | arcinogens (AT |  | yrs |  |  |  |  |  |  |
| Oral Slope Facto |  | chemical | $(\mathrm{mg} / \mathrm{Kg}-\mathrm{day})^{-1}$ |  |  |  |  |  |  |
| Body Weight (B |  | 70 | Kg |  |  |  |  |  |  |
| Oral Reference Dose on ( $\mathrm{RfD}_{0}$ ) |  | chemical-specific mg/Kg-day |  |  |  |  |  |  |  |
| COPC ${ }^{\text {a/ }}$ | CAS ${ }^{\text {b/ }}$ | $\begin{gathered} \text { EPC } \\ (\mathrm{mg} / \mathrm{kg})^{c /} \end{gathered}$ | $\begin{gathered} \mathrm{SF}_{\mathrm{o}} \\ (\mathrm{mg} / \mathrm{kg}-\mathrm{day})^{-1} \mathrm{~d} / \end{gathered}$ |  | $\begin{gathered} \text { RFD } \\ \text { (mg/kg-day) } \end{gathered}$ | Cancer Risk | \% Of <br> Total | Hazard Quotient | $\begin{aligned} & \text { \% Of } \\ & \text { Total } \end{aligned}$ |
| Inorganics |  |  |  |  |  |  |  |  |  |
| Aluminum | 7429-90-5 | 23116.14 | --e/ |  | 1 | -- | -- | 3.2E-02 | 12\% |
| Cobalt | 7440-48-4 | 42 | -- |  | 0.0003 | -- | -- | $1.9 \mathrm{E}-01$ | 70\% |
| Copper | 7440-50-8 | 79.24033 | -- |  | 0.04 | -- | -- | 2.7E-03 | <1\% |
| Manganese | 7439-96-5 | 603.7308 | -- |  | 0.14 | -- | -- | $5.9 \mathrm{E}-03$ | 2\% |
| Mercury | 7439-97-6 | 0.1459022 | -- |  | 0.0003 | -- | -- | 6.7E-04 | <1\% |
| Nickel | 7440-02-0 | 73.72868 | -- |  | 0.02 | -- | -- | 5.0E-03 | 2\% |
| Tellurium | 13494-80-9 | 2.5 | -- |  | -- | -- | -- | -- | -- |
| Thallium | 7440-28-0 | 1.089113 | -- |  | 0.00008 | -- | -- | $1.9 \mathrm{E}-02$ | 7\% |
| Vanadium | 7440-62-2 | 94.2606 | -- |  | 0.007 | -- | -- | $1.8 \mathrm{E}-02$ | 7\% |
|  |  |  |  |  | Pathway Sums | -- |  | 2.7E-01 |  |

[^13]TABLE F. 2
RME HYPOTHETICAL ADULT RESIDENT
CARCINOGENIC AND NONCOARCINOGENIC RISK ESTIMATES -- INHALATION OF VOLATILES/PARTICULATES FROM SURFACE SOIL
4835 GLENBROOK ROAD - SPRING VALLEY
WASHINGTON, DC


[^14]TABLE F. 3
RME HYPOTHETICAL ADULT RESIDENT
CARCINOGENIC AND NONCARCINOGENIC RISK ESTIMATES -- INGESTION OF MIXED SOIL 4835 GLENBROOK ROAD - SPRING VALLEY

WASHINGTON, DC


[^15]TABLE F. 4
RME HYPOTHETICAL ADULT RESIDENT
CARCINOGENIC AND NONCOARCINOGENIC RISK ESTIMATES -- INHALATION OF VOLATILES/PARTICULATES FROM MIXED SOIL
4835 GLENBROOK ROAD - SPRING VALLEY WASHINGTON, DC


[^16]TABLE F. 5
RME HYPOTHETICAL CHILD RESIDENT

## CARCINOGENIC AND NONCARCINOGENIC RISK ESTIMATES -- INGESTION OF SURFACE SOIL 4835 GLENBROOK ROAD - SPRING VALLEY

## WASHINGTON, DC

| Exposure Assumptions |  |  |  | Risk and Hazard Equations |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Receptor |  | RME Hypothetical Child Resident |  | Carcinogenic: |  |  |  |  |  |
| COPC Concentration in Soil/Sediment ( $\mathrm{C}_{\text {soil/sed }}$ ) <br> chemical-specific $\quad \mathrm{mg} / \mathrm{Kg}$ |  |  |  | Risk $=\underline{\left(\mathrm{C}_{\text {soil/sed }}\right)\left(\mathrm{IR}_{\text {soil/sed }}\right)(\mathrm{EF})(\mathrm{ED})(\mathrm{FI})(\mathrm{CF})\left(\mathrm{SF}_{\mathrm{o}}\right)}$ |  |  |  |  |  |
| Soil Ingestion Rate ( $\mathrm{IR}_{\text {soil/sed }}$ ) |  |  | mg/day | (BW)( $\mathrm{AT}_{\mathrm{c}}$ )(365days/year) |  |  |  |  |  |
| Exposure Frequency (EF) |  |  | days/yr | Non-Carcinogenic: |  |  |  |  |  |
| Exposure Durati |  | 6 | yrs |  |  |  |  |  |  |
| Fraction Contan (FI) | Soil/Sedimen | ted $\quad 1$ | unitless | $\mathbf{H Q}=\quad\left(\mathrm{C}_{\text {soil/sed }}\right)\left(\mathrm{IR}_{\text {soil/sed }}\right)(\mathrm{EF})(\mathrm{ED})(\mathrm{FI})(\mathrm{CF})$ |  |  |  |  |  |
| Conversion Fact | nogens ( $A T_{c}$ ) | 0.000 7 | Kg/mg yrs | HQ - (RfD ${ }_{0}$ )(BW) $\left(\mathrm{AT}_{\mathrm{nc}}\right)(365 d a y s /$ year $)$ |  |  |  |  |  |
| Averaging Time | arcinogens (AT | 6 | yrs |  |  |  |  |  |  |
| Oral Slope Facto |  | chemic | $(\mathrm{mg} / \mathrm{Kg}-\mathrm{day})^{-1}$ |  |  |  |  |  |  |
| Body Weight (B |  | 1 | Kg |  |  |  |  |  |  |
| Oral Reference Dose on ( $\mathrm{RfD}_{\mathrm{o}}$ ) |  | chemical-specific mg/Kg-day |  |  |  |  |  |  |  |
| COPC ${ }^{\text {a/ }}$ | CAS ${ }^{\text {b/ }}$ | $\begin{gathered} \text { EPC } \\ (\mathrm{mg} / \mathrm{kg})^{c /} \end{gathered}$ | $\begin{gathered} \mathrm{SF}_{\mathrm{o}} \\ (\mathrm{mg} / \mathrm{kg}-\mathrm{day})^{-1} \mathrm{~d} / \end{gathered}$ |  | $\begin{gathered} \text { RFD } \\ \text { (mg/kg-day) } \end{gathered}$ | Cancer Risk | $\begin{aligned} & \text { \% Of } \\ & \text { Total } \end{aligned}$ | Hazard Quotient | $\begin{aligned} & \text { \% Of } \\ & \text { Total } \end{aligned}$ |
| Inorganics |  |  |  |  |  |  |  |  |  |
| Aluminum | 7429-90-5 | 23116.14 | --e/ |  | 1 | -- | -- | $1.5 \mathrm{E}-01$ | 12\% |
| Cobalt | 7440-48-4 | 42 | -- |  | 0.0003 | -- | -- | 8.9E-01 | 70\% |
| Copper | 7440-50-8 | 79.24033 | -- |  | 0.04 | -- | -- | $1.3 \mathrm{E}-02$ | <1\% |
| Manganese | 7439-96-5 | 603.7308 | -- |  | 0.14 | -- | -- | $2.8 \mathrm{E}-02$ | 2\% |
| Mercury | 7439-97-6 | 0.1459022 | -- |  | 0.0003 | -- | -- | $3.1 \mathrm{E}-03$ | <1\% |
| Nickel | 7440-02-0 | 73.72868 | -- |  | 0.02 | -- | -- | 2.4E-02 | 2\% |
| Tellurium | 13494-80-9 | 2.5 | -- |  | -- | -- | -- | -- | <1\% |
| Thallium | 7440-28-0 | 1.089113 | -- |  | 0.00008 | -- | -- | $8.7 \mathrm{E}-02$ | 7\% |
| Vanadium | 7440-62-2 | 94.2606 | -- |  | 0.007 | -- | -- | 8.6E-02 | 7\% |
|  |  |  |  |  | Pathway Sums | -- |  | $1.3 \mathrm{E}+00$ |  |

[^17]TABLE F. 6
RME HYPOTHETICAL CHILD RESIDENT
CARCINOGENIC AND NONCOARCINOGENIC RISK ESTIMATES -- INHALATION OF VOLATILES/PARTICULATES FROM SURFACE SOIL
4835 GLENBROOK ROAD - SPRING VALLEY
WASHINGTON, DC


[^18]TABLE F. 7
RME HYPOTHETICAL CHILD RESIDENT CARCINOGENIC AND NONCARCINOGENIC RISK ESTIMATES -- INGESTION OF MIXED SOIL 4835 GLENBROOK ROAD - SPRING VALLEY

WASHINGTON, DC

| Exposure Assumptions |  |  |  | Risk and Hazard Equations |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Receptor |  | RME Hypothetical Child Resident |  | Carcinogenic: |  |  |  |  |  |
| COPC Concentration in Soil/Sediment ( $\mathrm{C}_{\text {soil/sed }}$ ) <br> chemical-specific $\quad \mathrm{mg} / \mathrm{Kg}$ |  |  |  | Risk $=\underline{\left(\mathrm{C}_{\text {soil/sed }}\right)\left(\mathrm{IR}_{\text {soil/sed }}\right)(\mathrm{EF})(\mathrm{ED})(\mathrm{FI})(\mathrm{CF})\left(\mathrm{SF}_{\mathrm{o}}\right)}$ |  |  |  |  |  |
| Soil Ingestion Rate ( $\mathrm{IR}_{\text {soil/sed }}$ ) |  |  | mg /day | (BW)( $\mathrm{AT}_{\mathrm{c}}$ )(365days/year) |  |  |  |  |  |
| Exposure Frequency (EF) |  |  | days/yr | Non-Carcinogenic: |  |  |  |  |  |
| Exposure Duration (ED) |  |  | yrs |  |  |  |  |  |  |
| Fraction Contaminated Soil/Sediment Ingested(FI) |  |  | unitless | $\mathbf{H Q}=\frac{\left(\mathrm{C}_{\text {soil/sed }}\right)\left(\mathrm{IR}_{\text {soil/sed }}\right)(E F)(E D)(\mathrm{FI})(\mathrm{CF})}{\text { ( }}$ |  |  |  |  |  |
| Conversion Factor (CF) 0.00 |  |  | $\mathrm{Kg} / \mathrm{mg}$ <br> yrs | ( $\left.\mathrm{RfD}_{\mathrm{o}}\right)(\mathrm{BW})\left(\mathrm{AT}_{\text {nc }}\right)(365$ days/year) |  |  |  |  |  |
| Averaging Time, Noncarcinogens ( $\mathrm{AT}_{\mathrm{nc}}$ ) |  |  | yrs |  |  |  |  |  |  |
| Oral Slope Factor ( $\mathrm{SF}_{0}$ ) |  |  | $\left(\mathrm{mg} / \mathrm{Kg}\right.$-day) ${ }^{-1}$ |  |  |  |  |  |  |
| Body Weight (BW) |  |  | Kg |  |  |  |  |  |  |
| Oral Reference Dose on ( $\mathrm{RfD}_{0}$ ) |  | chemical-specific mg/Kg-day |  |  |  |  |  |  |  |
| COPC ${ }^{\text {a/ }}$ | CAS ${ }^{\text {b/ }}$ | $\begin{gathered} \text { EPC } \\ (\mathrm{mg} / \mathrm{kg})^{c /} \end{gathered}$ | $\begin{gathered} \mathrm{SF}_{\mathrm{o}} \\ (\mathrm{mg} / \mathrm{kg}-\mathrm{day})^{-1} \mathrm{~d} / \end{gathered}$ |  | $\begin{gathered} \text { RFD } \\ \text { (mg/kg-day) } \end{gathered}$ | Cancer Risk | \% Of <br> Total | Hazard Quotient | $\% \text { Of }$ <br> Total |
| Inorganics |  |  |  |  |  |  |  |  |  |
| Aluminum | 7429-90-5 | 25532.93 | --e/ |  | 1 | -- | -- | $1.6 \mathrm{E}-01$ | 12\% |
| Cobalt | 7440-48-4 | 42 | -- |  | 0.0003 | -- | -- | $8.9 \mathrm{E}-01$ | 67\% |
| Copper | 7440-50-8 | 107.5559 | -- |  | 0.04 | -- | -- | $1.7 \mathrm{E}-02$ | 1\% |
| Manganese | 7439-96-5 | 772.9605 | -- |  | 0.14 | -- | -- | 3.5E-02 | 3\% |
| Mercury | 7439-97-6 | 0.1202667 | -- |  | 0.0003 | -- | -- | $2.6 \mathrm{E}-03$ | <1\% |
| Nickel | 7440-02-0 | 71.95466 | -- |  | 0.02 | -- | -- | 2.3E-02 | 2\% |
| Tellurium | 13494-80-9 | 6.6 | -- |  | -- | -- | -- | -- | <1\% |
| Thallium | 7440-28-0 | 1.346168 | -- |  | 0.00008 | -- | -- | $1.1 \mathrm{E}-01$ | 8\% |
| Vanadium | 7440-62-2 | 108.5226 | -- |  | 0.007 | -- | -- | $9.9 \mathrm{E}-02$ | 7\% |
|  |  |  |  |  | Pathway Sums | -- |  | $1.3 \mathrm{E}+00$ |  |

[^19]TABLE F. 8
RME HYPOTHETICAL CHILD RESIDENT
CARCINOGENIC AND NONCOARCINOGENIC RISK ESTIMATES -- INHALATION OF VOLATILES/PARTICULATES FROM MIXED SOIL
4835 GLENBROOK ROAD - SPRING VALLEY WASHINGTON, DC


[^20]TABLE F. 9
RME OUTDOOR WORKER

## CARCINOGENIC AND NONCARCINOGENIC RISK ESTIMATES -- INGESTION OF SURFACE SOIL 4835 GLENBROOK ROAD - SPRING VALLEY WASHINGTON, DC



[^21]TABLE F. 10
RME OUTDOOR WORKER
CARCINOGENIC AND NONCOARCINOGENIC RISK ESTIMATES -- INHALATION OF VOLATILES/PARTICULATES FROM SURFACE SOIL
4835 GLENBROOK ROAD - SPRING VALLEY
WASHINGTON, DC


[^22]TABLE F. 11
RME OUTDOOR WORKER
CARCINOGENIC AND NONCARCINOGENIC RISK ESTIMATES -- INGESTION OF MIXED SOIL 4835 GLENBROOK ROAD - SPRING VALLEY

WASHINGTON, DC

| Exposure Assumptions |  |  |  | Risk and Hazard Equations |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Receptor |  | RME Outdoor Worker |  | Carcinogenic: |  |  |  |  |  |
| COPC Concentration in Soil/Sediment$\left(\mathrm{C}_{\text {soil/sed }}\right)$ |  |  |  | Risk $=\frac{\left(\mathrm{C}_{\text {soil/sed }}\right)\left(\mathrm{IR}_{\text {soil/sed }}\right)(\mathrm{EF})(\mathrm{ED})(\mathrm{FI})(\mathrm{CF})\left(\mathrm{SF}_{\mathrm{o}}\right)}{}$ |  |  |  |  |  |
| Soil Ingestion Rate ( $\mathrm{IR}_{\text {soil/sed }}$ ) |  |  | mg/day | (BW)( $\mathrm{AT}_{\mathrm{c}}$ )(365days/year) |  |  |  |  |  |
| Exposure Frequency (EF) |  |  | days/yr | Non-Carcinogenic: |  |  |  |  |  |
| Exposure Durati |  | 2 | yrs |  |  |  |  |  |  |
| Fraction Contaminated Soil/Sediment Ingested <br> (FI) <br> Conversion Factor (CF) |  |  | unitless | $\mathbf{H Q}=\frac{\left(\mathrm{C}_{\text {soil/sed }}\right)\left(\mathrm{IR}_{\text {soil/sed }}\right)(\mathrm{EF})(\mathrm{ED})(\mathrm{FI})(\mathrm{CF})}{}$ |  |  |  |  |  |
| Conversion Fact Averaging Time | $\text { nogens }\left(\mathrm{AT}_{\mathrm{c}}\right)$ | 0.00 | Kg/mg yrs | $\mathbf{H Q ~ - ~}\left(\mathrm{RfD}_{\mathrm{o}}\right)(\mathrm{BW})\left(\mathrm{AT}_{\text {nc }}\right)(365 d a y s /$ year $)$ |  |  |  |  |  |
| Averaging Time | arcinogens (AT |  | yrs |  |  |  |  |  |  |
| Oral Slope Factor |  | chemical | (mg/Kg-day) ${ }^{-1}$ |  |  |  |  |  |  |
| Body Weight (B |  | 70 | Kg |  |  |  |  |  |  |
| Oral Reference Dose on ( $\mathrm{RfD}_{0}$ ) |  | chemical-specific mg/Kg-day |  |  |  |  |  |  |  |
| COPC ${ }^{\text {a/ }}$ | CAS ${ }^{\text {b/ }}$ | $\begin{gathered} \text { EPC } \\ (\mathrm{mg} / \mathrm{kg})^{c /} \end{gathered}$ | $\begin{gathered} \mathrm{SF}_{\mathrm{o}} \\ (\mathrm{mg} / \mathrm{kg}-\mathrm{day})^{-1} \end{gathered}$ |  | $\begin{gathered} \text { RFD } \\ \text { (mg/kg-day) } \end{gathered}$ | Cancer Risk | \% Of <br> Total | Hazard Quotient | $\% \text { Of }$ <br> Total |
| Inorganics |  |  |  |  |  |  |  |  |  |
| Aluminum | 7429-90-5 | 25532.93 | --e/ |  | 1 | -- | -- | $1.2 \mathrm{E}-01$ | 12\% |
| Cobalt | 7440-48-4 | 42 | -- |  | 0.0003 | -- | -- | $6.6 \mathrm{E}-01$ | 67\% |
| Copper | 7440-50-8 | 107.5559 | -- |  | 0.04 | -- | -- | $1.3 \mathrm{E}-02$ | 1\% |
| Manganese | 7439-96-5 | 772.9605 | -- |  | 0.14 | -- | -- | $2.6 \mathrm{E}-02$ | 3\% |
| Mercury | 7439-97-6 | 0.1202667 | -- |  | 0.0003 | -- | -- | $1.9 \mathrm{E}-03$ | <1\% |
| Nickel | 7440-02-0 | 71.95466 | -- |  | 0.02 | -- | -- | $1.7 \mathrm{E}-02$ | 2\% |
| Tellurium | 13494-80-9 | 6.6 | -- |  | -- | -- | -- | -- | -- |
| Thallium | 7440-28-0 | 1.346168 | -- |  | 0.00008 | -- | -- | 7.9E-02 | 8\% |
| Vanadium | 7440-62-2 | 108.5226 | -- |  | 0.007 | -- | -- | 7.3E-02 | 7\% |
|  |  |  |  |  | Pathway Sums | -- |  | $9.9 \mathrm{E}-01$ |  |

[^23]TABLE F. 12
RME OUTDOOR WORKER CARCINOGENIC AND NONCOARCINOGENIC RISK ESTIMATES -- INHALATION OF VOLATILES/PARTICULATES FROM MIXED SOIL
4835 GLENBROOK ROAD - SPRING VALLEY WASHINGTON, DC


[^24]TABLE F. 13
CT HYPOTHETICAL ADULT RESIDENT
CARCINOGENIC AND NONCARCINOGENIC RISK ESTIMATES -- INGESTION OF SURFACE SOIL 4835 GLENBROOK ROAD - SPRING VALLEY

WASHINGTON, DC


[^25]TABLE F. 14
CT HYPOTHETICAL ADULT RESIDENT CARCINOGENIC AND NONCOARCINOGENIC RISK ESTIMATES -- INHALATION OF VOLATILES/PARTICULATES FROM SURFACE SOIL
4835 GLENBROOK ROAD - SPRING VALLEY
WASHINGTON, DC

| Exposure Assumptions |  |  |  | Risk and Hazard Equations |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Receptor | CT Hypothetical Adult Resident |  |  | Carcinogenic: |  |  |  |  |  |
| COPC Ambient Air Concentration due to volatile or particulate emissions from soil (Cair) | chemical-specific |  | $\mu \mathrm{g} / \mathrm{m}^{3}$ | Risk $=\frac{\left(\mathrm{C}_{\mathrm{air}}\right)(\mathrm{EF})(\mathrm{ED})(\mathrm{ET})(\mathrm{URF})}{}$ |  |  |  |  |  |
| Exposure Frequency (EF) |  | 350 | days/yr |  |  | $\mathrm{T}_{\mathrm{c}}$ )(365day | ear) |  |  |
| Exposure Duration (ED) |  | 9 | yrs | Non-Carcinog | genic: |  |  |  |  |
| Fraction of EF in Contact with Soil (ET) Averaging Time, Carcinogens ( $\mathrm{AT}_{\mathrm{C}}$ ) |  | 0.0625 70 | unitless yrs | HO = |  | $\left.{ }_{i r}\right)(E F)(E D$ |  |  |  |
| Averaging Time, Noncarcinogens $\left(\mathrm{AT}_{\mathrm{N}}\right)$ |  | 9 | yrs | HQ - |  | $\mathrm{AT}_{n \mathrm{nc}}$ )(365 | /year) |  |  |
| Inhalation Unit Risk Factor(URF) |  | mical-specific | $\left(\mu \mathrm{g} / \mathrm{m}^{3}\right)^{-1}$ | where: |  |  |  |  |  |
| Inhalation Reference Concentration (RfC) |  | mical-specific | $\mu \mathrm{g} / \mathrm{m}^{3}$ | C ${ }_{\text {ar-VO }}$ |  |  | VOCs |  |  |
| Volatilization Factor (VF) |  | ical-specific | $\mathrm{m}^{3} / \mathrm{Kg}$ | air-VOC |  |  | OCs |  |  |
| Particulate emission factor (PEF) | $3.2 \mathrm{E}+09$ |  | $\mathrm{m}^{3} / \mathrm{kg}$ | $\mathrm{Cair}_{\text {air-Particulate }}=$ |  | for non-VOCs |  |  |  |
| COPC $^{\text {a }}$ CAS ${ }^{\text {b/ }}$ | $\begin{gathered} \text { EPC } \\ (\mu \mathrm{g} / \mathrm{kg})^{\mathrm{c} l} \end{gathered}$ | Volatilization Factor $\left(\mathrm{m}^{3} / \mathrm{kg}\right)^{\mathrm{d} /}$ | $\begin{gathered} \mathrm{C}_{\mathrm{air}} \\ \left(\mu \mathrm{~g} / \mathrm{m}^{3}\right)^{\mathrm{e}} \end{gathered}$ | $\begin{gathered} \text { URF } \\ \left(\mu \mathrm{g} / \mathrm{m}^{3}\right)_{1} \end{gathered}$ | $\begin{gathered} \text { RFC } \\ \left(\mu \mathrm{g} / \mathrm{m}^{3}\right) \end{gathered}$ | Cancer Risk | \% Of <br> Tota | Hazard Quotient | \% |
| Inorganics |  |  |  |  |  |  |  |  |  |
| Aluminum 7429-90-5 | $2.1 \mathrm{E}+07$ | --f/ | 6.5E-03 | -- | $5.0 \mathrm{E}+00$ | -- | -- | 7.8E-05 | 18\% |
| Cobalt 7440-48-4 | 42000 | -- | $1.3 \mathrm{E}-05$ | $9.0 \mathrm{E}-03$ | 6.0E-03 | 9.0E-10 | 96\% | $1.3 \mathrm{E}-04$ | 31\% |
| Copper 7440-50-8 | 70000 | -- | 2.2E-05 | -- | -- | -- | -- | -- | -- |
| Manganese 7439-96-5 | 543000 | -- | $1.7 \mathrm{E}-04$ | -- | $5.0 \mathrm{E}-02$ | -- | -- | 2.0E-04 | 48\% |
| Mercury 7439-97-6 | 110 | -- | 3.4E-08 | -- | $2.0 \mathrm{E}-01$ | -- | -- | $1.0 \mathrm{E}-08$ | <1\% |
| Nickel 7440-02-0 | 63720 | -- | $2.0 \mathrm{E}-05$ | 2.6E-04 | 9.0E-02 | 3.9E-11 | 4\% | $1.3 \mathrm{E}-05$ | 3\% |
| Tellurium 13494-80-9 | 2500 | -- | 7.7E-07 | -- | -- | -- | -- | -- | -- |
| Thallium 7440-28-0 | 970 | -- | $3.0 \mathrm{E}-07$ | -- | -- | -- | -- | -- | -- |
| Vanadium 7440-62-2 | 83800 | -- | $2.6 \mathrm{E}-05$ | -- | -- | -- | -- | -- | -- |
|  |  |  |  | Pathwa | y Sums | 9.4E-10 |  | 4.2E-04 |  |

[^26]TABLE F. 15
CT HYPOTHETICAL ADULT RESIDENT
CARCINOGENIC AND NONCARCINOGENIC RISK ESTIMATES -- INGESTION OF MIXED SOIL 4835 GLENBROOK ROAD - SPRING VALLEY

WASHINGTON, DC

| Exposure Assumptions |  |  |  | Risk and Hazard Equations |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Receptor CT Hypothetical Adult Resident |  |  |  | Carcinogenic: |  |  |  |  |  |
| COPC Concentration in Soil/Sediment$\left(\mathrm{C}_{\text {soil/sed }}\right)$ |  |  |  | Risk $=\frac{\left(\mathrm{C}_{\text {soil/sed }}\right)\left(\mathrm{IR}_{\text {soil/sed }}\right)(\mathrm{EF})(\mathrm{ED})(\mathrm{FI})(\mathrm{CF})\left(\mathrm{SF}_{0}\right)}{\text { ) }}$ |  |  |  |  |  |
| Soil Ingestion Rate ( $\mathrm{IR}_{\text {soil/sed }}$ ) |  |  | mg /day | $(\mathrm{BW})\left(\mathrm{AT}_{\mathrm{c}}\right)(365 \mathrm{days} / \mathrm{year})$ |  |  |  |  |  |
| Exposure Frequency (EF) |  |  | days/yr |  |  |  |  |  |  |
| Exposure Duration (ED) <br> Fraction Contaminated Soil/Sediment Ingested <br> (FI) |  |  | yrs | Non-Carcinogenic: |  |  |  |  |  |
|  |  |  | unitless | $\mathbf{H Q}=\quad\left(\mathrm{C}_{\text {soil/sed }}\right)\left(\mathrm{IR}_{\text {soil/sed }}\right)(\mathrm{EF})(\mathrm{ED})(\mathrm{FI})(\mathrm{CF})$ |  |  |  |  |  |
| Conversion Factor (CF) |  |  | $\begin{gathered} \mathrm{Kg} / \mathrm{mg} \\ \mathrm{yrs} \end{gathered}$ | HQ $\quad\left(\mathrm{RfD}_{\mathrm{o}}\right)(\mathrm{BW})\left(\mathrm{AT}_{\text {nc }}\right)(365 d a y s /$ year $)$ |  |  |  |  |  |
| Averaging Time, Noncarcinogens ( $\mathrm{AT}_{\mathrm{nc}}$ ) |  |  | yrs |  |  |  |  |  |  |
| Oral Slope Factor ( $\mathrm{SF}_{\mathrm{o}}$ ) |  |  | $(\mathrm{mg} / \mathrm{Kg}-\text { day })^{-1}$ |  |  |  |  |  |  |
| Body Weight (BW) |  |  | Kg |  |  |  |  |  |  |
| Oral Reference Dose on ( $\mathrm{RfD}_{0}$ ) |  | chemical-specific $\quad \mathrm{mg} / \mathrm{Kg}$-day |  |  |  |  |  |  |  |
| COPC ${ }^{\text {a/ }}$ | CAS ${ }^{\text {b/ }}$ | $\begin{gathered} \text { EPC } \\ (\mathrm{mg} / \mathrm{kg})^{c l} \end{gathered}$ | $\begin{gathered} \mathrm{SF}_{\mathrm{o}} \\ (\mathrm{mg} / \mathrm{kg}-\mathrm{day})^{-1 d} \end{gathered}$ |  | $\begin{gathered} \text { RFD }_{\mathrm{o}} \\ \text { (mg/kg-day) } \end{gathered}$ | Cancer Risk | \% Of <br> Total | Hazard Quotient | \% Of <br> Total |
| Inorganics |  |  |  |  |  |  |  |  |  |
| Aluminum | 7429-90-5 | 24020.22 | -_e/ |  | 1 | -- | -- | $1.6 \mathrm{E}-02$ | 15\% |
| Cobalt | 7440-48-4 | 28 | -- |  | 0.0003 | -- | -- | 6.4E-02 | 60\% |
| Copper | 7440-50-8 | 63 | -- |  | 0.04 | -- | -- | $1.1 \mathrm{E}-03$ | 1\% |
| Manganese | 7439-96-5 | 669 | -- |  | 0.14 | -- | -- | 3.3E-03 | 3\% |
| Mercury | 7439-97-6 | 0.1 | -- |  | 0.0003 | -- | -- | 2.3E-04 | <1\% |
| Nickel | 7440-02-0 | 66.05 | -- |  | 0.02 | -- | -- | 2.3E-03 | 2\% |
| Tellurium | 13494-80-9 | 3.77 | -- |  | -- | -- | -- | -- | -- |
| Thallium | 7440-28-0 | 1.17 | -- |  | 0.00008 | -- | -- | $1.0 \mathrm{E}-02$ | 9\% |
| Vanadium | 7440-62-2 | 93.7 | -- |  | 0.007 | -- | -- | $9.2 \mathrm{E}-03$ | 9\% |
|  |  |  |  |  | Pathway Sums | -- |  | $1.1 \mathrm{E}-01$ |  |

[^27]TABLE F. 16
CT HYPOTHETICAL ADULT RESIDENT CARCINOGENIC AND NONCOARCINOGENIC RISK ESTIMATES -- INHALATION OF VOLATILES/PARTICULATES FROM MIXED SOIL
4835 GLENBROOK ROAD - SPRING VALLEY WASHINGTON, DC


[^28]TABLE F. 17
CT HYPOTHETICAL CHILD RESIDENT

## CARCINOGENIC AND NONCARCINOGENIC RISK ESTIMATES -- INGESTION OF SURFACE SOIL

 4835 GLENBROOK ROAD - SPRING VALLEYWASHINGTON, DC

| Exposure Assumptions |  |  |  | Risk and Hazard Equations |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Receptor |  | CT Hypothetical Child Resident |  | Carcinogenic: |  |  |  |  |  |
| COPC Concentration in Soil/Sediment ( $\mathrm{C}_{\text {soil/sed }}$ ) <br> chemical-specific $\quad \mathrm{mg} / \mathrm{Kg}$ |  |  |  | Risk $=\frac{\left(\mathrm{C}_{\text {soil/sed }}\right)\left(\mathrm{IR}_{\text {soil/sed }}\right)(\mathrm{EF})(\mathrm{ED})(\mathrm{FI})(\mathrm{CF})\left(\mathrm{SF}_{\mathrm{o}}\right)}{}$ |  |  |  |  |  |
| Soil Ingestion Rate ( $\mathrm{IR}_{\text {soil/sed }}$ ) |  |  | mg/day | (BW)( $\mathrm{AT}_{\mathrm{c}}$ )(365days/year) |  |  |  |  |  |
| Exposure Frequency (EF) |  |  | days/yr | Non-Carcinogenic: |  |  |  |  |  |
| Exposure Duration (ED) |  |  | yrs |  |  |  |  |  |  |
| Fraction Contaminated Soil/Sediment Ingested(FI) |  |  | unitless | $\mathbf{H Q}=\frac{\left(\mathrm{C}_{\text {soil/sed }}\right)\left(\mathrm{IR}_{\text {soil/sed }}\right)(E F)(E D)(\mathrm{FI})(\mathrm{CF})}{\text { ( }}$ |  |  |  |  |  |
| Conversion Fact Averaging Time | $\text { nogens }\left(\mathrm{AT}_{\mathrm{c}}\right)$ | 0.000 70 | Kg/mg <br> yrs | HQ = | $\left(\mathrm{RfD} \mathrm{o}_{0}\right)(\mathrm{BW})\left(\mathrm{AT}_{\mathrm{nc}}\right)(365 d a y s /$ year $)$ |  |  |  |  |
| Averaging Time, Noncarcinogens ( $\mathrm{AT}_{\mathrm{nc}}$ ) |  |  | yrs |  |  |  |  |  |  |
| Oral Slope Factor ( $\mathrm{SF}_{\mathrm{o}}$ ) |  |  | $(\mathrm{mg} / \mathrm{Kg}-\mathrm{day})^{-1}$ |  |  |  |  |  |  |
| Body Weight (BW) |  | 15 | Kg |  |  |  |  |  |  |
| Oral Reference Dose on ( $\mathrm{RfD}_{0}$ ) |  | chemical | mg/Kg-day |  |  |  |  |  |  |
| COPC ${ }^{\text {a/ }}$ | CAS ${ }^{\text {b/ }}$ | $\begin{gathered} \text { EPC } \\ (\mathrm{mg} / \mathrm{kg})^{\mathrm{cl}} \end{gathered}$ | $\begin{gathered} \mathrm{SF}_{\mathrm{o}} \\ (\mathrm{mg} / \mathrm{kg}-\mathrm{day})^{-1 \mathrm{~d} /} \end{gathered}$ |  | $\begin{gathered} \mathrm{RFD}_{\mathrm{o}} \\ \text { (mg/kg-day) } \end{gathered}$ | Cancer Risk | $\begin{aligned} & \text { \% Of } \\ & \text { Total } \end{aligned}$ | Hazard Quotient | \% Of <br> Total |
| Inorganics |  |  |  |  |  |  |  |  |  |
| Aluminum | 7429-90-5 | 21000 | --e/ |  | 1 | -- | -- | $1.3 \mathrm{E}-01$ | 11\% |
| Cobalt | 7440-48-4 | 42 | -- |  | 0.0003 | -- | -- | 8.9E-01 | 72\% |
| Copper | 7440-50-8 | 70 | -- |  | 0.04 | -- | -- | $1.1 \mathrm{E}-02$ | <1\% |
| Manganese | 7439-96-5 | 543 | -- |  | 0.14 | -- | -- | $2.5 \mathrm{E}-02$ | 2\% |
| Mercury | 7439-97-6 | 0.11 | -- |  | 0.0003 | -- | -- | 2.3E-03 | <1\% |
| Nickel | 7440-02-0 | 63.72 | -- |  | 0.02 | -- | -- | $2.0 \mathrm{E}-02$ | 2\% |
| Tellurium | 13494-80-9 | 2.5 | -- |  | -- | -- | -- | -- | <1\% |
| Thallium | 7440-28-0 | 0.97 | -- |  | 0.00008 | -- | -- | $7.8 \mathrm{E}-02$ | 6\% |
| Vanadium | 7440-62-2 | 83.8 | -- |  | 0.007 | -- | -- | 7.7E-02 | 6\% |
|  |  |  |  |  | Pathway Sums | -- |  | $1.2 \mathrm{E}+00$ |  |

[^29]TABLE F. 18
CT HYPOTHETICAL CHILD RESIDENT

## CARCINOGENIC AND NONCOARCINOGENIC RISK ESTIMATES -- INHALATION OF VOLATILES/PARTICULATES FROM SURFACE SOIL <br> 4835 GLENBROOK ROAD - SPRING VALLEY <br> WASHINGTON, DC

| Exposure Assumptions |  |  |  | Risk and Hazard Equations |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Receptor | CT Hypothetical Child Resident |  |  | Carcinogenic: |  |  |  |  |  |
| COPC Ambient Air Concentration due to volatile or particulate emissions from soil (Cair) | chemical-specific |  | $\mu \mathrm{g} / \mathrm{m}^{3}$ | Risk $=\frac{\left(\mathrm{C}_{\mathrm{air}}\right)(\mathrm{EFF})(\mathrm{ED})(\mathrm{ET})(\mathrm{URF})}{}$ |  |  |  |  |  |
| Exposure Frequency (EF) |  | 350 | days/yr |  |  | $\mathrm{T}_{\mathrm{c}}$ )(365da | ear) |  |  |
| Exposure Duration (ED) |  | 6 | yrs | Non-Carcinog | genic: |  |  |  |  |
| Fraction of EF in Contact with Soil (ET) |  | 0.074 | unitless |  |  | )(EF)(ED) |  |  |  |
| Averaging Time, Carcinogens ( $\mathrm{AT}_{\mathrm{C}}$ ) |  | 70 | yrs | HQ = |  | ir $)(E F)(E D$ |  |  |  |
| Averaging Time, Noncarcinogens $\left(\mathrm{AT}_{\mathrm{N}}\right)$ |  | 6 | yrs |  |  | $\left(T_{n c}\right)(365$ | /year) |  |  |
| Inhalation Unit Risk Factor(URF) |  | ical-specific | $\left(\mu \mathrm{g} / \mathrm{m}^{3}\right)^{-1}$ | where: |  |  |  |  |  |
| Inhalation Reference Concentration (RfC) |  | mical-specific | $\mu \mathrm{g} / \mathrm{m}^{3}$ | C |  |  | VOCs; |  |  |
| Volatilization Factor (VF) |  | cal-specific | $\mathrm{m}^{3} / \mathrm{Kg}$ | air-VOC |  |  | OCs, |  |  |
| Particulate emission factor (PEF) | $3.2 \mathrm{E}+09$ |  | $\mathrm{m}^{3} / \mathrm{kg}$ | $\mathrm{Cair}_{\text {air-Particulate }}=$ |  | for non-VOCs |  |  |  |
| COPC $^{\text {a }}$ CAS ${ }^{\text {b/ }}$ | $\begin{gathered} \text { EPC } \\ (\mu \mathrm{g} / \mathrm{kg})^{c /} \end{gathered}$ | Volatilization Factor $\left(\mathrm{m}^{3} / \mathrm{kg}\right)^{\mathrm{d} /}$ | $\begin{gathered} \mathrm{C}_{\mathrm{air}} \\ \left(\mu \mathrm{~g} / \mathrm{m}^{3}\right)^{\mathrm{e}} \end{gathered}$ | $\begin{aligned} & \text { URF } \\ & \left(\mu \mathrm{g} / \mathrm{m}^{3}\right)^{-} \end{aligned}$ | $\begin{gathered} \text { RFC } \\ \left(\mu \mathrm{g} / \mathrm{m}^{3}\right) \end{gathered}$ | Cancer Risk | \% | Hazard Quotient | \% |
| Inorganics |  |  |  |  |  |  |  |  |  |
| Aluminum 7429-90-5 | $2.1 \mathrm{E}+07$ | --f/ | 6.5E-03 | -- | $5.0 \mathrm{E}+00$ | -- | -- | 9.2E-05 | 18\% |
| Cobalt 7440-48-4 | 42000 | -- | $1.3 \mathrm{E}-05$ | $9.0 \mathrm{E}-03$ | 6.0E-03 | 7.1E-10 | 96\% | $1.5 \mathrm{E}-04$ | 31\% |
| Copper 7440-50-8 | 70000 | -- | $2.2 \mathrm{E}-05$ | -- | -- | -- | -- | -- | -- |
| Manganese 7439-96-5 | 543000 | -- | $1.7 \mathrm{E}-04$ | -- | 5.0E-02 | -- | -- | $2.4 \mathrm{E}-04$ | 48\% |
| Mercury 7439-97-6 | 110 | -- | 3.4E-08 | -- | 2.0E-01 | -- | -- | $1.2 \mathrm{E}-08$ | <1\% |
| Nickel 7440-02-0 | 63720 | -- | $2.0 \mathrm{E}-05$ | $2.6 \mathrm{E}-04$ | 9.0E-02 | 3.1E-11 | 4\% | $1.6 \mathrm{E}-05$ | 3\% |
| Tellurium 13494-80-9 | 2500 | -- | $7.7 \mathrm{E}-07$ | -- | -- | -- | -- | -- | -- |
| Thallium 7440-28-0 | 970 | -- | $3.0 \mathrm{E}-07$ | -- | -- | -- | -- | -- | -- |
| Vanadium 7440-62-2 | 83800 | -- | 2.6E-05 | -- | -- | -- | -- | -- | -- |
|  |  |  |  | Pathway | y Sums | 7.4E-10 |  | $5.0 \mathrm{E}-04$ |  |

[^30]TABLE F. 19
CT HYPOTHETICAL CHILD RESIDENT
CARCINOGENIC AND NONCARCINOGENIC RISK ESTIMATES -- INGESTION OF MIXED SOIL 4835 GLENBROOK ROAD - SPRING VALLEY

WASHINGTON, DC

| Exposure Assumptions |  |  |  | Risk and Hazard Equations |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Receptor |  | CT Hypothetical Child Resident |  | Carcinogenic: |  |  |  |  |  |
| COPC Concentration in Soil/Sediment ( $\mathrm{C}_{\text {soil/sed }}$ ) <br> chemical-specific $\quad \mathrm{mg} / \mathrm{Kg}$ |  |  |  | Risk $=\underline{\left(\mathrm{C}_{\text {soil/sed }}\right)\left(\mathrm{IR}_{\text {soil/sed }}\right)(\mathrm{EF})(\mathrm{ED})(\mathrm{FI})(\mathrm{CF})\left(\mathrm{SF}_{\mathrm{o}}\right)}$ |  |  |  |  |  |
| Soil Ingestion Rate ( $\mathrm{IR}_{\text {soil/sed }}$ ) |  |  | mg/day | (BW)( $\mathrm{AT}_{\mathrm{c}}$ )(365days/year) |  |  |  |  |  |
| Exposure Frequency (EF) |  |  | days/yr | Non-Carcinogenic: |  |  |  |  |  |
| Exposure Duration (ED) |  |  | yrs |  |  |  |  |  |  |
| Fraction Contaminated Soil/Sediment Ingested(FI) |  |  | unitless | $\mathbf{H Q}=\frac{\left(\mathrm{C}_{\text {soil/sed }}\right)\left(\mathrm{IR}_{\text {soil/sed }}\right)(\mathrm{EF})(\mathrm{ED})(\mathrm{FI})(\mathrm{CF})}{}$ |  |  |  |  |  |
| Conversion Fact Averaging Time | $\text { nogens }\left(\mathrm{AT}_{\mathrm{c}}\right)$ | 0.00 | $\mathrm{Kg} / \mathrm{mg}$ <br> yrs | HQ - ( $\left.\mathrm{RfD}_{\mathrm{o}}\right)(\mathrm{BW})\left(\mathrm{AT}_{\text {nc }}\right)(365 d a y s /$ year $)$ |  |  |  |  |  |
| Averaging Time | arcinogens (AT | 6 | yrs |  |  |  |  |  |  |
| Oral Slope Facto |  | chemical | $(\mathrm{mg} / \mathrm{Kg}-\mathrm{day})^{-1}$ |  |  |  |  |  |  |
| Body Weight (B |  |  | Kg |  |  |  |  |  |  |
| Oral Reference Dose on ( $\mathrm{RfD}_{0}$ ) |  | chemical-specific mg/Kg-day |  |  |  |  |  |  |  |
| COPC ${ }^{\text {a/ }}$ | CAS ${ }^{\text {b/ }}$ | $\begin{gathered} \text { EPC } \\ (\mathrm{mg} / \mathrm{kg})^{\mathrm{cl}} \end{gathered}$ | $\begin{gathered} \mathrm{SF}_{\mathrm{o}} \\ (\mathrm{mg} / \mathrm{kg}-\mathrm{day})^{-1 \mathrm{~d} /} \end{gathered}$ |  | $\begin{gathered} \text { RFD } \\ \text { (mg/kg-day) } \end{gathered}$ | Cancer Risk | $\begin{aligned} & \text { \% Of } \\ & \text { Total } \end{aligned}$ | Hazard Quotient | \% Of <br> Total |
| Inorganics |  |  |  |  |  |  |  |  |  |
| Aluminum | 7429-90-5 | 24020.22 | --e/ |  | 1 | -- | -- | $1.5 \mathrm{E}-01$ | 15\% |
| Cobalt | 7440-48-4 | 28 | -- |  | 0.0003 | -- | -- | $6.0 \mathrm{E}-01$ | 60\% |
| Copper | 7440-50-8 | 63 | -- |  | 0.04 | -- | -- | $1.0 \mathrm{E}-02$ | 1\% |
| Manganese | 7439-96-5 | 669 | -- |  | 0.14 | -- | -- | $3.1 \mathrm{E}-02$ | 3\% |
| Mercury | 7439-97-6 | 0.1 | -- |  | 0.0003 | -- | -- | $2.1 \mathrm{E}-03$ | <1\% |
| Nickel | 7440-02-0 | 66.05 | -- |  | 0.02 | -- | -- | 2.1E-02 | 2\% |
| Tellurium | 13494-80-9 | 3.77 | -- |  | -- | -- | -- | -- | <1\% |
| Thallium | 7440-28-0 | 1.17 | -- |  | 0.00008 | -- | -- | $9.3 \mathrm{E}-02$ | 9\% |
| Vanadium | 7440-62-2 | 93.7 | -- |  | 0.007 | -- | -- | 8.6E-02 | 9\% |
|  |  |  |  |  | Pathway Sums | -- |  | 9.9E-01 |  |

[^31]TABLE F. 20
CT HYPOTHETICAL CHILD RESIDENT

## CARCINOGENIC AND NONCOARCINOGENIC RISK ESTIMATES -- INHALATION OF VOLATILES/PARTICULATES FROM MIXED SOIL <br> 4835 GLENBROOK ROAD - SPRING VALLEY WASHINGTON, DC



[^32]TABLE F. 21

## CT OUTDOOR WORK

## CARCINOGENIC AND NONCARCINOGENIC RISK ESTIMATES -- INGESTION OF SURFACE SOIL 4835 GLENBROOK ROAD - SPRING VALLEY

 WASHINGTON, DC

[^33]TABLE F. 22
CT OUTDOOR WORK
CARCINOGENIC AND NONCOARCINOGENIC RISK ESTIMATES -- INHALATION OF VOLATILESIPARTICULATES FROM SURFACE SOIL
4835 GLENBROOK ROAD - SPRING VALLEY
WASHINGTON, DC
Exposure Assumptions
Receptor
COPC Ambient Air Concentration due to volatile or
particulate emissions from soil (Cair)
Exposure Frequency (EF)
Exposure Duration (ED)
Fraction of EF in Contact with Soil (ET)
Averaging Time, Carcinogens (AT $\mathrm{C}_{\mathrm{C}}$ )
Averaging Time, Noncarcinogens ( $\mathrm{AT}_{\mathrm{N}}$ )
Inhalation Unit Risk Factor(URF)
Inhalation Reference Concentration (RfC)
Volatilization Factor (VF)
Particulate emission factor (PEF)

| CT Outdoor Work chemical-specific | $\mu \mathrm{g} / \mathrm{m}^{3}$ | Risk and Hazard Equations Carcinogenic: |  |  |
| :---: | :---: | :---: | :---: | :---: |
|  |  | Risk $=\frac{\left(\mathrm{C}_{\text {air }}\right)(\mathrm{EF})(\mathrm{ED})(\mathrm{ET})(\mathrm{URF})}{}$ |  |  |
| 219 | days/yr | ( $\mathrm{AT}_{\mathrm{c}}$ )(365days/year) |  |  |
| 9 | yrs | Non-Carcinogenic: |  |  |
| 0.333 70 | unitless | $\mathbf{H Q}=\quad\left(\mathrm{C}_{\mathrm{air}}\right)(\mathrm{EF})(\mathrm{ED})(\mathrm{ET})$ |  |  |
| 9 | yrs | (RfC)( $\mathrm{AT}_{\mathrm{nc}}$ )(365days/year) |  |  |
| chemical-specific | $\left(\mu \mathrm{g} / \mathrm{m}^{3}\right)^{-1}$ | where: |  |  |
| chemical-specific | $\mu \mathrm{g} / \mathrm{m}^{3}$ | $\mathrm{Cair-VOC}^{=}$ | $\left(\mathrm{C}_{\text {soil }}\right)$ | for VOCs; and |
| chemical-specific | $\mathrm{m}^{3} / \mathrm{Kg}$ |  | (VF) |  |
| $3.2 \mathrm{E}+09$ | $\mathrm{m}^{3} / \mathrm{kg}$ | $\mathrm{C}_{\text {air-Particulate }}=$ | $\left(\mathrm{C}_{\text {soil }}\right)$ | for non-VOCs |
|  |  |  | (PEF) |  |


| COPC ${ }^{\text {a }}$ | CAS ${ }^{\text {b/ }}$ | $\begin{gathered} \text { EPC } \\ (\mu \mathrm{g} / \mathrm{kg})^{c l} \end{gathered}$ | Volatilization Factor $\left(\mathrm{m}^{3} / \mathrm{kg}\right)^{\mathrm{d} /}$ | $\begin{gathered} C_{\text {air }} \\ \left(\mu \mathrm{g} / \mathrm{m}^{3}\right)^{\mathrm{e}} \end{gathered}$ | $\begin{gathered} \text { URF } \\ \left(\mu \mathrm{g} / \mathrm{m}^{3}\right)^{-} \end{gathered}$ | RFC ( $\mu \mathrm{g} / \mathrm{m}^{3}$ ) | Cancer Risk | $\begin{aligned} & \text { \% } \\ & \text { Of } \\ & \text { Total } \end{aligned}$ | Hazard Quotient | $\begin{aligned} & \text { \% } \\ & \text { Of } \\ & \text { Total } \end{aligned}$ |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Inorganics |  |  |  |  |  |  |  |  |  |  |
| Aluminum | 7429-90-5 | $2.1 \mathrm{E}+07$ | --f/ | 6.5E-03 | -- | $5.0 \mathrm{E}+00$ | -- | -- | $2.6 \mathrm{E}-04$ | 18\% |
| Cobalt | 7440-48-4 | 42000 | -- | $1.3 \mathrm{E}-05$ | $9.0 \mathrm{E}-03$ | 6.0E-03 | 3.0E-09 | 96\% | $4.3 \mathrm{E}-04$ | 31\% |
| Copper | 7440-50-8 | 70000 | -- | $2.2 \mathrm{E}-05$ | -- | -- | -- | -- | -- | -- |
| Manganese | 7439-96-5 | 543000 | -- | $1.7 \mathrm{E}-04$ | -- | 5.0E-02 | -- | -- | $6.7 \mathrm{E}-04$ | 48\% |
| Mercury | 7439-97-6 | 110 | -- | 3.4E-08 | -- | 2.0E-01 | -- | -- | $3.4 \mathrm{E}-08$ | <1\% |
| Nickel | 7440-02-0 | 63720 | -- | $2.0 \mathrm{E}-05$ | 2.6E-04 | 9.0E-02 | 1.3E-10 | 4\% | $4.4 \mathrm{E}-05$ | 3\% |
| Tellurium | 13494-80-9 | 2500 | -- | 7.7E-07 | -- | -- | -- | -- | -- | -- |
| Thallium | 7440-28-0 | 970 | -- | 3.0E-07 | -- | -- | -- | -- | -- | -- |
| Vanadium | 7440-62-2 | 83800 | -- | $2.6 \mathrm{E}-05$ | -- | -- | -- | -- | -- | -- |
|  |  |  |  |  | Pathway Sums |  | $3.1 \mathrm{E}-09$ |  | $1.4 \mathrm{E}-03$ |  |

[^34]TABLE F. 23

## CT OUTDOOR WORK

## CARCINOGENIC AND NONCARCINOGENIC RISK ESTIMATES -- INGESTION OF MIXED SOIL

 4835 GLENBROOK ROAD - SPRING VALLEYWASHINGTON, DC

| Exposure Assumptions |  |  |  | Risk and Hazard Equations |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Receptor |  | CT Outdoor Work |  | Carcinogenic: |  |  |  |  |  |
| COPC Concentration in Soil/Sediment$\left(\mathrm{C}_{\text {soil/sed }}\right)$ |  |  |  | Risk $=\underline{\left(\mathrm{C}_{\text {soil/sed }}\right)\left(\mathrm{IR}_{\text {soil/sed }}\right)(\mathrm{EF})(\mathrm{ED})(\mathrm{FI})(\mathrm{CF})\left(\mathrm{SF}_{\mathrm{o}}\right)}$ |  |  |  |  |  |
| Soil Ingestion Rate ( $\mathrm{IR}_{\text {soil/sed }}$ ) |  |  | mg /day | $(\mathrm{BW})\left(\mathrm{AT}_{\mathrm{c}}\right)(365$ days/year) |  |  |  |  |  |
| Exposure Frequency (EF) |  |  | days/yr | Non-Carcinogenic: |  |  |  |  |  |
| Exposure Duration (ED) |  |  | yrs |  |  |  |  |  |  |
| Fraction Contaminated Soil/Sediment Ingested <br> (FI) <br> Conversion Factor (CF) |  |  | unitless | $\mathbf{H Q}=\frac{\left(\mathrm{C}_{\text {soil/sed }}\right)\left(\mathrm{IR}_{\text {soil/sed }}\right)(E F)(E D)(\mathrm{FI})(\mathrm{CF})}{\text { ( }}$ |  |  |  |  |  |
| Conversion Factor (CF) |  |  | Kg/mg yrs | $\left(\mathrm{RfD}{ }_{0}\right)(\mathrm{BW})\left(\mathrm{AT}_{\mathrm{nc}}\right)(365 d a y s /$ year $)$ |  |  |  |  |  |
| Averaging Time | arcinogens (AT | 9 | yrs |  |  |  |  |  |  |
| Oral Slope Facto |  | chemical | $(\mathrm{mg} / \mathrm{Kg}-\text { day })^{-1}$ |  |  |  |  |  |  |
| Body Weight (B |  | 7 | Kg |  |  |  |  |  |  |
| Oral Reference Dose on ( $\mathrm{RfD}_{\mathrm{o}}$ ) |  | chemical-specific mg/Kg-day |  |  |  |  |  |  |  |
| COPC ${ }^{\text {a/ }}$ | CAS ${ }^{\text {b/ }}$ | $\begin{gathered} \text { EPC } \\ (\mathrm{mg} / \mathrm{kg})^{c l} \end{gathered}$ | $\begin{gathered} \mathrm{SF}_{\mathrm{o}} \\ (\mathrm{mg} / \mathrm{kg}-\text { day })^{-1} \end{gathered}$ |  | $\begin{gathered} R F D_{o} \\ \text { (mg/kg-day) } \end{gathered}$ | Cancer Risk | $\begin{aligned} & \text { \% Of } \\ & \text { Total } \end{aligned}$ | Hazard Quotient | \% Of <br> Total |
| Inorganics |  |  |  |  |  |  |  |  |  |
| Aluminum | 7429-90-5 | 24020.22 | --e/ |  | 1 | -- | -- | 2.1E-02 | 15\% |
| Cobalt | 7440-48-4 | 28 | -- |  | 0.0003 | -- | -- | 8.0E-02 | 60\% |
| Copper | 7440-50-8 | 63 | -- |  | 0.04 | -- | -- | $1.4 \mathrm{E}-03$ | 1\% |
| Manganese | 7439-96-5 | 669 | -- |  | 0.14 | -- | -- | $4.1 \mathrm{E}-03$ | 3\% |
| Mercury | 7439-97-6 | 0.1 | -- |  | 0.0003 | -- | -- | $2.9 \mathrm{E}-04$ | <1\% |
| Nickel | 7440-02-0 | 66.05 | -- |  | 0.02 | -- | -- | $2.8 \mathrm{E}-03$ | 2\% |
| Tellurium | 13494-80-9 | 3.77 | -- |  | -- | -- | -- | -- | -- |
| Thallium | 7440-28-0 | 1.17 | -- |  | 0.00008 | -- | -- | 1.3E-02 | 9\% |
| Vanadium | 7440-62-2 | 93.7 | -- |  | 0.007 | -- | -- | $1.1 \mathrm{E}-02$ | 9\% |
|  |  |  |  |  | Pathway Sums | -- |  | 1.3E-01 |  |

[^35]TABLE F. 24
CT OUTDOOR WORK

## CARCINOGENIC AND NONCOARCINOGENIC RISK ESTIMATES -- INHALATION OF VOLATILES/PARTICULATES FROM MIXED SOIL <br> 4835 GLENBROOK ROAD - SPRING VALLEY WASHINGTON, DC

## Exposure Assumptions

## Receptor

COPC Ambient Air Concentration due to volatile or particulate emissions from soil (Cair)
Exposure Frequency (EF)
Exposure Duration (ED)
Fraction of EF in Contact with Soil (ET)
Averaging Time, Carcinogens ( $\mathrm{AT}_{\mathrm{C}}$ )
Averaging Time, Noncarcinogens $\left(\mathrm{AT}_{\mathrm{N}}\right)$
Inhalation Unit Risk Factor(URF)
Inhalation Reference Concentration (RfC)
Volatilization Factor (VF)
Particulate emission factor (PEF)


| COPC ${ }^{\text {a/ }}$ | CAS ${ }^{\text {b/ }}$ | $\begin{gathered} \text { EPC } \\ (\mu \mathrm{g} / \mathrm{kg})^{c /} \end{gathered}$ | Volatilization Factor $\left(\mathrm{m}^{3} / \mathrm{kg}\right)^{\mathrm{d} /}$ | $\begin{gathered} \mathrm{C}_{\mathrm{air}} \\ \left(\mu \mathrm{~g} / \mathrm{m}^{3}\right)^{\mathrm{e}} \end{gathered}$ | $\begin{gathered} \text { URF } \\ \left(\mu \mathrm{g} / \mathrm{m}^{3}\right)^{-} \end{gathered}$ | $\begin{gathered} \text { RFC } \\ \left(\mu \mathrm{g} / \mathrm{m}^{3}\right) \end{gathered}$ | Cancer Risk | \% Of <br> Total | Hazard Quotient | $\begin{aligned} & \text { \% } \\ & \text { Of } \\ & \text { Total } \end{aligned}$ |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Inorganics |  |  |  |  |  |  |  |  |  |  |
| Aluminum | 7429-90-5 | $2.402022 \mathrm{E}+07$ | -- ${ }^{\text {f/ }}$ | 7.4E-03 | -- | $5.0 \mathrm{E}+00$ | -- | -- | 3.0E-04 | 20\% |
| Cobalt | 7440-48-4 | 28000 | -- | 8.7E-06 | 9.0E-03 | 6.0E-03 | 2.0E-09 | 94\% | 2.9E-04 | 20\% |
| Copper | 7440-50-8 | 63000 | -- | $1.9 \mathrm{E}-05$ | -- | -- | -- | -- | -- | -- |
| Manganese | 7439-96-5 | 669000 | -- | $2.1 \mathrm{E}-04$ | -- | $5.0 \mathrm{E}-02$ | -- | -- | 8.3E-04 | 57\% |
| Mercury | 7439-97-6 | 100 | -- | 3.1E-08 | -- | $2.0 \mathrm{E}-01$ | -- | -- | $3.1 \mathrm{E}-08$ | <1\% |
| Nickel | 7440-02-0 | 66050 | -- | $2.0 \mathrm{E}-05$ | 2.6E-04 | $9.0 \mathrm{E}-02$ | 1.4E-10 | 6\% | $4.5 \mathrm{E}-05$ | 3\% |
| Tellurium | 13494-80-9 | 3770 | -- | $1.2 \mathrm{E}-06$ | -- | -- | -- | -- | -- | -- |
| Thallium | 7440-28-0 | 1170 | -- | 3.6E-07 | -- | -- | -- | -- | -- | -- |
| Vanadium | 7440-62-2 | 93700 | -- | $2.9 \mathrm{E}-05$ | -- | -- | -- | -- | -- | -- |
|  |  |  |  |  | Pathway Sums |  | 2.1E-09 |  | $1.5 \mathrm{E}-03$ |  |

[^36]
## APPENDIX G HOMEGROWN VEGETABLE INTAKE PARAMETERS

## Appendix G. 1

## Moisture Content and Dry Weight of Various Vegetables <br> 4835 Glenbrook Road <br> Spring Valley, Washington, D.C.

|  | Moisture Content | Dry Weight |
| :---: | :---: | :---: |
| Vegetable | (\%) | (\%) |
| Alfalfa sprouts | 91.14 | 8.86 |
| Artichokes - globe \& French | 84.38 | 15.62 |
| Artichokes - Jerusalem | 78.01 | 21.99 |
| Asparagus | 92.25 | 7.75 |
| Bamboo shoots | 91 | 9 |
| Beans - dry - blackeye peas (cowpeas) | 66.8 | 33.2 |
| Beans - dry - hyacinth (mature seeds) | 87.87 | 12.13 |
| Beans - dry - navy (pea) | 79.15 | 20.85 |
| Beans - dry - pinto | 81.3 | 18.7 |
| Beans - lima | 70.24 | 29.76 |
| Beans - snap - Italian - green - yellow | 90.27 | 9.73 |
| Beets | 87.32 | 12.68 |
| Beets - tops (greens) | 92.15 | 7.85 |
| Broccoli | 90.69 | 9.31 |
| Brussel sprouts | 86 | 14 |
| Cabbage - Chinese/celery, including bok choy | 95.32 | 4.68 |
| Cabbage - red | 91.55 | 8.45 |
| Cabbage - savoy | 91 | 9 |
| Carrots | 87.79 | 12.21 |
| Cassava (yucca blanca) | 68.51 | 31.49 |
| Cauliflower | 92.26 | 7.74 |
| Celeriac | 88 | 12 |
| Celery | 94.7 | 5.3 |
| Chili peppers | 87.74 | 12.26 |
| Chives | 92 | 8 |
| Cole slaw | 81.5 | 18.5 |
| Collards | 93.9 | 6.1 |
| Corn - sweet | 75.96 | 24.04 |
| Cress - garden - field | 89.4 | 10.6 |
| Cress - garden | 89.4 | 10.6 |
| Cucumbers | 96.05 | 3.95 |
| Dandelion - greens | 85.6 | 14.4 |
| Eggplant | 91.93 | 8.07 |
| Endive | 93.79 | 6.21 |
| Garlic | 58.58 | 41.42 |
| Kale | 84.46 | 15.54 |
| Kohlrabi | 91 | 9 |
| Lambsquarter | 84.3 | 15.7 |
| Leeks | 83 | 17 |
| Lentils - whole | 67.34 | 32.66 |
| Lettuce - iceberg | 95.89 | 4.11 |
| Lettuce - romaine | 94.91 | 5.09 |
| Mung beans (sprouts) | 90.4 | 9.6 |
| Mushrooms | 91.81 | 8.19 |

Appendix G. 1
Moisture Content and Dry Weight of Various Vegetables
4835 Glenbrook Road
Spring Valley, Washington, D.C.

| Vegetable | Moisture Content <br> (\%) | Dry Weight <br> (\%) |
| :--- | :---: | :---: |
| Mustard greens | 90.8 | 9.2 |
| Okra | 89.58 | 10.42 |
| Onions | 90.82 | 9.18 |
| Onions - dehydrated or dried | 3.93 | 96.07 |
| Parsley | 88.31 | 11.69 |
| Parsley roots | 88.31 | 11.69 |
| Parsnips | 79.53 | 20.47 |
| Peas (garden) - mature seeds - dry | 88.89 | 11.11 |
| Peppers - sweet - garden | 92.77 | 7.23 |
| Potatoes (white) - peeled | 78.96 | 21.04 |
| Potatoes (white) - whole | 83.29 | 16.71 |
| Pumpkin | 91.6 | 8.4 |
| Radishes - roots | 94.84 | 5.16 |
| Rhubarb | 93.61 | 6.39 |
| Rutabagas - unspecified | 89.66 | 10.34 |
| Salsify (oyster plant) | 77 | 23 |
| Shallots | 79.8 | 20.2 |
| Soybeans - sprouted seeds | 69.05 | 30.95 |
| Spinach | 91.58 | 8.42 |
| Squash - summer | 93.68 | 6.32 |
| Squash - winter | 88.71 | 11.29 |
| Sweetpotatoes (including yams) | 72.84 | 27.16 |
| Swiss chard | 92.66 | 7.34 |
| Tapioca - pearl | 10.99 | 89.01 |
| Taro - greens | 85.66 | 14.34 |
| Taro - root | 70.64 | 29.36 |
| Tomatoes - raw | 93.95 | 6.05 |
| Tomatoes - whole | 93.95 | 6.05 |
| Towelgourd | 93.85 | 6.15 |
| Turnips - roots | 91.87 | 8.13 |
| Turnips - tops | 91.07 | 8.93 |
| Water chestnuts | 73.46 | 26.54 |
| Yambean - tuber | 89.15 | 10.85 |
|  | $\mathbf{8 4 . 4 3}$ | 15.57 |
| Saut |  |  |

Source: USEPA (1997a), Table 9-27

## Appendix G. 2

## Percent Weight Loss from Preparation of Various Vegetables 4835 Glenbrook Road Spring Valley, Washington, D.C.

| Vegetable | Preparation Loss <br> (\%) |
| :--- | :---: |
| Asparagus | 23 |
| Beets | 28 |
| Broccoli | 14 |
| Cabbage | 11 |
| Carrots | 19 |
| Corn | 26 |
| Cucumbers | 18 |
| Lettuce | 22 |
| Lima beans | -12 |
| Okra | 12 |
| Onions | 5 |
| Peas, green | 2 |
| Peppers | 13 |
| Pumpkins | 19 |
| Snap beans | 18 |
| Tomatoes | 15 |
| Potatoes | 15 |
|  | -22 |
| Source: USEPA (1097a), Tablage: | 12.41 |

Source: USEPA (1997a), Table 13-7

## APPENDIX H EVALUATION OF ARSENIC

## Table H. 1

## Arsenic Summary Statistics

4835 Glenbrook Road
Spring Valley, Washington, D.C.

| Depth | Chemical | $\mathbf{N}$ | \#D | \%D | Units | MinD | MaxD | Distribution | UCL Calculated Using ${ }^{\mathbf{1}}$ | Central Tendency $^{2}$ |
| :--- | :--- | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| UCL | RME |  |  |  |  |  |  |  |  |  |
| 0-2 | Arsenic | 97 | 97 | $100 \%$ | $\mathrm{mg} / \mathrm{kg}$ | 1.1 | 19.9 | None | $95 \%$ Student's-t UCL |  |
| $0-10$ | Arsenic | 151 | 151 | $100 \%$ | $\mathrm{mg} / \mathrm{kg}$ | 0.69 | 19.9 | None | 9.4 |  |

Notes
1 UCLs were calculated by ProUCL using the indicated technique
2 Value presented as the Central Tendency is determined by the distribution as follows
Kaplan-Meier: the Kaplan-Meier mean
None: data is not parametrically distributed. The median is presented.
Lognormal: the backtransformed mean of the lognormal data
Gamma: k star * theta star
3 UCLs and Central Tendencies not calculated for datasets with less than ten samples [ $\mathrm{n}<10$ ] and/or less than 20 percent detections.

## Definitions:

N Total number of samples analyzed
NA Not applicable
ND Number of non-detects
\%D Percentage of detects
MinD Minimum detected value
MaxD Maximum detected value
UCL Upper confidence limit
RME Reasonable maximum exposure

Table H. 2

## Arsenic ProUCL Output 4835 Glenbrook Road Spring Valley, Washington, D.C.

| Full Precision | OFF |
| :--- | ---: |
| Confidence Coefficient | $95 \%$ |
| Number of Bootstrap Operations | 2000 |


| Arsenic 0-2 ft bgs |  |  |
| :---: | :---: | :---: |
| General Statistics |  |  |
| Number of Valid Samples | 97 Number of Unique Samples | 86 |
| Raw Statistics | Log-transformed Statistics |  |
| Minimum | 1.1 Minimum of Log Data | 0.0953 |
| Maximum | 19.9 Maximum of Log Data | 2.991 |
| Mean | 9.686 Mean of log Data | 2.083 |
| Median | 9.4 SD of log Data | 0.679 |
| SD | 5.14 |  |
| Coefficient of Variation | 0.531 |  |
| Skewness | 0.176 |  |
| Relevant UCL Statistics |  |  |
| Normal Distribution Test | Lognormal Distribution Test |  |
| Lilliefors Test Statistic | 0.072 Lilliefors Test Statistic | 0.117 |
| Lilliefors Critical Value | 0.09 Lilliefors Critical Value | 0.09 |
| Data appear Normal at 5\% Significance Level | Data not Lognormal at 5\% Significance Level |  |
| Assuming Normal Distribution | Assuming Lognormal Distribution |  |
| 95\% Student's-t UCL | 10.55 95\% H-UCL | 11.58 |
| 95\% UCLs (Adjusted for Skewness) | 95\% Chebyshev (MVUE) UCL | 13.41 |
| 95\% Adjusted-CLT UCL | 10.55 97.5\% Chebyshev (MVUE) UCL | 14.86 |
| 95\% Modified-t UCL | 10.55 99\% Chebyshev (MVUE) UCL | 17.69 |
| Gamma Distribution Test | Data Distribution |  |
| k star (bias corrected) | 2.737 Data appear Normal at 5\% Significance Level |  |
| Theta Star | 3.538 |  |
| nu star | 531.1 |  |
| Approximate Chi Square Value (.05) | 478.6 Nonparametric Statistics |  |
| Adjusted Level of Significance | 0.0475 95\% CLT UCL | 10.54 |
| Adjusted Chi Square Value | 477.9 95\% Jackknife UCL | 10.55 |
|  | 95\% Standard Bootstrap UCL | 10.54 |
| Anderson-Darling Test Statistic | 1.328 95\% Bootstrap-t UCL | 10.59 |
| Anderson-Darling 5\% Critical Value | 0.76 95\% Hall's Bootstrap UCL | 10.59 |
| Kolmogorov-Smirnov Test Statistic | 0.0962 95\% Percentile Bootstrap UCL | 10.54 |
| Kolmogorov-Smirnov 5\% Critical Value | 0.0916 95\% BCA Bootstrap UCL | 10.58 |
| Data not Gamma Distributed at 5\% Significance Level | 95\% Chebyshev(Mean, Sd) UCL | 11.96 |
|  | 97.5\% Chebyshev(Mean, Sd) UCL | 12.95 |
| Assuming Gamma Distribution | 99\% Chebyshev(Mean, Sd) UCL | 14.88 |
| 95\% Approximate Gamma UCL | 10.75 |  |
| 95\% Adjusted Gamma UCL | 10.76 |  |
|  | Potential UCL to Use Use 95\% Student's-t UCL | 10.55 |

Table H. 2

## Arsenic ProUCL Output 4835 Glenbrook Road Spring Valley, Washington, D.C.

| Full Precision | OFF |
| :--- | ---: |
| Confidence Coefficient | $95 \%$ |
| Number of Bootstrap Operations | 2000 |


| Arsenic 0-10 ft bgs |  |  |  |
| :---: | :---: | :---: | :---: |
| General Statistics |  |  |  |
| Number of Valid Samples | 151 | Number of Unique Samples | 117 |
| Raw Statistics |  | Log-transformed Statistics |  |
| Minimum | 0.69 | Minimum of Log Data | -0.371 |
| Maximum | 19.9 | Maximum of Log Data | 2.991 |
| Mean | 9.275 | Mean of log Data | 1.998 |
| Median |  | SD of log Data | 0.761 |
| SD | 5.341 |  |  |
| Coefficient of Variation | 0.576 |  |  |
| Skewness | 0.235 |  |  |
| Relevant UCL Statistics |  |  |  |
| Normal Distribution Test |  | Lognormal Distribution Test |  |
| Lilliefors Test Statistic | 0.0742 | Lilliefors Test Statistic | 0.132 |
| Lilliefors Critical Value | 0.0721 | Lilliefors Critical Value | 0.0721 |
| Data not Normal at 5\% Significance Level Data not Lognormal at 5\% Significance Level |  |  |  |
| Assuming Normal Distribution Assuming Lognormal Distribution |  |  |  |
| 95\% Student's-t UCL | 9.994 | 95\% H-UCL | 11.14 |
| 95\% UCLs (Adjusted for Skewness) |  | 95\% Chebyshev (MVUE) UCL | 12.81 |
| 95\% Adjusted-CLT UCL | 9.998 | 97.5\% Chebyshev (MVUE) UCL | 14.11 |
| 95\% Modified-t UCL | 9.995 | 99\% Chebyshev (MVUE) UCL | 16.65 |
| Gamma Distribution Test |  | Data Distribution |  |
| k star (bias corrected) | 2.288 | Data do not follow a Discernable Distribution (0.05) |  |
| Theta Star | 4.054 |  |  |
| nu star | 690.9 |  |  |
| Approximate Chi Square Value (.05) | 630.9 | Nonparametric Statistics |  |
| Adjusted Level of Significance | 0.0484 | 95\% CLT UCL | 9.989 |
| Adjusted Chi Square Value | 630.3 | 95\% Jackknife UCL | 9.994 |
|  |  | 95\% Standard Bootstrap UCL | 9.973 |
| Anderson-Darling Test Statistic | 1.971 | 95\% Bootstrap-t UCL | 9.99 |
| Anderson-Darling 5\% Critical Value | 0.764 | 95\% Hall's Bootstrap UCL | 9.958 |
| Kolmogorov-Smirnov Test Statistic | 0.0906 | 95\% Percentile Bootstrap UCL | 9.964 |
| Kolmogorov-Smirnov 5\% Critical Value | 0.0771 | 95\% BCA Bootstrap UCL | 9.952 |
| Data not Gamma Distributed at 5\% Significance Level |  | 95\% Chebyshev(Mean, Sd) UCL | 11.17 |
|  |  | 97.5\% Chebyshev(Mean, Sd) UCL | 11.99 |
| Assuming Gamma Distribution |  | 99\% Chebyshev(Mean, Sd) UCL | 13.6 |
| 95\% Approximate Gamma UCL | 10.16 |  |  |
| 95\% Adjusted Gamma UCL | 10.17 |  |  |
|  |  |  | 11.17 |

Table H-3
Homegrown Vegetables Bioaccumulation Factors
4835 Glenbrook Road
Spring Valley, Washington, D.C.

|  | Bioaccumulation Factors for Vegetables |  |
| :--- | :---: | :---: |
| cOPC | Transfer Equation | Source |
| Arsenic | $C p=0.03752 * C s$ | USEPA $(2007)$ |
| Copper | $\ln (\mathrm{Cp})=0.394 * \ln (\mathrm{Cs})+0.668$ | USEPA $(2007)$ |
| Nickel | $\ln (\mathrm{Cp})=0.748 * \ln (\mathrm{Cs})-2.223$ | USEPA $(2007)$ |
| Notes: |  |  |
|  | Cp - Concentration of contaminant in the homegrown vegetables |  |
|  | Cs - Concentration of contaminants in soil |  |

Table H-4
Oral and Inhalation Toxicity Values
4835 Glenbrook Road
Spring Valley, Washington, D.C.

|  | SF。 | URF |  |  | RfD。 |  |  | RfC |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| COPC | $\left(\mathrm{mg} / \mathrm{kg}\right.$-day) ${ }^{-1}$ | $\left(\mu \mathrm{g} / \mathrm{m}^{3}\right)^{-1}$ | Source | Date | (mg/kg-day) | Source | Date | $\left(\mu \mathrm{g} / \mathrm{m}^{3}\right)$ | Source | Date |
| Aluminum | - | - | - | - | $1.00 \mathrm{E}+00$ | PPRTV | Oct-06 | $5.00 \mathrm{E}+00$ | PPRTV | Oct-06 |
| Arsenic | 1.5 | 4.30E-03 | IRIS | May-09 | 3.00E-04 | IRIS | May-09 | 1.50E-02 | Cal EPA | May-09 |
| Cobalt | - | 9.00E-03 | PPRTV | Aug-08 | $3.00 \mathrm{E}-04$ | PPRTV | Aug-08 | 6.00E-03 | PPRTV | Aug-08 |
| Copper | - | - | - | - | $4.00 \mathrm{E}-02$ | HEAST | Jul-97 | - | - | - |
| Manganese | - | - | - | - | $1.40 \mathrm{E}-01$ | IRIS | May-09 | 5.00E-02 | IRIS | May-09 |
| Mercury | - | - | - | - | $3.00 \mathrm{E}-04$ | IRIS;1 | May-09 | 2.00E-01 | ATSDR | May-09 |
| Nickel | - | $2.60 \mathrm{E}-04$ | OEHHA | May-09 | $2.00 \mathrm{E}-02$ | IRIS | May-09 | 9.00E-02 | ATSDR | May-09 |
| Thallium | - | - | - | - | $8.00 \mathrm{E}-05$ | IRIS;2 | May-09 | - | - | - |
| Vanadium | - | - | - | - | $7.00 \mathrm{E}-03$ | HEAST | Jul-97 | - | - | - |

Notes:
1 - Mercuric chloride used.
2 - Thallium (I) sulfate used
Definitions:
ATSDR Agency for Toxic Substances and Disease Registry Minimal Risk Levels
Available online at: http://www.atsdr.cdc.gov/mrls/index.html
HEAST USEPA (1997b) Health Effects Assessment Tables
IRIS USEPA's Integrated Risk Information System. Available online at: http://cfpub.epa.gov/ncea/iris/index.cfm
OEHHA Office of Environmental Health Hazard Assessment Toxicity Criteria Database.
Available online at: http://www.oehha.org/risk/chemicalDB/index.asp
USEPA Provisional Peer Reviewed Toxicity Values
RfC Reference concentation
RfD Reference dose
SF Slope factor
URF Inhalation unit risk

Table H-5
Dermal Toxicity Values
4835 Glenbrook Road
Spring Valley, Washington, D.C.

| $\mathbf{S F}_{\mathrm{d}}$ |  | RfD $_{\mathrm{d}}$ | $\mathbf{D A F}^{1}$ | OAF |  |
| :--- | :---: | :---: | :---: | :---: | :---: |
| COPC | $(\mathbf{m g} / \mathrm{kg}-\mathrm{day})^{-1}$ | $(\mathrm{mg} / \mathrm{kg}-$ day $)$ | (unitless) | (unitless) | Source |
| Aluminum | - | $1.00 \mathrm{E}-01$ | - | 0.1 | Bast and Borges (1996) |
| Arsenic | $1.50 \mathrm{E}+00$ | $3.00 \mathrm{E}-04$ | 0.03 | 1 | USEPA (2004a) |
| Cobalt | - | $3.00 \mathrm{E}-04$ | - | 1 | USEPA (2004a) |
| Copper | - | $1.20 \mathrm{E}-02$ | - | 0.3 | Bast and Borges (1996) |
| Manganese | - | $5.60 \mathrm{E}-03$ | - | 0.04 | USEPA (2004a) |
| Mercury | - | $2.10 \mathrm{E}-05$ | - | 0.07 | USEPA (2004a) |
| Nickel | - | $8.00 \mathrm{E}-04$ | - | 0.04 | USEPA (2004a) |
| Thallium | - | $8.00 \mathrm{E}-05$ | - | 1 | USEPA (2004a) |
| Vanadium | - | $1.82 \mathrm{E}-04$ | - | 0.026 | USEPA (2004a) |

Notes:
1 - From USEPA (2004a).

## Definitions:

| DAF | Dermal absorption fraction from soil |
| :--- | :--- |
| OAF | Oral absorption fraction |
| $\mathrm{RfD}_{\mathrm{d}}$ | Dermal reference dose, which equals RfD |
| $\mathrm{RfD}_{0}$ | OAF |
| $\mathrm{SF}_{\mathrm{d}}$ | Oral reference dose |
| $\mathrm{SF}_{\mathrm{o}}$ | Dermal slope factor, which equals $\mathrm{SF}_{0} / \mathrm{OAF}$ |
|  | Oral slope factor |

Table H. 6
Risks and Noncancer Hazards from Assumed Exposures to Arsenic
4835 Glenbrook Road
Spring Valley, Washington, D.C.
Carcinogenic Risks

| Scenario Depth (ft bgs) | Receptor | EPC | Ingestion | Dermal | Inhalation (Dust) | Vegetable ingestion | Total |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| RME Exposures |  |  |  |  |  |  |  |
| 0-2 | Adult resident | 10.55 | 9.29E-06 | 1.11E-06 | 3.60E-10 | 8.16E-06 | 2.E-05 |
| 0-10 | Adult resident | 11.17 | 9.84E-06 | 1.18E-06 | $3.82 \mathrm{E}-10$ | $8.64 \mathrm{E}-06$ | 2.E-05 |
| 0-2 | Child resident | 10.55 | 8.67E-06 | 1.46E-06 | 8.53E-11 | $1.63 \mathrm{E}-06$ | 1.E-05 |
| 0-10 | Child resident | 11.17 | 9.18E-06 | 1.54E-06 | $9.04 \mathrm{E}-11$ | $1.73 \mathrm{E}-06$ | 1.E-05 |
| 0-2 | Outdoor worker | 10.55 | $2.65 \mathrm{E}-05$ | 3.83E-07 | 3.43E-09 | - | 3.E-05 |
| 0-10 | Outdoor worker | 11.17 | 2.81E-05 | 4.06E-07 | 3.63E-09 | - | 3.E-05 |
| Central Tendency Exposures |  |  |  |  |  |  |  |
| 0-2 | Adult resident | 9.4 | 8.28E-06 | $9.91 \mathrm{E}-07$ | $3.21 \mathrm{E}-10$ | $7.27 \mathrm{E}-06$ | 2.E-05 |
| 0-10 | Adult resident | 9.1 | 8.01E-06 | 9.59E-07 | $3.11 \mathrm{E}-10$ | $7.04 \mathrm{E}-06$ | 2.E-05 |
| 0-2 | Child resident | 9.4 | 7.73E-06 | 1.30E-06 | $7.60 \mathrm{E}-11$ | $1.45 \mathrm{E}-06$ | 1.E-05 |
| 0-10 | Child resident | 9.1 | 7.48E-06 | 1.26E-06 | $7.36 \mathrm{E}-11$ | $1.41 \mathrm{E}-06$ | 1.E-05 |
| 0-2 | Outdoor worker | 9.4 | 2.37E-05 | 3.41E-07 | 3.06E-09 | - | 2.E-05 |
| 0-10 | Outdoor worker | 9.1 | 2.29E-05 | 3.31E-07 | $2.96 \mathrm{E}-09$ | - | 2.E-05 |

Hazard Index

| Scenario | Depth (ft bgs) | Receptor | EPC | Ingestion | Dermal | Inhalation (Dust) | Vegetable ingestion | Total |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| RME Exposures |  |  |  |  |  |  |  |  |
|  | 0-2 | Adult resident | 10.55 | 4.82E-02 | $5.77 \mathrm{E}-03$ | $1.30 \mathrm{E}-05$ | 4.23E-02 | 0.10 |
|  | 0-10 | Adult resident | 11.17 | 5.10E-02 | 6.11E-03 | $1.38 \mathrm{E}-05$ | 4.48E-02 | 0.10 |
|  | 0-2 | Child resident | 10.55 | 2.25E-01 | $3.78 \mathrm{E}-02$ | $1.54 \mathrm{E}-05$ | 4.23E-02 | 0.30 |
|  | 0-10 | Child resident | 11.17 | 2.38E-01 | $4.00 \mathrm{E}-02$ | 1.63E-05 | $4.48 \mathrm{E}-02$ | 0.32 |
|  | 0-2 | Outdoor worker | 10.55 | 1.65E-01 | $2.38 \mathrm{E}-03$ | $1.49 \mathrm{E}-04$ | - | 0.17 |
|  | 0-10 | Outdoor worker | 11.17 | $1.75 \mathrm{E}-01$ | 2.52E-03 | $1.58 \mathrm{E}-04$ | - | 0.18 |
| Central Tendency Exposures |  |  |  |  |  |  |  |  |
|  | 0-2 | Adult resident | 9.4 | 4.29E-02 | 5.14E-03 | 1.16E-05 | $3.77 \mathrm{E}-02$ | 0.09 |
|  | 0-10 | Adult resident | 9.1 | 4.16E-02 | $4.97 \mathrm{E}-03$ | $1.12 \mathrm{E}-05$ | 3.65E-02 | 0.08 |
|  | 0-2 | Child resident | 9.4 | 2.00E-01 | 3.37E-02 | $1.38 \mathrm{E}-05$ | $3.77 \mathrm{E}-02$ | 0.27 |
|  | 0-10 | Child resident | 9.1 | $1.94 \mathrm{E}-01$ | 3.26E-02 | $1.33 \mathrm{E}-05$ | 3.65E-02 | 0.26 |
|  | 0-2 | Outdoor worker | 9.4 | $1.47 \mathrm{E}-01$ | 2.12E-03 | 1.33E-04 | - | 0.15 |
|  | 0-10 | Outdoor worker | 9.1 | 1.42E-01 | $2.06 \mathrm{E}-03$ | 1.29E-04 | - | 0.14 |

Table H.7.1
Adult Resident Risk Estimates
Surface Soils ( $0-2 \mathrm{ft} \mathrm{bgs}$ ) and Mixed Soils ( $0-10 \mathrm{ft} \mathrm{bgs}$ )
4835 Gl enbrook Road
4835 Glenbrook Road
Spring Valley, Washington, D.C.

| COPC | RME Risk Probabilities |  |  |  |  |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | Surface Soils (0-2 ft bgs) |  |  |  | Summation | Percent Contribution | Mixed Soils (0-10 ft bgs) |  |  |  | Summation | Percent Contribution |
|  | Ingestion | Dermal Contact | Inhalation of VOC/Dust in Outdoor Air | Ingestion of Home-Grown Vegetables |  |  | Ingestion | Dermal Contact | Inhalation of vOC/Dust in Outdoor Air | Ingestion of Home-Grown Vegetables |  |  |
| Aluminum | - | - | - | - |  | - |  |  |  |  |  | - |
| Arsenic | 9.29E-06 | 1.11E-06 | 3.60E-10 | 8.16E-06 | $1.86 \mathrm{E}-05$ | 100\% | $9.84 \mathrm{E}-06$ | $1.18 \mathrm{E}-06$ | $3.82 \mathrm{E}-10$ | $8.64 \mathrm{E}-06$ | $1.97 \mathrm{E}-05$ | 100\% |
| Cobalt | - | - | 3.00E-09 | - | 3.00E-09 | 0\% | - | - | 3.00E-09 | - | 3.00E-09 | 0\% |
| Copper | - | - | - | - | - | - | - | - | - | - | - | - |
| Manganese | - | - | - | - | - | - | - | - | - | - | - | - |
| Mercury | - | - | - | - | - | - | - | - | - | - | - | - |
| Nickel | - | - | $1.52 \mathrm{E}-10$ | - | $1.52 \mathrm{E}-10$ | 0\% | - | - | 1.49E-10 | - | 1.49E-10 | 0\% |
| Thallium | - | - | - | - | - | - | - | - | - | - | - | - |
| Vanadium | - | - | - | - | - | - | - |  | - | - |  | - |
| Summation | 9E-06 | 1E-06 | 4E-09 | 8E-06 | $2 \mathrm{E}-05$ |  | 1E-05 | 1E-06 | 4E-09 | 9E-06 | $2 \mathrm{E}-05$ |  |


| COPC | RME Hazard Index (HI) |  |  |  |  |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | Surface Soils (0-2 ft bgs) |  |  |  | Summation |  | Mixed Soils (0-10 ft bgs) |  |  |  | Summation | Percent Contribution |
|  | Ingestion | Dermal Contact | Inhalation of VOC/Dust in Outdoor Air | Ingestion of Home-Grown Vegetables |  |  | Ingestion | Dermal Contact | Inhalation of vOC/Dust in Outdoor Air | Ingestion of Home-Grown Vegetables |  |  |
| Aluminum | 3.17E-02 | - | $8.57 \mathrm{E}-05$ | - | 3.18E-02 | 8\% | 3.50E-02 | - | 9.47E-05 | - | 3.51E-02 | 9\% |
| Arsenic | $4.82 \mathrm{E}-02$ | 5.77E-03 | $1.30 \mathrm{E}-05$ | $4.23 \mathrm{E}-02$ | $9.63 \mathrm{E}-02$ | 25\% | $5.10 \mathrm{E}-02$ | 6.11E-03 | $1.38 \mathrm{E}-05$ | $4.48 \mathrm{E}-02$ | $1.02 \mathrm{E}-01$ | 25\% |
| Cobalt | 1.92E-01 | - | 1.30E-04 | - | $1.92 \mathrm{E}-01$ | 50\% | $1.92 \mathrm{E}-01$ | - | 1.30E-04 | - | $1.92 \mathrm{E}-01$ | 47\% |
| Copper | 2.71E-03 | - |  | $8.76 \mathrm{E}-03$ | $1.15 \mathrm{E}-02$ | 3\% | $3.68 \mathrm{E}-03$ | . |  | $9.88 \mathrm{E}-03$ | $1.36 \mathrm{E}-02$ | 3\% |
| Manganese | 5.91E-03 | - | 2.24E-04 | - | $6.13 \mathrm{E}-03$ | 2\% | 7.56E-03 | - | $2.87 \mathrm{E}-04$ | - | $7.85 \mathrm{E}-03$ | 2\% |
| Mercury | 6.66E-04 | - | 1.35E-08 | - | $6.66 \mathrm{E}-04$ | 0\% | 5.49E-04 | - | $1.11 \mathrm{E}-08$ | - | $5.49 \mathrm{E}-04$ | 0\% |
| Nickel | $5.05 \mathrm{E}-03$ | - | $1.52 \mathrm{E}-05$ | 4.33E-03 | 9.40E-03 | 2\% | $4.93 \mathrm{E}-03$ | - | $1.48 \mathrm{E}-05$ | 4.25E-03 | $9.20 \mathrm{E}-03$ | 2\% |
| Thallium | 1.86E-02 | - | - | - | $1.86 \mathrm{E}-02$ | 5\% | $2.31 \mathrm{E}-02$ | - | - | - | $2.31 \mathrm{E}-02$ | 6\% |
| Vanadium | 1.84E-02 | - | - | - | 1.84E-02 | 5\% | $2.12 \mathrm{E}-02$ | - | - | - | $2.12 \mathrm{E}-02$ | 5\% |
| Summation | 3E-01 | 6E-03 | 5E-04 | 6E-02 | 4E-01 |  | 3E-01 | 6E-03 | 5E-04 | 6E-02 | $4 \mathrm{E}-01$ |  |


| COPC | CT Risk Probabilities |  |  |  |  |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | Surface Soils (0-2 ft bgs) |  |  |  | Summation | Percent Contribution | Mixed Soils (0-10 ft bgs) |  |  |  | Summation | $\begin{gathered} \text { Percent } \\ \text { Contribution } \\ \hline \end{gathered}$ |
|  | Ingestion | Dermal Contact | Inhalation of voc/Dust in Outdoor Air | Ingestion of Home-Grown Vegetables |  |  | Ingestion | Dermal Contact | Inhalation of voc/Dust in Outdoor Air | Ingestion of Home-Grown Vegetables |  |  |
| Aluminum |  | - | - | - |  |  |  |  |  |  |  | - |
| Arsenic | $8.28 \mathrm{E}-06$ | 9.91E-07 | 3.21E-10 | $7.27 \mathrm{E}-06$ | 1.65E-05 | 100\% | 8.01E-06 | 9.59E-07 | 3.11E-10 | 7.04E-06 | 1.60E-05 | 100\% |
| Cobalt |  | - | $9.01 \mathrm{E}-10$ | - | $9.01 \mathrm{E}-10$ | 0\% |  |  | 6.01E-10 |  | 6.01E-10 | 0\% |
| Copper | - | - | - | - | - |  | . | . | - | - | - | - |
| Manganese | . | - | . | . | - | . | . | . | . | . | - | . |
| Mercury | - | - | - | - | - | - | - | - | - | - | - | - |
| Nickel | - | - | 3.95E-11 | - | 3.95E-11 | 0\% | - | - | 4.09E-11 | - | 4.09E-11 | 0\% |
| Thallium | - | - | - | - | - | - | - | - | - | - | - | - |
| Vanadium |  | - | - | - | - | - | - | - | - | . | - | - |
| Summation | 8E-06 | 1E-06 | 1E-09 | 7E-06 | $2 \mathrm{E}-05$ |  | 8E-06 | 1E-06 | 1E-09 | 7E-06 | $2 \mathrm{E}-05$ |  |


| COPC | CT Hazard Index (HI) |  |  |  |  |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | Surface Soils ( $0-2 \mathrm{ft} \mathrm{bgs}$ ) |  |  |  | Summation | Percent Contribution | Mixed Soils ( $0-10 \mathrm{ft} \mathrm{bgs}$ ) |  |  |  | Summation | Percent Contribution |
|  | Ingestion | Dermal Contact | Inhalation of voc/Dust in Outdoor Air | Ingestion of Home-Grown Vegetables |  |  | Ingestion | Dermal Contact | Inhalation of voc/Dust in Outdoor Air | Ingestion of Home-Grown Vegetables |  |  |
| Aluminum | 1.44E-02 | - | $7.79 \mathrm{E}-05$ | - | 1.45E-02 | 7\% | $1.65 \mathrm{E}-02$ | - | $8.91 \mathrm{E}-05$ |  | $1.65 \mathrm{E}-02$ | 9\% |
| Arsenic | 4.29E-02 | 5.14E-03 | $1.16 \mathrm{E}-05$ | 3.77E-02 | $8.58 \mathrm{E}-02$ | 39\% | $4.16 \mathrm{E}-02$ | 4.97E-03 | 1.12E-05 | 3.65E-02 | $8.30 \mathrm{E}-02$ | 43\% |
| Cobalt | 9.59E-02 | - | 1.30E-04 | - | 9.60E-02 | 43\% | $6.39 \mathrm{E}-02$ | - | 8.65E-05 |  | $6.40 \mathrm{E}-02$ | 33\% |
| Copper | 1.20E-03 | - | - | 1.39E-03 | $2.59 \mathrm{E}-03$ | 1\% | $1.07 \mathrm{E}-03$ | - | - | $1.33 \mathrm{E}-03$ | $2.40 \mathrm{E}-03$ | 1\% |
| Manganese | $2.66 \mathrm{E}-03$ | - | $2.01 \mathrm{E}-04$ |  | $2.86 \mathrm{E}-03$ | 1\% | $3.27 \mathrm{E}-03$ | - | $2.48 \mathrm{E}-04$ |  | 3.52E-03 | 2\% |
| Mercury | $2.45 \mathrm{E}-04$ | - | 9.94E-09 | - | $2.45 \mathrm{E}-04$ | 0\% | $2.28 \mathrm{E}-04$ | - | 9.24E-09 | - | $2.28 \mathrm{E}-04$ | 0\% |
| Nickel | $2.18 \mathrm{E}-03$ | - | $1.31 \mathrm{E}-05$ | 6.46E-04 | $2.84 \mathrm{E}-03$ | 1\% | $2.26 \mathrm{E}-03$ | - | $1.36 \mathrm{E}-05$ | $6.63 \mathrm{E}-04$ | $2.94 \mathrm{E}-03$ | 2\% |
| Thallium | 8.26E-03 | - | - | - | 8.26E-03 | 4\% | 1.00E-02 | - | - | - | $1.00 \mathrm{E}-02$ | 5\% |
| Vanadium | 8.20E-03 | - | - | - | 8.20E-03 | 4\% | 9.17E-03 | - | - | - | $9.17 \mathrm{E}-03$ | 5\% |
| Summation | $2 \mathrm{E}-01$ | 5E-03 | 4E-04 | 4E-02 | 2E-01 |  | 1E-01 | 5E-03 | 4E-04 | 4E-02 | $2 \mathrm{E}-01$ |  |

Child Resident Risk Estimates
Surface Soils ( $0-2 \mathrm{ft} \mathrm{bgs}$ ) and Mixed Soils ( $0-10 \mathrm{ft} \mathrm{bgs}$ )
4835 Gl enbrook Road
4835 Glenbrook Road
Spring Valley, Washington, D.C.

| COPC | RME Risk Probabilities |  |  |  |  |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | Surface Soils (0-2 ft bgs) |  |  |  | Summation | Percent Contribution | Mixed Soils (0-10 ft bgs) |  |  |  | Summation | Percent Contribution |
|  | Ingestion | Dermal Contact | Inhalation of vOC/Dust in Outdoor Air | Ingestion of Home-Grown Vegetables |  |  | Ingestion | Dermal Contact | Inhalation of VOC/Dust in Outdoor Air | Ingestion of Home-Grown Vegetables |  |  |
| Aluminum | - | - | - | - | - | - | - | - | - | - | - | - |
| Arsenic | 8.67E-06 | $1.46 \mathrm{E}-06$ | 8.53E-11 | 1.63E-06 | 1.18E-05 | 100\% | $9.18 \mathrm{E}-06$ | 1.54E-06 | 9.04E-11 | 1.73E-06 | 1.25E-05 | 100\% |
| Cobalt | - | - | 7.11E-10 | - | 7.11E-10 | 0\% | - | - | 7.11E-10 | - | 7.11E-10 | 0\% |
| Copper | - | - | - | - | - | - | - | - | - | - | - | - |
| Manganese | - | - | - | - | - | - | - | - | - | - | - | - |
| Mercury | - | - | - | - | - | - | - | - | - | - | - | - |
| Nickel | - | - | 3.61E-11 | . | 3.61E-11 | 0\% | - | - | 3.52E-11 | - | 3.52E-11 | 0\% |
| Thallium | - | - | - | - | - | - | - | - | - | - | - | - |
| Vanadium | - | - | - | - | - | - | - | - | - | - | - | - |
| Summation | 9E-06 | 1E-06 | 8E-10 | 2E-06 | 1E-05 |  | 9E-06 | $2 \mathrm{E}-06$ | 8E-10 | 2E-06 | 1E-05 |  |


|  |  |  |  |  |  | RME Haza | Index (HI) |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  |  | Surface | Soils (0-2 ft bgs) |  |  |  |  | Mixed | oils (0-10 ft bgs |  |  |  |
| COPC | Ingestion | Dermal Contact | Inhalation of voc/Dust in Outdoor Air | Ingestion of Home-Grown Vegetables | Summation | Percent Contribution | Ingestion | Dermal Contact | Inhalation of voc/Dust in Outdoor Air | Ingestion of Home-Grown Vegetables | Summation | Percent Contribution |
| Aluminum | 1.48E-01 | - | 1.01E-04 |  | $1.48 \mathrm{E}-01$ | 9\% | 1.63E-01 |  | 1.12E-04 |  | $1.63 \mathrm{E}-01$ | 10\% |
| Arsenic | $2.25 \mathrm{E}-01$ | $3.78 \mathrm{E}-02$ | $1.54 \mathrm{E}-05$ | 4.23E-02 | 3.05E-01 | 19\% | $2.38 \mathrm{E}-01$ | 4.00E-02 | $1.63 \mathrm{E}-05$ | $4.48 \mathrm{E}-02$ | 3.23E-01 | 19\% |
| Cobalt | $8.95 \mathrm{E}-01$ | - | 1.54E-04 |  | 8.95E-01 | 56\% | 8.95E-01 |  | $1.54 \mathrm{E}-04$ |  | 8.95E-01 | 53\% |
| Copper | 1.27E-02 | - | - | 8.76E-03 | $2.14 \mathrm{E}-02$ | 1\% | 1.72E-02 | - |  | $9.88 \mathrm{E}-03$ | $2.71 \mathrm{E}-02$ | 2\% |
| Manganese | 2.76E-02 | - | $2.65 \mathrm{E}-04$ |  | $2.78 \mathrm{E}-02$ | 2\% | 3.53E-02 |  | 3.39E-04 |  | 3.56E-02 | 2\% |
| Mercury | 3.11E-03 | - | $1.60 \mathrm{E}-08$ |  | 3.11E-03 | 0\% | 2.56E-03 |  | $1.32 \mathrm{E}-08$ |  | $2.56 \mathrm{E}-03$ | 0\% |
| Nickel | $2.36 \mathrm{E}-02$ | - | $1.80 \mathrm{E}-05$ | $4.33 \mathrm{E}-03$ | $2.79 \mathrm{E}-02$ | 2\% | $2.30 \mathrm{E}-02$ |  | $1.75 \mathrm{E}-05$ | 4.25E-03 | $2.73 \mathrm{E}-02$ | 2\% |
| Thallium | 8.70E-02 | - | - |  | 8.70E-02 | 5\% | 1.08E-01 | - |  |  | $1.08 \mathrm{E}-01$ | 6\% |
| Vanadium | $8.61 \mathrm{E}-02$ | - | - |  | 8.61E-02 | 5\% | 9.91E-02 | - | - | . | $9.91 \mathrm{E}-02$ | 6\% |
| Summation | $2 \mathrm{E}+00$ | 4E-02 | $6 \mathrm{E}-04$ | 6E-02 | $2 \mathrm{E}+00$ |  | $2 \mathrm{E}+00$ | 4E-02 | $6 \mathrm{E}-04$ | 6E-02 | $2 \mathrm{E}+00$ |  |


| COPC | CT Risk Probabilities |  |  |  |  |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | Surface Soils (0-2 ft bgs) |  |  |  | Summation | $\begin{array}{\|c\|} \hline \text { Percent } \\ \text { Contribution } \\ \hline \end{array}$ | Mixed Soils ( $0-10 \mathrm{ft} \mathrm{bgs}$ ) |  |  |  | Summation | Percent Contribution |
|  | Ingestion | Dermal | Inhalation of vOC/Dust in Outdoor Air | Ingestion of Home-Grown Vegetables |  |  | Ingestion | $\begin{aligned} & \text { Dermal } \\ & \text { Contact } \end{aligned}$ | Inhalation of vOC/Dust in Outdoor Air | Ingestion of Home-Grown Vegetables |  |  |
| Aluminum | - | - | - | - | - | - | - | - | - | - | - | - |
| Arsenic | 7.73E-06 | 1.30E-06 | 7.60E-11 | $1.45 \mathrm{E}-06$ | 1.05E-05 | 100\% | 7.48E-06 | 1.26E-06 | 7.36E-11 | $1.41 \mathrm{E}-06$ | 1.01E-05 | 100\% |
| Cobalt | - | - | 7.11E-10 | - | 7.11E-10 | 0\% | - | - | 4.74E-10 | - | 4.74E-10 | 0\% |
| Copper | - | - | - | - | - | - | - | - | - | - | - | - |
| Manganese | - | - | - | - | - | - | - | - | - | - | - | - |
| Mercury | - | - | - | - | - | - | - | - | - | - | - | - |
| Nickel | - | - | 3.12E-11 | - | 3.12E-11 | 0\% | - | - | 3.23E-11 | - | 3.23E-11 | 0\% |
| Thallium | - | - | - | . | - | - | - | - | - | - | - | - |
| Vanadium | - | - | - | - | - | - | - | - | - | - | - | - |
| Summation | 8E-06 | 1E-06 | 8E-10 | 1E-06 | 1E-05 |  | 7E-06 | 1E-06 | 6E-10 | 1E-06 | 1E-05 |  |


| COPC | CT Hazard Index (HI) |  |  |  |  |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | Surface Soils ( $0-2 \mathrm{ft} \mathrm{bgs}$ ) |  |  |  | Summation | Percent Contribution | Mixed Soils (0-10 ft bgs) |  |  |  | Summation | Percent Contribution |
|  | Ingestion | Dermal Contact | Inhalation of vOC/Dust in Outdoor Air | Ingestion of Home-Grown Vegetables |  |  | Ingestion | Dermal Contact | Inhalation of voc/Dust in Outdoor Air | Ingestion of Home-Grown Vegetables |  |  |
| Aluminum | 1.34E-01 |  | 9.22E-05 |  | $1.34 \mathrm{E}-01$ | 9\% | 1.54E-01 |  | 1.05E-04 |  | 1.54E-01 | 12\% |
| Arsenic | $2.00 \mathrm{E}-01$ | 3.37E-02 | $1.38 \mathrm{E}-05$ | 3.77E-02 | $2.72 \mathrm{E}-01$ | 18\% | 1.94E-01 | 3.26E-02 | $1.33 \mathrm{E}-05$ | 3.65E-02 | 2.63E-01 | 21\% |
| Cobalt | $8.95 \mathrm{E}-01$ | - | $1.54 \mathrm{E}-04$ |  | $8.95 \mathrm{E}-01$ | 59\% | 5.97E-01 |  | 1.02E-04 |  | 5.97E-01 | 47\% |
| Copper | 1.12E-02 | - |  | 1.39E-03 | 1.26E-02 | 1\% | 9.99E-03 |  |  | 1.33E-03 | 1.13E-02 | 1\% |
| Manganese | $2.48 \mathrm{E}-02$ | - | $2.38 \mathrm{E}-04$ | - | $2.50 \mathrm{E}-02$ | 2\% | 3.06E-02 | - | $2.94 \mathrm{E}-04$ | - | $3.09 \mathrm{E}-02$ | 2\% |
| Mercury | $2.28 \mathrm{E}-03$ | - | $1.18 \mathrm{E}-08$ | - | $2.28 \mathrm{E}-03$ | 0\% | 2.12E-03 | - | $1.09 \mathrm{E}-08$ | - | 2.12E-03 | 0\% |
| Nickel | $2.04 \mathrm{E}-02$ | - | $1.55 \mathrm{E}-05$ | 6.46E-04 | $2.10 \mathrm{E}-02$ | 1\% | $2.11 \mathrm{E}-02$ | - | 1.61E-05 | 6.63E-04 | 2.18E-02 | 2\% |
| Thallium | 7.71E-02 | - | - | - | 7.71E-02 | 5\% | 9.36E-02 | - | - | - | 9.36E-02 | 7\% |
| Vanadium | $7.65 \mathrm{E}-02$ | - | - | - | $7.65 \mathrm{E}-02$ | 5\% | 8.56E-02 | - | - | - | 8.56E-02 | 7\% |
| Summation | 1E+00 | 3E-02 | 5E-04 | 4E-02 | 2E+00 |  | $1 \mathrm{E}+00$ | 3E-02 | 5E-04 | 4E-02 | 1E+00 |  |

Table H.7.3
Outdoor Worker Risk Estimates
Surface Soils (0-2 ft bgs) and Mixed Soils (0-10 ft bgs) 4835 Glenbrook Road
Spring Valley, Washington, D.C.

| COPC | RME Risk Probabilities |  |  |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | Surface Soils (0-2 ft bgs) |  |  | Summation | Percent Contribution | Mixed Soils (0-10 ft bgs) |  |  | Summation | Percent Contribution |
|  | Ingestion | Dermal Contact | Inhalation of VOC/Dust in Outdoor Air |  |  | Ingestion | Dermal Contact | Inhalation of VOC/Dust in Outdoor Air |  |  |
| Aluminum | - | - | - | - | - | - | - | - | - | - |
| Arsenic | 2.65E-05 | 3.83E-07 | 3.43E-09 | 2.69E-05 | 100\% | 2.81E-05 | 4.06E-07 | 3.63E-09 | 2.85E-05 | 100\% |
| Cobalt | - | - | $2.86 \mathrm{E}-08$ | $2.86 \mathrm{E}-08$ | 0\% | - | - | $2.86 \mathrm{E}-08$ | $2.86 \mathrm{E}-08$ | 0\% |
| Copper | - | - | - | - | - | - | - | - | - | - |
| Manganese | - | - | - | - | - | - | - | - | - | - |
| Mercury | - | - | - | - | - | - | - | - | - | - |
| Nickel | - | - | 1.45E-09 | 1.45E-09 | 0\% | - | - | 1.42E-09 | 1.42E-09 | 0\% |
| Thallium | - | - | - | - | - | - | - | - | - | - |
| Vanadium | - | - | - | - | - | - | - | - | - | - |
| Summation | 3E-05 | 4E-07 | 3E-08 | 3E-05 |  | 3E-05 | 4E-07 | 3E-08 | 3E-05 |  |


| COPC | RME Hazard Index (HI) |  |  |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | Surface Soils (0-2 ft bgs) |  |  | Summation | Percent Contribution | Mixed Soils (0-10 ft bgs) |  |  | Summation | Percent Contribution |
|  | Ingestion | Dermal Contact | Inhalation of voc/Dust in Outdoor Air |  |  | Ingestion | Dermal Contact | Inhalation of VOC/Dust in Outdoor Air |  |  |
| Aluminum | 1.09E-01 | - | $9.79 \mathrm{E}-04$ | $1.10 \mathrm{E}-01$ | 10\% | 1.20E-01 | - | $1.08 \mathrm{E}-03$ | $1.21 \mathrm{E}-01$ | 10\% |
| Arsenic | $1.65 \mathrm{E}-01$ | 2.38E-03 | $1.49 \mathrm{E}-04$ | $1.68 \mathrm{E}-01$ | 15\% | 1.75E-01 | 2.52E-03 | $1.58 \mathrm{E}-04$ | $1.78 \mathrm{E}-01$ | 15\% |
| Cobalt | $6.58 \mathrm{E}-01$ | - | $1.48 \mathrm{E}-03$ | $6.59 \mathrm{E}-01$ | 59\% | 6.58E-01 |  | $1.48 \mathrm{E}-03$ | $6.59 \mathrm{E}-01$ | 56\% |
| Copper | $9.30 \mathrm{E}-03$ | - | - | $9.30 \mathrm{E}-03$ | 1\% | 1.26E-02 | - | - | $1.26 \mathrm{E}-02$ | 1\% |
| Manganese | 2.03E-02 | - | $2.56 \mathrm{E}-03$ | $2.28 \mathrm{E}-02$ | 2\% | 2.59E-02 | - | 3.28E-03 | $2.92 \mathrm{E}-02$ | 2\% |
| Mercury | $2.28 \mathrm{E}-03$ | - | $1.55 \mathrm{E}-07$ | $2.28 \mathrm{E}-03$ | 0\% | $1.88 \mathrm{E}-03$ | - | $1.27 \mathrm{E}-07$ | $1.88 \mathrm{E}-03$ | 0\% |
| Nickel | 1.73E-02 | - | $1.74 \mathrm{E}-04$ | $1.75 \mathrm{E}-02$ | 2\% | 1.69E-02 | - | $1.69 \mathrm{E}-04$ | $1.71 \mathrm{E}-02$ | 1\% |
| Thallium | 6.39E-02 | - | - | 6.39E-02 | 6\% | 7.90E-02 | - | - | 7.90E-02 | 7\% |
| Vanadium | 6.32E-02 | - | - | 6.32E-02 | 6\% | 7.28E-02 | - | - | $7.28 \mathrm{E}-02$ | 6\% |
| Summation | 1E+00 | 2E-03 | 5E-03 | 1E+00 |  | 1E+00 | 3E-03 | 6E-03 | 1E+00 |  |


| COPC | CT Risk Probabilities |  |  |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | Surface Soils (0-2 ft bgs ) |  |  | Summation | Percent Contribution | Mixed Soils (0-10 ft bgs) |  |  | Summation | Percent Contribution |
|  | Ingestion | Dermal Contact | Inhalation of VOC/Dust in Outdoor Air |  |  | Ingestion | Dermal Contact | Inhalation of VOC/Dust in Outdoor Air |  |  |
| Aluminum | - | - | - | - | - | - | - | - | - | - |
| Arsenic | $2.37 \mathrm{E}-05$ | 3.41E-07 | 3.06E-09 | $2.40 \mathrm{E}-05$ | 100\% | 2.29E-05 | 3.31E-07 | 2.96E-09 | 2.32E-05 | 100\% |
| Cobalt | - | - | 3.00E-09 | 3.00E-09 | 0\% | - | - | $2.00 \mathrm{E}-09$ | $2.00 \mathrm{E}-09$ | 0\% |
| Copper | - | - | - | - | - | - | - | - | - | - |
| Manganese | - | - | - | - | - | - | - | - | - | - |
| Mercury | - | - | - | - | - | - | - | - | - | - |
| Nickel | - | - | $1.32 \mathrm{E}-10$ | $1.32 \mathrm{E}-10$ | 0\% | - | - | $1.36 \mathrm{E}-10$ | $1.36 \mathrm{E}-10$ | 0\% |
| Thallium | - | - | - | - | - | - | - | - | - | - |
| Vanadium | - | - | - | - | - | - | - | - | - | - |
| Summation | 2E-05 | 3E-07 | 6E-09 | 2E-05 |  | 2E-05 | 3E-07 | 5E-09 | 2E-05 |  |


| COPC | CT Hazard Index (HI) |  |  |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | Surface Soils (0-2 ft bgs) |  |  | Summation | Percent Contribution | Mixed Soils (0-10 ft bgs) |  |  | Summation | Percent Contribution |
|  | Ingestion | Dermal Contact | Inhalation of VOC/Dust in Outdoor Air |  |  | Ingestion | Dermal Contact | Inhalation of VOC/Dust in Outdoor Air |  |  |
| Aluminum | 1.80E-02 | - | $2.60 \mathrm{E}-04$ | 1.83E-02 | 6\% | 2.06E-02 | - | 2.97E-04 | 2.09E-02 | 7\% |
| Arsenic | $1.47 \mathrm{E}-01$ | 2.12E-03 | $1.33 \mathrm{E}-04$ | $1.49 \mathrm{E}-01$ | 47\% | 1.42E-01 | 2.06E-03 | $1.29 \mathrm{E}-04$ | $1.45 \mathrm{E}-01$ | 52\% |
| Cobalt | 1.20E-01 | - | $4.33 \mathrm{E}-04$ | $1.20 \mathrm{E}-01$ | 38\% | 8.00E-02 | - | $2.88 \mathrm{E}-04$ | 8.03E-02 | 29\% |
| Copper | 1.51E-03 | - | - | $1.51 \mathrm{E}-03$ | 0\% | $1.34 \mathrm{E}-03$ | - | - | $1.34 \mathrm{E}-03$ | 0\% |
| Manganese | 3.32E-03 | - | $6.71 \mathrm{E}-04$ | $3.99 \mathrm{E}-03$ | 1\% | 4.10E-03 | - | 8.27E-04 | $4.93 \mathrm{E}-03$ | 2\% |
| Mercury | 3.06E-04 | - | 3.31E-08 | 3.06E-04 | 0\% | 2.85E-04 | - | 3.08E-08 | $2.85 \mathrm{E}-04$ | 0\% |
| Nickel | $2.73 \mathrm{E}-03$ | - | $4.38 \mathrm{E}-05$ | $2.77 \mathrm{E}-03$ | 1\% | 2.83E-03 | - | $4.54 \mathrm{E}-05$ | $2.88 \mathrm{E}-03$ | 1\% |
| Thallium | 1.03E-02 | - | - | $1.03 \mathrm{E}-02$ | 3\% | 1.25E-02 | - | - | $1.25 \mathrm{E}-02$ | 4\% |
| Vanadium | 1.03E-02 | - | - | $1.03 \mathrm{E}-02$ | 3\% | 1.15E-02 | - | - | $1.15 \mathrm{E}-02$ | 4\% |
| Summation | 3E-01 | 2E-03 | 2E-03 | 3E-01 |  | 3E-01 | 2E-03 | 2E-03 | 3E-01 |  |

Table H.7.4
COPC Toxic Endpoints
4835 Glenbrook Road
Spring Valley, Washington, D.C.

| COPC | No Adverse Effects | Autoimmune effects | Respiratory Effects | Developmental Effects | Thyroid Effects | Central Nervous System (CNS) | Decreased Body and Organ Weight | Skin | Cardiovascular Effects | Gastrointestinal Effects | Hematopoietic Effects |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Ingestion |  |  |  |  |  |  |  |  |  |  |  |
| Aluminum |  |  |  | X |  | X |  |  |  |  |  |
| Arsenic |  |  |  |  |  |  |  | X | X |  |  |
| Cobalt |  |  |  | X | X |  |  |  |  |  | X |
| Copper |  |  |  |  |  |  |  |  |  | X |  |
| Manganese |  |  |  |  |  | X |  |  |  |  |  |
| Mercury |  | X |  |  |  |  |  |  |  |  |  |
| Nickel |  |  |  |  |  |  | X |  |  |  |  |
| Thallium | X |  |  |  |  |  |  |  |  |  |  |
| Vanadium | X |  |  |  |  |  |  |  |  |  |  |
| Inhalation |  |  |  |  |  |  |  |  |  |  |  |
| Aluminum |  |  |  |  |  | X |  |  |  |  |  |
| Arsenic |  |  | X | X |  | X |  |  | X |  |  |
| Cobalt |  |  | X |  |  |  |  |  |  |  |  |
| Copper |  |  |  |  |  |  |  |  |  |  |  |
| Manganese |  |  |  |  |  | X |  |  |  |  |  |
| Mercury |  |  |  |  |  | X |  |  |  |  |  |
| Nickel |  |  | X |  |  |  |  |  |  |  |  |
| Thallium |  |  |  |  |  |  |  |  |  |  |  |
| Vanadium |  |  |  |  |  |  |  |  |  |  |  |

Table H.7.5
RME Child Residential Hazard Indices by Toxic Endpoint 4835 Glenbrook Road
Spring Valley, Washington, D.C.

| Toxic Endpoint | Child Resident |  |
| :--- | :---: | :---: |
|  | Surface Soil | Mixed Soil |
| No adverse effects | 0.2 | 0.2 |
| Autoimmune effects | 0.003 | 0.003 |
| Respiratory | $2 \mathrm{E}-04$ | $2 \mathrm{E}-04$ |
| Developmental Effects | 1 | 1 |
| Thyroid Effects | 0.9 | 0.9 |
| Central Nervous System (CNS) | 0.2 | 0.2 |
| Decreased body and organ weights | 0.03 | 0.03 |
| Skin | 0.4 | 0.4 |
| Cardiovascular Effects | 0.3 | 0.3 |
| Gastrointestinal Effects | 0.02 | 0.03 |
| Hematopoietic Effects | 0.9 | 0.9 |

Note:
Toxic endpoints - see Table H.7.4
RME Child Residential Hazard Indices - see Table H.7.2
The hazard indices are sumed for the toxic endpoint depend on ingestion or inhalation.


[^0]:    P: \ISEH $746040($ NewDA01)\05_Suppl RA \& MEC Haz Assess\4835 RISK ASSESSMENT\Final\Revision 1\Final 4835 Glenbrook Rd HHRA9-11-09.doc Rev 2

[^1]:    P: ISEH $7446040($ NewDA01) 105 _Suppl RA \& MEC Haz Assess 14835 RISK ASSESSMENT/Final1Revision 11Final 4835 Glenbrook Rd HHRA9-11-09.doc Rev 2 Contract No. DACA87-02-D-0005

[^2]:    P:IISEH $746040($ NewDA01) 105 _Suppl RA \& MEC Haz Assess 14835 RISK ASSESSMENT/Final\Revision 1\Final 4835 Glenbrook Rd HHRA9-11-09.doc Rev 2
    Contract No. DACA87-02-D-0005
    Delivery Order No. DA01

[^3]:    P:IISEH $1746040($ NewDA01) 105 _Suppl RA \& MEC Haz Assess 14835 RISK ASSESSMENT\FinallRevision 11Final 4835 Glenbrook Rd HHRA9-11-09.doc Rev 2

[^4]:    P:ISEH $1746040($ NewDA01) 105 _Suppl RA \& MEC Haz Assess 14835 RISK ASSESSMENT/Final\Revision 1\Final 4835 Glenbrook Rd HHRA9-11-09.doc Rev 2 Contract No. DACA87-02-D-0005

[^5]:    P:IISEH $1746040($ NewDA01) 105 _Suppl RA \& MEC Haz Assess 14835 RISK ASSESSMENT\FinallRevision 11Final 4835 Glenbrook Rd HHRA9-11-09.doc Rev 2 Contract No. DACA87-02-D-0005

[^6]:    P:\ISEH $1746040($ NewDA01) 105 _Suppl RA \& MEC Haz Assess 4835 RISK ASSESSMENT\Final\Revision 1\Final 4835 Glenbrook Rd HHRA9-11-09.doc Rev 2
    Contract No. DACA87-02-D-0005
    Delivery Order No. DA01

[^7]:    P: IISEH $1746040($ NewDA01) 105 _Suppl RA \& MEC Haz Assess 48835 RISK ASSESSMENTVFinal1Revision 1/Final 4835 Glenbrook Rd HHRA9-11-09.doc Rev 2 Contract No. DACA87-02-D-0005

[^8]:    P: IISEH $1746040($ NewDA01) 105 _Suppl RA \& MEC Haz Assess 48835 RISK ASSESSMENTVFinal1Revision 1/Final 4835 Glenbrook Rd HHRA9-11-09.doc Rev 2

[^9]:    P:IISEH $746040($ NewDA01) 105 _Suppl RA \& MEC Haz Assess 14835 RISK ASSESSMENT/Final\Revision 1\Final 4835 Glenbrook Rd HHRA9-11-09.doc Rev 2 Contract No. DACA87-02-D-0005

[^10]:    P:\ISEH 746040 (NewDA01) 105 _Suppl RA \& MEC Haz Assess 4835 RISK ASSESSMENT\Final\Revision 1\Final 4835 Glenbrook Rd HHRA9-11-09.doc Rev 2
    Contract No. DACA87-02-D-0005
    Delivery Order No. DA01

[^11]:    P: IISEH $1746040($ NewDA01) 105 _Suppl RA \& MEC Haz Assess 48835 RISK ASSESSMENTVFinal1Revision 1/Final 4835 Glenbrook Rd HHRA9-11-09.doc Rev 2

[^12]:    P: IISEH $1746040($ NewDA01) 105 _Suppl RA \& MEC Haz Assess 48835 RISK ASSESSMENTVFinal1Revision 1/Final 4835 Glenbrook Rd HHRA9-11-09.doc Rev 2

[^13]:    a/ COPC = Chemical of potential concern.
    b/ CAS $=$ Chemical Abstracts Service number.
    c/ Exposure Point Concentration,
    d/ mg/Kg-day = Milligrams per kilogram-day.
    e/ "--" = Data unavailable.

[^14]:    a/ COPC $=$ Chemical of potential concern.
    b/ CAS = Chemical Abstracts Service number.
    c/ $\mu \mathrm{g} / \mathrm{Kg}=$ Micrograms per kilogram.
    d/ $\mathrm{m}^{3} / \mathrm{kg}=$ Cubic meters per kilogram. Volatilization Factors used for volatile organic compounds only.
    e/ $\mu \mathrm{g} / \mathrm{m}^{3}=$ Micrograms per cubic meter.
    f/ "--" = Data unavailable.

[^15]:    a/ COPC = Chemical of potential concern.
    b/ CAS $=$ Chemical Abstracts Service number.
    c/ Exposure Point Concentration,
    $\mathrm{d} / \mathrm{mg} / \mathrm{Kg}$-day $=$ Milligrams per kilogram-day.
    e/ "--" = Data unavailable.

[^16]:    a/ COPC $=$ Chemical of potential concern.
    b/ CAS = Chemical Abstracts Service number.
    c/ $\mu \mathrm{g} / \mathrm{Kg}=$ Micrograms per kilogram.
    d/ $\mathrm{m}^{3} / \mathrm{kg}=$ Cubic meters per kilogram. Volatilization Factors used for volatile organic compounds only.
    e/ $\mu \mathrm{g} / \mathrm{m}^{3}=$ Micrograms per cubic meter.
    f/ "--" = Data unavailable.

[^17]:    a/ COPC = Chemical of potential concern.
    b/ CAS $=$ Chemical Abstracts Service number.
    c/ Exposure Point Concentration,
    $\mathrm{d} / \mathrm{mg} / \mathrm{Kg}$-day $=$ Milligrams per kilogram-day.
    e/ "--" = Data unavailable.

[^18]:    a/ COPC $=$ Chemical of potential concern.
    b/ CAS $=$ Chemical Abstracts Service number.
    c/ $\mu \mathrm{g} / \mathrm{Kg}=$ Micrograms per kilogram.
    d/ $\mathrm{m}^{3} / \mathrm{kg}=$ Cubic meters per kilogram. Volatilization Factors used for volatile organic compounds only.
    e/ $\mu \mathrm{g} / \mathrm{m}^{3}=$ Micrograms per cubic meter.
    f/ "--" = Data unavailable.

[^19]:    a/ COPC = Chemical of potential concern.
    b/ CAS = Chemical Abstracts Service number.
    c/ Exposure Point Concentration,
    d/ mg/Kg-day = Milligrams per kilogram-day.
    e/ "--" = Data unavailable.

[^20]:    a/ COPC $=$ Chemical of potential concern.
    b/ CAS = Chemical Abstracts Service number.
    c/ $\mu \mathrm{g} / \mathrm{Kg}=$ Micrograms per kilogram.
    $\mathrm{d} / \mathrm{m}^{3} / \mathrm{kg}=$ Cubic meters per kilogram. Volatilization Factors used for volatile organic compounds only.
    e/ $\mu \mathrm{g} / \mathrm{m}^{3}=$ Micrograms per cubic meter.
    f/ "--" = Data unavailable.

[^21]:    a/ COPC = Chemical of potential concern.
    b/ CAS $=$ Chemical Abstracts Service number.
    c/ Exposure Point Concentration,
    d/ mg/Kg-day = Milligrams per kilogram-day.
    e/ "--" = Data unavailable.

[^22]:    a/ COPC $=$ Chemical of potential concern.
    b/ CAS $=$ Chemical Abstracts Service number.
    c/ $\mu \mathrm{g} / \mathrm{Kg}=$ Micrograms per kilogram.
    d/ $\mathrm{m}^{3} / \mathrm{kg}=$ Cubic meters per kilogram. Volatilization Factors used for volatile organic compounds only.
    e/ $\mu \mathrm{g} / \mathrm{m}^{3}=$ Micrograms per cubic meter.
    f/ "--" = Data unavailable.

[^23]:    a/ COPC = Chemical of potential concern.
    b/ CAS $=$ Chemical Abstracts Service number.
    c/ Exposure Point Concentration,
    d/ mg/Kg-day = Milligrams per kilogram-day.
    e/ "--" = Data unavailable.

[^24]:    a/ COPC $=$ Chemical of potential concern.
    b/ CAS = Chemical Abstracts Service number.
    c/ $\mu \mathrm{g} / \mathrm{Kg}=$ Micrograms per kilogram.
    d/ $\mathrm{m}^{3} / \mathrm{kg}=$ Cubic meters per kilogram. Volatilization Factors used for volatile organic compounds only.
    e/ $\mu \mathrm{g} / \mathrm{m}^{3}=$ Micrograms per cubic meter.
    f/ "--" = Data unavailable.

[^25]:    a/ COPC = Chemical of potential concern.
    b/ CAS $=$ Chemical Abstracts Service number.
    c/ Exposure Point Concentration,
    d/ mg/Kg-day = Milligrams per kilogram-day.
    e/ "--" = Data unavailable.

[^26]:    a/ COPC $=$ Chemical of potential concern.
    b/ CAS = Chemical Abstracts Service number.
    c/ $\mu \mathrm{g} / \mathrm{Kg}=$ Micrograms per kilogram.
    d/ $\mathrm{m}^{3} / \mathrm{kg}=$ Cubic meters per kilogram. Volatilization Factors used for volatile organic compounds only.
    e/ $\mu \mathrm{g} / \mathrm{m}^{3}=$ Micrograms per cubic meter.
    f/ "--" = Data unavailable.

[^27]:    a/ COPC = Chemical of potential concern.
    b/ CAS $=$ Chemical Abstracts Service number.
    c/ Exposure Point Concentration,
    d/ mg/Kg-day = Milligrams per kilogram-day.
    e/ "--" = Data unavailable.

[^28]:    a/ COPC $=$ Chemical of potential concern.
    b/ CAS = Chemical Abstracts Service number.
    c/ $\mu \mathrm{g} / \mathrm{Kg}=$ Micrograms per kilogram.
    d/ $\mathrm{m}^{3} / \mathrm{kg}=$ Cubic meters per kilogram. Volatilization Factors used for volatile organic compounds only.
    e/ $\mu \mathrm{g} / \mathrm{m}^{3}=$ Micrograms per cubic meter.
    f/ "--" = Data unavailable.

[^29]:    a/ COPC = Chemical of potential concern.
    b/ CAS $=$ Chemical Abstracts Service number
    c/ Exposure Point Concentration,
    d/ mg/Kg-day = Milligrams per kilogram-day.
    e/ "--" = Data unavailable

[^30]:    a/ COPC $=$ Chemical of potential concern
    b/ CAS $=$ Chemical Abstracts Service number.
    c/ $\mu \mathrm{g} / \mathrm{Kg}=$ Micrograms per kilogram.
    $\mathrm{d} / \mathrm{m}^{3} / \mathrm{kg}=$ Cubic meters per kilogram. Volatilization Factors used for volatile organic compounds only.
    e/ $\mu \mathrm{g} / \mathrm{m}^{3}=$ Micrograms per cubic meter.
    f/ "--" = Data unavailable.

[^31]:    a/ COPC = Chemical of potential concern.
    b/ CAS $=$ Chemical Abstracts Service number.
    c/ Exposure Point Concentration,
    d/ mg/Kg-day = Milligrams per kilogram-day.
    e/ "--" = Data unavailable.

[^32]:    a/ COPC $=$ Chemical of potential concern.
    b/ CAS = Chemical Abstracts Service number.
    c/ $\mu \mathrm{g} / \mathrm{Kg}=$ Micrograms per kilogram.
    $\mathrm{d} / \mathrm{m}^{3} / \mathrm{kg}=$ Cubic meters per kilogram. Volatilization Factors used for volatile organic compounds only.
    e/ $\mu \mathrm{g} / \mathrm{m}^{3}=$ Micrograms per cubic meter.
    f/ "--" = Data unavailable.

[^33]:    a/ COPC = Chemical of potential concern.
    b/ CAS $=$ Chemical Abstracts Service number.
    c/ Exposure Point Concentration,
    d/ mg/Kg-day = Milligrams per kilogram-day.
    e/ "--" = Data unavailable.

[^34]:    a/ COPC $=$ Chemical of potential concern.
    b/ CAS = Chemical Abstracts Service number.
    c/ $\mu \mathrm{g} / \mathrm{Kg}=$ Micrograms per kilogram.
    d/ $\mathrm{m}^{3} / \mathrm{kg}=$ Cubic meters per kilogram. Volatilization Factors used for volatile organic compounds only.
    e/ $\mu \mathrm{g} / \mathrm{m}^{3}=$ Micrograms per cubic meter.
    f/ "--" = Data unavailable.

[^35]:    a/ COPC = Chemical of potential concern.
    b/ CAS $=$ Chemical Abstracts Service number.
    c/ Exposure Point Concentration,
    d/ mg/Kg-day = Milligrams per kilogram-day.
    e/ "--" = Data unavailable.

[^36]:    a/ COPC $=$ Chemical of potential concern.
    b/ CAS $=$ Chemical Abstracts Service number.
    c/ $\mu \mathrm{g} / \mathrm{Kg}=$ Micrograms per kilogram.
    d/ $\mathrm{m}^{3} / \mathrm{kg}=$ Cubic meters per kilogram. Volatilization Factors used for volatile organic compounds only.
    e/ $\mu \mathrm{g} / \mathrm{m}^{3}=$ Micrograms per cubic meter.
    f/ "--" = Data unavailable.

